

This discussion paper is/has been under review for the journal Hydrology and Earth System Sciences (HESS). Please refer to the corresponding final paper in HESS if available.

Impact of climate change on groundwater in a confined Mediterranean aquifer

Y. Caballero and B. Ladouche

BRGM, Water Environment and Ecotechnologies Division, 1039 rue de Pinville, 34000, Montpellier, France

Received: 23 July 2015 – Accepted: 10 August 2015 – Published: 2 October 2015

Correspondence to: Y. Caballero (y.caballero@brgm.fr)

Published by Copernicus Publications on behalf of the European Geosciences Union.

10109

Abstract

This paper presents an inverse modeling method based on wavelet analysis, devoted to assessment of the impacts of climate change on the groundwater resources of a confined coastal multi-layer aquifer, located in the south of France (*Pyrénées-Orientales*).
5 The hydraulic behavior of the aquifer is described based on the results of a model calibrated to simulate the groundwater dynamics observed on two representative piezometers. The relative contributions of the climate and pumping forcings to the piezometric variations are quantified. The results illustrate in quantitative terms the dominant influence of pumping on the temporal variations of the hydraulic head of the aquifer.
10 Based on this specific behavior simulation, we show the moderate vulnerability of such confined aquifers to climate change. Some insights regarding pumping strategies for confined coastal aquifers that could contribute towards preserving their good status in future are also provided.

1 Introduction

15 Although groundwater resources are often critical for agricultural and drinking water uses, the impact of climate change on groundwater is less frequently analyzed than for surface water. Existing studies at the regional scale focus on changes of aquifer recharge using reservoir models employing precipitation and PET (Thomsen, 1990; Chen et al., 2002), aquifer storage (groundwater level) and drainage (spring flow) using water balance models (Döll, 2009; Rivard et al., 2014; Wada et al., 2014), three-dimensional finite difference groundwater models (Loaiciga et al., 2000; Allen et al., 2004; Serrat-Capdevila et al., 2007; van Roosmalen et al., 2007; Woldeamlak et al., 2007; Carneiro et al., 2010; Goderniaux et al., 2011; Habets et al., 2013; Castaño et al., 2013) or soil-vegetation-atmosphere transfer schemes (Bouraoui et al., 1999; 25 Eckhardt and Ulbrich, 2003; Brouyère et al., 2004; Scibek and Allen, 2006; Caballero

10110

et al., 2007; Crosbie et al., 2013; Portmann et al., 2013). For a complete review of climate change impacts on groundwater see Green et al. (2011).

Existing studies generally focus on unconfined aquifers because (1) they are generally shallow (2) they are often in hydraulic connection with surface water where river-groundwater interactions control the impact of recharge on both systems (Allen et al., 2004; Scibek and Allen, 2006). Nevertheless, confined aquifers are of great interest especially for water supply since they generally provide a water resource protected from anthropogenic pollution. The few existing assessments of climate change impact on confined aquifers are usually performed using complex hydrodynamic models (see Green et al., 2011 or more recent papers such as Ali et al., 2012). For such models, detailed descriptions of the geometry and the hydraulic properties of the different layers are required, for a correct estimation of the transient flow component of the recharge (Seiler et al., 2008), as the aquifer response to loading from surface recharge mainly depends on the change of pressure in the overlying confining units (Maliva et al., 2011).

For complex geological contexts such as those that can be found in the Mediterranean sedimentary rim (Duvail et al., 2005; Clauzon et al., 2015), the description of the layers geometry and hydraulic properties may be challenging enough to prevent the use of a hydrodynamic model. Alternative modeling methods able to simulate the climate change impact without a detailed description of the aquifer geometry are thus required.

This paper aims to present a modeling method based on wavelet analysis for the assessment of the climate change impact on a confined multi-layer aquifer affected by major pumping exploitation. Such a method allows fitting a numerical relation between climate, pumping and groundwater level variation, without explicitly taking into account the aquifer geometry and its hydraulic properties. It thus represents an interesting option to explore global (ie. both climate and economic – considering the link between future economic situation and the pumping withdrawals evolution to for the latter) change impacts on such kind of complex aquifers.

10111

The following section describes the study area, the main hydrogeological characteristics of the confined aquifer studied, and the available data used for modeling. The third section presents a general description of the wavelet analysis and the way it has been used for the estimation of the pumped volumes and the groundwater level variation simulation. It also presents the climate scenarios used to force the model built. Results are presented and discussed in the fourth section, and main conclusions are summarized in the last section.

2 The Roussillon plain Plio–Quaternary (PQ) aquifer

The Roussillon basin is located along the southernmost part of the French Mediterranean coast, near the Spanish border. This 700 km² sedimentary basin is bordered by the foothills of the Pyrenean Mountains on the south and west, the Corbières karst region on the north and the Mediterranean Sea on the east (Fig. 1). It is embedded within a Miocene structured margin, the consequence of a marine regression following the Messinian salinity crisis (partial drying up of the Mediterranean Sea, approximately 6 My ago, Benson et al., 1991). The regression caused the excavation of deep canyons by fluvial erosion, which were then filled up with fluvial and marine sands, which now constitute various sedimentological units deposited according to the “Gilbert Delta” genetic model (Clauzon, 1990; Guennoc et al., 2000; Aunay et al., 2006). Two major aquifers, the Quaternary aquifer and the underlying continental and marine Pliocene aquifer, occupy these sediments.

The Quaternary one lies along the principal rivers (Agly, Têt and Tech) and the coastline, with thicknesses ranging from 10 m in the upper part of the valleys to more than 20 m in the coastal fringe (Aunay et al., 2006). It is mainly exploited by farmers, private individuals and campsite owners.

The Pliocene is divided into an upper system formed by prograding fluvial sands and a lower system of prograding marine sands and clays, separated by diachronous layers of lignite with plant remains alternating with marshy plastic-clays (Aunay et al.,

10112

2006). Its depth increases towards the coastline, where its maximum thickness reaches 300 m (Duvail et al., 2005). Both sedimentological units constitute productive confined aquifers, with a transmissivity (and associated storage coefficient) ranging from $10^{-2} \text{ m}^2 \text{ s}^{-1}$ ($4 \times 10^{-5} \text{ m}^{-1}$) for the continental sands, to $10^{-3} \text{ m}^2 \text{ s}^{-1}$ ($1 \times 10^{-5} \text{ m}^{-1}$) for the marine sands (Chabart, 1996). These latter layers are not present everywhere, implying potential local connections between the upper and lower Pliocene aquifers (Aunay et al., 2006).

A network for monitoring groundwater status has enabled the compilation of a database of piezometric levels over periods in some cases exceeding 30 years for certain sites (Fig. 1).

The data acquired indicate that the groundwater flows from west to east in the Pliocene aquifers. The hydraulic gradient is low (about 1 to 4‰), and flow rates are probably less than ten meters per year (Chery, 1992). Forty years ago the piezometric level of groundwater in the deep marine sands was artesian (Got, 1965; Aunay, 2007). Since then, major exploitation of the aquifers by drilling (essentially for drinking water supplies and agriculture) has caused a net decrease in the hydraulic loads, especially during the summer. In recent years and at certain sites, this decline has lowered the groundwater level below sea level during the summer months, when withdrawals are most intense.

The multilayer Plio–Quaternary Roussillon aquifer accounts for almost 80 % of the resources used for drinking water in the Département of Pyrénées Orientales. Withdrawals for this purpose have doubled in the past 30 years to 47 Mm^3 in 2007, including 29 Mm^3 from Pliocene formations (Terrasson et al., 2014). This increase in withdrawals is in line with the increase in the population within the study area which depend on this resource (+56 % in 30 years). However, a trend towards the stabilization of withdrawals for this purpose has been observed in recent years. Withdrawals from the Plio–Quaternary aquifer for uses other than drinking water supply are less well documented. They could be around $28 \text{ Mm}^3 \text{ yr}^{-1}$ (including 5 Mm^3 from the Pliocene) for agriculture, 4 Mm^3 from the Pliocene for industry, and 6 Mm^3 (including 1 Mm^3 in the Pliocene) from

10113

private wells. Thus, the total annual withdrawals from the Plio–Quaternary aquifer for all uses would reach a total of 83 Mm^3 , including 39 Mm^3 from the Pliocene (Terrasson et al., 2014). Accordingly, the importance of this aquifer for the Département's water resources quickly established a need to understand its functioning, so as to be able to manage the available resource in a sustainable manner.

A lot of work has been done to describe the structure and geometry of the Pliocene aquifer (Duvail et al., 2005; Chabart, 1996; Aunay, 2007); it emphasizes the strong spatial variation of the sedimentological units involved. The first difficulty faced when trying to deduce the hydrodynamic processes of the corresponding aquifers from their geological geometry is that there is a great lithological heterogeneity in the various sedimentological units described, e.g., continental sands and marine sands and clays. A consequence of this is that the hydrodynamic behavior of such units is also spatially variable. There are also connections between the units, which are not simple horizontal layers separated by units of low permeability but the distributary channels of a deltaic complex filled by sand. Moreover, many more than 500 boreholes have been drilled to exploit the available groundwater resource in the Roussillon basin, 35 % of which tap more than one sedimentological unit (Aunay, 2007). Thus pumping groundwater in a single borehole may have an impact on the hydraulic head of several sedimentological units. These major difficulties have so far prevented the construction of a 3-D gridded hydrodynamic model. Nonetheless, the monitoring network (Fig. 2) shows a rather coherent seasonal evolution of the groundwater levels at the scale of the whole aquifer system. Consequently, the evolution of one particular groundwater level time series of the piezometric network can roughly be considered representative of that of the whole aquifer system. Based on this fact, the data from two piezometers (10972X0137/PONT and 10908X0263/FIGUER), along those with the longer time series and located at contrasted distance of the sea shore (Fig. 1), have been analyzed to provide some insights about the dynamic of the whole aquifer system under present and future conditions.

10114

3 Modeling method

3.1 General approach

Two piezometers The daily variations of the water levels observed in the piezometers at Perpignan (French Database Index 10908X0263/FIGUER) and Argelès (French Database Index 10972X0137/PONT) are shown in Fig. 3. The groundwater level's variation is affected by recharge from rainfall, and pumping for domestic, industrial or agricultural needs. The groundwater level time series show a seasonal cycle at the annual scale and a long term tendency to fall (Fig. 3). To analyze whether the pumping could explain this, the monthly variation of the total pumping volume of all the boreholes located on the Pliocene aquifer during the 1998–2007 period, is compared to both time series (Fig. 3). As regards confined groundwater, it is considered that all the pumping operations targeting the aquifer will impact the piezometric levels measured at any point on it (Sen, 1995, page 170). In this study, various areas containing pumping operations likely to affect the Perpignan and Argelès piezometers were defined, on the assumption that the water flows are mainly oriented from west to east (Fig. 1). On the pumping measurement series, the seasonal pumping (SP) can be distinguished from the permanent pumping (PP) (Fig. 3). The PP is estimated from the pump discharges during the winter months (October to March). The SP is calculated by subtracting the PP from the pump discharges (total pumping).

We assume that annual variations in the piezometry are linked to the SP, while the long-term trend is linked to the PP from drinking water withdrawals, which have continued to grow for nearly 30 years (Terrasson et al., 2014). To test and verify the effect of pumping on the piezometry we applied the techniques of signal processing, and in particular the use of continuous wavelets (wavelet analysis sub-section). Cross characterizations of pumping and piezometry for the period 1998–2008 were then used to estimate the volumes pumped over the period 1974–1997, for which the volumes pumped from Pliocene aquifers are poorly known (sub-section on estimation of pumped volumes).

10115

Transfer models were then developed to model the observed piezometric variations (sub-section on Inverse modeling), taking as input components the computed withdrawals and the recharge (effective rainfall). Finally, these transfer models were used to simulate the impact of climate change on the Roussillon aquifer's water resources, and to facilitate its discussion.

3.2 Wavelet analysis

The variability and oscillations observed in the long-term hydrogeological time series in the monitored boreholes (Fig. 3) were studied using wavelet analysis. This analytic technique is now extensively used in geoscience for the visualization, identification and filtering of the main spectral components in time series (Labat, 2005, 2008, 2010; Lafrenière and Sharp, 2003; Maraun and Kurths, 2004; Yan and Jones, 2008). Choice of the appropriate wavelet depends on the nature of the signal and on the type of information to be extracted from it. We used the Continuous Wavelet Transform (CWT) to determine the dominant modes of variability and how those modes vary over time. The Paul wavelet method described in Torrence and Compo (1998), was chosen because, as reported by De Moortel et al. (2004), it allows a better time resolution, for any value of the wavelet parameter, than the Morlet wavelet.

Our wavelet analysis was aimed at the characterization of water-table variations in order to identify the energy associated with seasonal and long-term increases in pumping. The long-term trend was characterized by using the wavelet-filtered time series (Torrence and Compo, 1998). Calculations were performed using the Paul wavelet, taking into account the entire pass-band of the piezometric measurement series. The piezometric variables filtered by the Paul wavelet enable access to figures not influenced by the long-term trend linked to the increase in permanent pumping (Fig. 4). The difference between the wavelength-filtered record (y_1) and the measured piezometric record (y_0) enables the long-term variation of a trend ($y_1 - y_0$) to be characterized (Fig. 4). The situation at the beginning of the measurement series constitutes the reference value for each of the measurement series studied. The processing applied

10116

3.5 Climate scenarios

While early studies used simplified climate scenarios involving fixed ranges of precipitation or recharge against which to compare the groundwater response (Woldeamlak et al., 2007; Wilkinson and Cooper, 1993; Ferguson and Maxwell, 2010), an increasing number of studies now use climate scenarios derived from the outputs of climate models (Green et al., 2011). To build these climate scenarios, various methods of bias correction and downscaling are applied. These methods range from the simple but robust so-called perturbation method, based on the mean monthly differences between future and present measurements (Déqué, 2007), to more sophisticated methods (called statistical or dynamic) which mainly enable the distributions of future meteorological variables to be compared to present-day ones (Déqué, 2007; Boé et al., 2006; Fowler et al., 2007; Goderniaux et al., 2009, 2011).

In this study a set of five climate scenarios were constructed by using the perturbation method. This method consists of taking a meteorological data set representative of the present, and applying the mean monthly anomalies between future and present-day temperatures and precipitations simulated by the climate models (Déqué, 2007; Xu, 1999). Daily precipitation and potential evapotranspiration (PET) for the 1970–2006 period were extracted from the SAFRAN meteorological database (Vidal et al., 2010) available for the study zone on a regular 8 km × 8 km grid (Chaouche et al., 2010). Among the 21 climate-model outputs from the World Climate Research Program's (WCRP) Coupled Model Intercomparison Project Phase 3 (CMIP3) multi-model dataset (Meehl et al., 2007), which are available through the Program for Climate Model Diagnosis and Intercomparison (PCMDI), five climate models were selected over the French region. They were chosen as they were considered as able to capture both regional precipitation and temperature climatology for the Mediterranean region (Mariotti et al., 2008). These models are CNRM-CM3 (Salas y Melia et al., 2005), HadGEM1 (Johns et al., 2006), IPSL-CM4 (Hourdin et al., 2006), MPI-ECHAM5 (Roeckner et al., 2006; Jungclaus et al., 2006), and NCAR-CCSM3.0 (Collins et al., 2006). Atmospheric

10121

temperature and precipitation (spatial resolution from 1 to 3.75°) were extracted for the French region (8° W to 10° E, 40° to 52° N) and for the 1980–2000 (reference), 2020–2040 (near-term), and 2040–2060 (medium-term) periods, and for the SRES-A1B scenario. This greenhouse-gas emission scenario is considered to be a compromise between a more optimistic scenario such as the SRES-B1 and a more pessimistic one such as the SRES-A2, although they differ only slightly for the period studied (Hourdin et al., 2006; Nakicenovic and Swart, 2000).

The mean monthly anomalies of temperature and precipitation calculated from the five climate models provide a description of the average climate change between the two future periods considered and the present, together with an illustration of the uncertainties associated with climate modeling. A multi-model average increase in temperature of between +1.0 and +1.8 °C is projected, depending on the season, for the near-term (2020–40) compared to the present. The projected increase is even greater for the medium term (2040–60), between +1.7 and 3.3 °C, with in addition a greater inter-seasonal variation compared to the near term. For precipitation, there is no clear signal for the near term (–10.9 to +10.5 % depending on the season), while a decrease (–21.9 to –2.1 %) is projected for the medium term. The large uncertainty associated with the calculated anomalies is higher for precipitation than for temperature, for summer than for winter, and for the medium term than for the near term.

More details about climate scenarios can be found in Caballero et al. (2008). While more sophisticated downscaling methods than the perturbation one (Quintana-Seguí et al., 2010) and new climate model simulations, like those carried out for the AR5 IPCC report, (Terry and Boé, 2013) have been performed on our study area, they have so far not produced sharply different results in terms of monthly average anomalies.

10122

4 Results

4.1 Modeling of the piezometric levels at Argelès and Perpignan

The transfer model for the Argelès piezometer provides a satisfactory simulation of the 10 day sampling interval variations in the measured piezometry for the calibration period (Nash criterion = 0.82, Fig. 10a). These results were obtained without taking into account any recharge by the Tech River. The model was validated against the water levels measured for the 1990–1998 period (Fig. 12); it satisfactorily reproduces the seasonal pumping periods and the Nash criterion of 0.82 is considered acceptable.

The mean transit time (T_m) associated with the normalized impulse response is the mean value of the lag time (Pinault et al., 2005). The response of the aquifer to recharge by effective rainfall is highly inertial, since T_m is 220 days, for an impulse response length of 520 days (Fig. 10c). The relative contribution of a hydrological cycle's recharge (fast + slow component) results in piezometric variations of about 0.5 to 0.8 m (Fig. 10b). The relative contribution of SP to the variations in piezometry is about –0.8 m over a hydrological cycle (Fig. 10b). The maximum response of the aquifer to seasonal pumping (SP) is detected during the month of August (Fig. 6f). It occurs 75 days after the start of pumping (Fig. 10d). The increase in PP over the period 1991–2007 is expressed by a decrease in the piezometry of about –0.6 m over the 17 years considered.

For the Perpignan site, the transfer model allows reproduction of the measured piezometric variations only if a water supply from the Têt River is considered. The transfer model enables a satisfactory reproduction of the variation in the groundwater level measured at the Perpignan piezometer (Nash criterion = 0.85, Fig. 11a). The model was validated with the water levels observed for the 1976–1994 period (Fig. 11a) which were satisfactorily reproduced (Nash criterion = 0.66). The flow rate of the Têt River upstream of the Roussillon plain is sustained in spring and summer by the release of water from the Vinca dam, which introduces a phase shift in the river's flow regime with respect to its natural regime. The contribution of recharge by precipitation thus differs

10123

both temporally and quantitatively from the releases at the dam on the Têt (the recharge causes an increase in the piezometric level of a maximum of 0.5 m and is expressed mainly in winter, whereas the Têt's contribution causes a rise in the level of between 0.3 and 0.8 m, generally expressed in the summer – Fig. 11b). The response of the aquifer to recharge by effective rainfall appears less inertial than that characterized at Argelès (maximum and mean response time 60 and 89 days respectively – Fig. 11c). The aquifer's response to recharge by releases from the dam reaches a maximum after 70–80 days (Fig. 11c). This maximum response is detected in late July, which is consistent with the fact that the dam releases occur generally in April of each year.

Seasonal pumping (SP) causes piezometric variations of between –1.8 and –3 m over a hydrological cycle (Fig. 11b). The aquifer's maximum response to seasonal pumping (SP) occurs 75 days after the start of pumping (Fig. 11d); this maximum response is detected in late June (Fig. 6e) vs. late July for Argelès (Fig. 11b). For Perpignan, recharge by the Têt River, which is at a maximum during July, masks the effect of seasonal pumping (SP), which should reach its maximum in late July.

The increase in PP over the period 1974–2007 is expressed by a decrease in the piezometry of about –0.9 m over the 33 year period studied. Table 1 summarizes the characteristics of the models applied to reproduce the piezometric variations observed in the Argelès and Perpignan piezometers. The greater inertia observed at Argelès is illustrated by smaller piezometric variations (cf. Figs. 11a and 12, lower panel), a greater impulse-response length (520 days for Argelès and 300 days for Perpignan) and a markedly longer average transit time (T_m : 220 days for Argelès vs. 89 days for Perpignan). It also shows the spatial heterogeneity of the hydraulic behavior of the sandy Pliocene formations.

Near Perpignan, we note that the Pliocene aquifer responds to releases from the Têt dam more slowly than to recharge (τ_3 : 70–80 d vs. τ : 60 d in Table 1). This is probably connected to the fact that the water released at the dams is mostly directed to irrigated areas, sometimes located more than 20 km downstream, via canals which may or may not be leakproof. These canals enable the irrigation of tree and truck-farming crops

10124

This is probably related to the fact that, for Argelès, the contribution of the Tech river was not explicitly taken into account because it is synchronous with the contribution from precipitation. However, this contribution is probably included in the two seasonal components used in the model (recharge and SP). In this case, it leads to an overestimation of the recharge component and/or an underestimation of the SP component (hence its smaller contribution to water level changes as compared to that seen at Perpignan), although it is impossible to quantify them. Since the contribution of the Têt river is explicitly taken into account in modeling the piezometric level at Perpignan (via the releases from the Vinca dam), the quantification of the SP component is probably more realistic than at Argelès.

Figure 12 (case a) shows that if pumping stops, the simulated water level rises abruptly back up, reaching an equilibrium level in less than one hydrological cycle. This is related to the mean transit time of the impulse response to effective rainfall employed in the model, whose duration is less than one year (Table 1). Similarly, if we consider a zero recharge (Fig. 12b and c), the simulated piezometric level remains at an equilibrium level and does not fall over time, in spite of the pumping. The model thus expresses the highly confined nature of the aquifer in the Argelès area, where variations in the piezometric level are controlled by rapid changes of pressure within the aquifer rather than by a flow dynamic known to be slow (flow rates of less than 10 m yr^{-1} , obtained from tritium and carbon-14 dating by Chery, 1992). However, although the functioning of the model in a no-recharge situation may be viewed as realistic, transient variations of the piezometric level are mainly controlled by the permanent pumping (the model was constructed on the assumption that the long-term trend is fully explained by the signal of increase in the PPs). This explains why the simulated piezometric level remains stable on Fig. 12b and c, instead of gradually falling under the effect of pumping, in the absence of recharge. Now, a rising trend in the temperatures of the study area was noted between 1971 and 2000 (Chaouche et al., 2010). This increase may have caused a reduction in the effective rainfall over the period 1970–2008. In this regard, a portion of the downward trend in water levels could also be related to long-

10129

term climate variability. This is the major limitation of the modeling approach adopted, and which we cannot escape when dealing with confined aquifers subjected to major exploitation by pumping.

The tests presented in Fig. 12 show that the equilibrium state of the water levels in the Argelès area is controlled by PPs. If permanent pumping on the aquifer were doubled, it is probable that the model would simulate a fall in the piezometric level until it reached a new steady state. From a hydrodynamic perspective it can therefore be considered that this aquifer would be capable of supporting a larger pumping operation. The aquifer thus appears to be relatively invulnerable to climate change. This idea is backed up by the projections in Fig. 13, which show that in the medium term the simulated piezometric level for Argelès would be affected by a decrease of no more than about 0.5 m. This decrease is less than that caused by the current exploitation of the aquifer (which produces a decrease of more than 1.5 m, taking PP and SP together). It would therefore appear possible to offset the declining recharge projected in the climate scenarios by a well thought-out exploitation of the aquifer, which could probably be achieved with improved water-use efficiency and water-demand management (Green et al., 2011). The relationship between the volume pumped by the PP and the drop in the water level (Fig. 7) indicates that over the near and medium term we could maintain the current piezometric level by retaining the current permanent pumping of about 7 million m^3 (vs. about 8 million currently) for Perpignan and 2 million m^3 (vs. some 2.2 million at present) for Argelès.

Nevertheless, the vulnerability of this coastal aquifer to saline intrusion problems may still be exacerbated by the effect of CC. If we extrapolate the simulated medium-term fall in piezometric level at Argelès (of about 50 cm) to the station at Barcarès (French database index 0912X0112/BAR3), located on the beach at the mouth of the Agly river, its piezometric level would fall to about 2 m below sea level under summer conditions. The great sensitivity of the water level to seasonal pumping identified by the modeling (Figs. 10 and 11) could then be used to limit this fall by reducing coastal withdrawals during the summer. This reduction could be offset by pumping in wintertime or alter-

10130

- ical model, *Environmental Earth Sciences*, 61, 241–252, doi:10.1007/s12665-009-0339-3, 2010.
- Castaño, S., Sanz, D., and Gómez-Alday, J. J.: Sensitivity of a groundwater flow model to both climatic variations and management scenarios in a semi-arid region of SE Spain, *Water Resour. Manag.*, 27, 2089–2101, doi:10.1007/s11269-013-0277-4, 2013.
- 5 Chabart, M.: La recharge de l'aquifère multicouche du Roussillon et les conséquences d'un éventuel changement climatique sur la gestion des ressources en eau (Pyrénées Orientales), Doctorat en Hydrogéologie, Université de Paris VI, 299 pp., 1996.
- Chaouche, K., Neppel, L., Dieulin, C., Pujol, N., Ladouche, B., Martin, E., Salas y Melia, D., and Caballero, Y.: Analyses of precipitation, temperature and evapotranspiration in a French Mediterranean region in the context of climate change, *CR Geosci.*, 342, 234–243, doi:10.1016/j.crte.2010.02.001, 2010.
- 10 Chen, Z., Grasby, S. E., and Osadez, K. G.: Predicting average annual groundwater levels from climatic variables: an empirical model, *J. Hydrol.*, 260, 102–117, 2002.
- 15 Chery, L.: Interpretation des résultats des analyses isotopiques des eaux du multicouche du Roussillon, n° BRGM/RR-36094-FR, 1–39, available at: <http://www.brgm.fr/publication/pubDetailRapportSP.jsp?id=RSP-BRGM/RR-36094-FR>, (last access: 17 July 2014), 1992.
- Clauzon, G.: Restitution de l'évolution géodynamique néogène du bassin du Roussillon et de l'unité adjacente des Corbières d'après les données écostratigraphiques et paléogéographiques, *Paléobiologie Continentale*, Montpellier, 17, 125–155, 1990.
- 20 Clauzon, G., Le Strat, P., Duvail, C., Do Couto, D., Suc, J., Molliex, S., Bache, F., Besson, D., Lindsay, E. H., Opdyke, N. D., Rubino, J., Popescu, S., Haq, B. U., and Gorini, C.: The Roussillon Basin (S. France): a case-study to distinguish local and regional events between 6 and 3 Ma, *Mar. Petrol. Geol.*, in press, doi:10.1016/j.marpetgeo.2015.03.012, 2015.
- 25 Collins, W. D., Bitz, C. M., Blackmon, M. L., Bonan, G. B., Bretherton, C. S., Carton, J. A., Chang, P. D., S. C., Hack, J. J., Henderson, T. B., Kiehl, J. T., Large, W. G., McKenna, D. S., Santer, B. D., and Smith, R. D.: The Community Climate System Model Version 3 (CCSM3), *J. Climate*, 19, 2122–2143, doi:10.1175/JCLI3761.1, 2006.
- Crosbie, R. S., Scanlon, B. R., Mpelasoka, F. S., Reedy, R. C., Gates, J. B., and Zhang, L.: Potential climate change effects on groundwater recharge in the High Plains Aquifer, USA, *Water Resour. Res.*, 49, 3936–3951, doi:10.1002/wrcr.20292, 2013.
- 30

10135

- De Moortel, I., Munday, S. A., and Hood, A. W.: Wavelet analysis: the effect of varying basic wavelet parameters, *Sol. Phys.*, 222, 203–228, doi:10.1023/B:SOLA.0000043578.01201.2d, 2004.
- Déqué, M.: Frequency of precipitation and temperature extremes over France in an anthropogenic scenario: model results and statistical correction according to observed values, *Global Planet. Change*, 57, 16–26, doi:10.1016/j.gloplacha.2006.11.030, 2007.
- 5 Dietrich, C. R. and Chapman, T. G.: Unit graph estimation and stabilization using quadratic programming and difference norms, *Water Resour. Res.*, 29, 2629–2635, 1993.
- Döll, P.: Vulnerability to the impact of climate change on renewable groundwater resources: a global-scale assessment, *Environ. Res. Lett.*, 4, 035006, doi:10.1088/1748-9326/4/3/035006, 2009.
- 10 Dörfli, N., Fleury, P., and Ladouche, B.: Inverse modeling approach to allogenic karst system characterization, *Ground Water*, 47, 414–426, doi:10.1111/j.1745-6584.2008.00517.x, 2009.
- 15 Duvail, C., Gorini, C., Lofi, J., Le Strat, P., Clauzon, G., and Dosreis, A.: Correlation between onshore and offshore Pliocene–Quaternary systems tracts below the Roussillon Basin (eastern Pyrenees, France), *Mar. Petrol. Geol.*, 22, 747–756, doi:10.1016/j.marpetgeo.2005.03.009, 2005.
- Eckhardt, K. and Ulbrich, U.: Potential impacts of climate change on groundwater recharge and streamflow in a central European low mountain range, *J. Hydrol.*, 284, 244–252, 2003.
- 20 Ferguson, I. M. and Maxwell, R. M.: Role of groundwater in watershed response and land surface feedbacks under climate change, *Water Resour. Res.*, 46, W00F02, doi:10.1029/2009WR008616, 2010.
- Fowler, H. J., Kilsby, C. G., and Stunell, J.: Modelling the impacts of projected future climate change on water resources in north-west England, *Hydrol. Earth Syst. Sci.*, 11, 1115–1126, doi:10.5194/hess-11-1115-2007, 2007.
- 25 Goderniaux, P., Brouyère, S., Fowler, H. J., Blenkinsop, S., Therrien, R., Orban, P., and Dassargues, A.: Large scale surface–subsurface hydrological model to assess climate change impacts on groundwater reserves, *J. Hydrol.*, 373, 122–138, doi:10.1016/j.jhydrol.2009.04.017, 2009.
- 30 Goderniaux, P., Brouyère, S., Blenkinsop, S., Burton, A., Fowler, H. J., Orban, P., and Dassargues, A.: Modeling climate change impacts on groundwater resources

10136

- Pinault, J. L., Plagnes, V., Aquilina, L., and Bakalowicz, M.: Inverse modeling of the hydrological and the hydrochemical behavior of hydrosystems: characterization of Karst System Functioning, *Water Resour. Res.*, 37, 2191, doi:10.1029/2001WR900018, 2001b.
- Pinault, J. L., Dörfli, N., Ladouche, B., and Bakalowicz, M.: Characterizing a coastal karst aquifer using an inverse modeling approach: the saline springs of Thau, southern France, *Water Resour. Res.*, 40, W08501, doi:10.1029/2003WR002553, 2004.
- Pinault, J. L., Amraoui, N., and Golaz, C.: Groundwater-induced flooding in macropore dominated hydrological system in the context of climate changes, *Water Resour. Res.*, 41, W05001, doi:10.1029/2004WR003169, 2005.
- Portmann, F. T., Döll, P., Eisner, S., and Fl, M.: Impact of climate change on renewable groundwater resources: assessing the benefits of avoided greenhouse gas emissions using selected CMIP5 climate projections, *Environ. Res. Lett.*, 8, 024023-, doi:10.1088/1748-9326/8/2/024023, 2013.
- Quintana-Seguí, P., Ribes, A., Martin, E., Habets, F., and Boé, J.: Comparison of three downscaling methods in simulating the impact of climate change on the hydrology of Mediterranean basins, *J. Hydrol.*, 383, 111–124, doi:10.1016/j.jhydrol.2009.09.050, 2010.
- Rivard, C., Paniconi, C., Vigneault, H., and Chaumont, D.: A watershed-scale study of climate change impacts on groundwater recharge (Annapolis Valley, Nova Scotia, Canada), *Hydrolog. Sci. J.*, 59, 1437–1456, doi:10.1080/02626667.2014.887203, 2014.
- Roeckner, E., Brokopf, R., Esch, M., Giorgetta, M., Hagemann, S., Kornbluh, L., Manzini, E., Schlese, U., and Schulzweida, U.: Sensitivity of simulated climate to horizontal and vertical resolution in the ECHAM5 atmosphere model, *J. Climate*, 19, 3771–3791, 2006.
- Salas-Méla, D., Chauvin, F., Déqué, M., Douville, H., Gueremy, J. F., Marquet, P., Planton, S., Royer, J. F., and Tyteca, S.: Description and validation of the CNRM-CM3 global coupled model, CNRM working note 103, available at: http://www.cnrm.meteo.fr/scenario2004/references_eng.html, (last access: 17 July 2014), 2005.
- Scibek, J. and Allen, D. M.: Modeled impacts of predicted climate change on recharge and groundwater levels, *Water Resour. Res.*, 42 doi:10.1029/2005WR004742, 2006, 2006.
- Seiler, K. P., Gu, W. Z., and Stichler, W.: Transient response of groundwater systems to climate changes, *Geol. Soc. Spec. Publ.*, 288, 111–119, doi:10.1144/SP288.9, 2008.
- Sen, Z.: *Applied Hydrogeology for Scientists and Engineers*, 464 pp., CRC Press, Boca Raton, Lewis Publishers, USA, ISBN 9781566700917, 1995.

10139

- Serrat-Capdevila, A., Valdes, J. B., González-Pérez, J., Baird, K., Mata, L. J., and Madock, T.: Modeling climate change impacts – and uncertainty – on the hydrology of a riparian system: the San Pedro Basin (Arizona/Sonora), *J. Hydrol.*, 347, 48–66, doi:10.1016/j.jhydrol.2007.08.028, 2007.
- Terrasson, I., Chazot, S., Maton, L., Rinaudo, J., and Caballero, Y.: Les Pyrénées Orientales face aux changements globaux: comparaison des effets des changements climatique et anthropique sur les besoins en eau, *La Houille Blanche*, n° 3, p. 53–63, 2014.
- Terray, L. and Boé, J.: Quantifying 21st-century France climate change and related uncertainties, *CR Geosci.*, 345, 136–149, doi:10.1016/j.crte.2013.02.003, 2013.
- Thomsen, R.: Effect of climate variability and change in groundwater in Europe, *Nord. Hydrol.*, 21, 185–194, 1990.
- Tikhonov, A. and Arsenine, V.: *Méthodes de Résolution de Problèmes mal Posés*, MIR, Moscow, 1976.
- Tikhonov, A. N. and Goncharsky, A. V. (Eds.): *Ill-Posed Problems in the Natural Sciences*, MIR, Moscow, 1987.
- Torrence, C. and Compo, G. P.: A practical guide for wavelet analysis, *B. Am. Meteorol. Soc.*, 79, 1, 61–78, 1998.
- Torrence, C. and Webster, P. J.: Interdecadal changes in the ENSO–monsoon system, *J. Climate*, 12, 2679–2690, doi:10.1175/1520-0442(1999)012<2679:ICITEM>2.0.CO;2, 1999.
- Van Roosmalen, L., Christensen, B. S. B., and Sonnenborg, T. O.: Regional differences in climate change impacts on groundwater and stream discharge in Denmark, *Vadose Zone J.*, 6, 554, doi:10.2136/vzj2006.0093, 2007.
- Vidal, J. P., Martin, E., Franchistéguy, L., Baillon, M., and Soubeyroux, J. M.: A 50 year high-resolution atmospheric reanalysis over France with the Safran system, *Int. J. Climatol.*, 30, 1627–1644, doi:10.1002/joc.2003, 2010.
- Wada, Y. and Bierkens, M. F. P.: Sustainability of global water use: past reconstruction and future projections, *Environ. Res. Lett.*, 9, 104003, doi:10.1088/1748-9326/9/10/104003, 2014.
- Wilkinson, W. B. and Cooper, D. M.: The response of idealized aquifer/river systems to climate change/Reponse de systemes aquifere-riviere schematiques aux modifications climatiques, *Hydrolog. Sci. J.*, 38, 379–390, doi:10.1080/026266693099492688, 1993.
- Woldeamlak, S. T., Batelaan, O., and De Smedt, F.: Effects of climate change on the groundwater system in the Grote-Nete catchment, Belgium, *Hydrogeol. J.*, 15, 891–901, 2007.

10140

- Xu, C.-Y.: From GCMs to river flow: a review of downscaling methods and hydrologic modelling approaches, *Prog. Phys. Geog.*, 23, 229–249, 1999.
- Yan, Z. and Jones, P. D.: Detecting inhomogeneity in daily climate series using wavelet analysis, *Adv. Atmos. Sci.*, 25, 157–163, doi:10.1007/s00376-008-0157-7, 2008

10141

Table 1. Main characteristics of the modeling for the Argelès and Perpignan piezometers. τ is the normalized impulse response of the recharge from effective rainfall; τ_1 is the normalized impulse response of the SP; τ_2 is the normalized impulse response of the PP; τ_3 is the normalized impulse response of the Têt river contribution.

Parameter	Argelès	Perpignan
Nash index for the calibration period	0.82	0.85
Nash index for the validation period	0.71	0.66
Relative contribution of effective recharge to water level changes	From +0.5 to +0.8 m	< 0.5 m
Length of the normalized impulse response of the effective rainfall (τ)	520 days	300 days
Mean transit time (T_m) of the normalized impulse response of (τ)	220 days	89 days
Lag for the maximum response of the normalized impulse response (τ)	30 days	60 days
Relative contribution of the seasonal pumping (SP) to water level changes	–1 m	From –1.8 m to 3 m
Date for the maximum contribution of the seasonal pumping (calculated from τ_1)	Mid-Aug	End of Jun
Relative contribution of permanent pumping (PP) to water level changes	–0.6 m over 17 years	–0.9 m over 33 years
Relative contribution of the river to water level changes	–	From +0.2 to 0.8 m
Lag for the maximum response of the normalized impulse response (τ_3)	–	70–80 days
Date for the maximum contribution of the river (calculated from τ_3)	–	End of Jul

10142

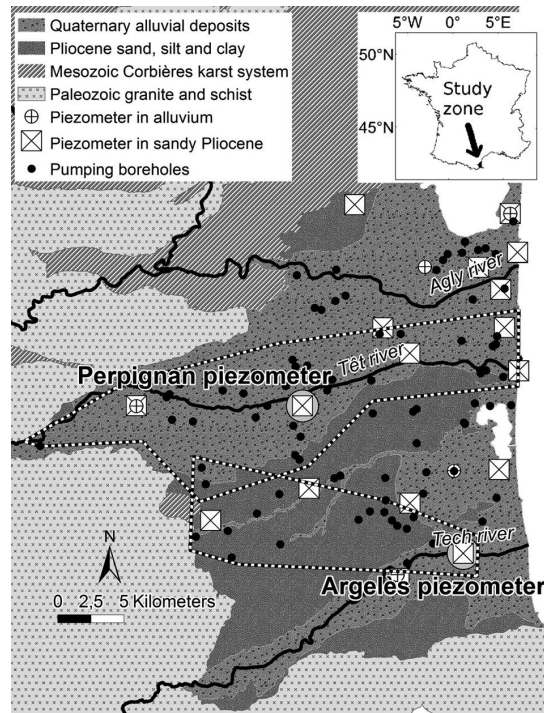


Figure 1. Geological context and piezometric network locations for monitoring the Roussillon Plio–Quaternary aquifer. Two specific piezometers (Perpignan and Argelès) are highlighted, over which modeling has been performed. Dashed outlines show the boreholes that were taken into account for each piezometer.

10143

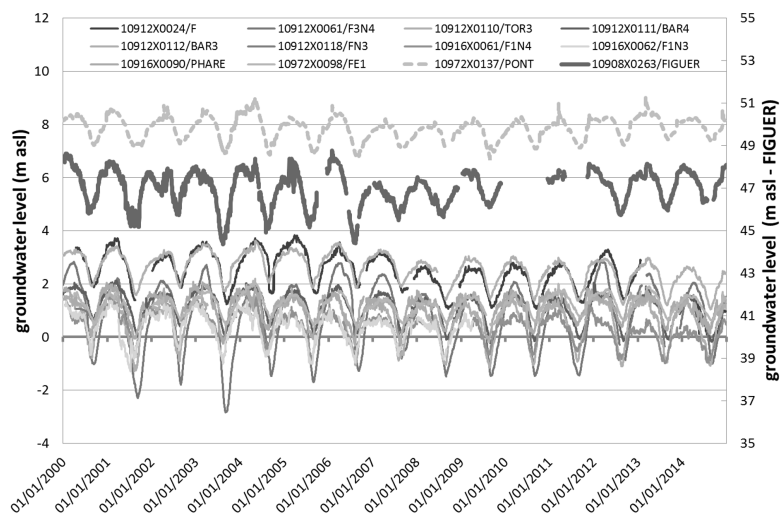


Figure 2. Groundwater level time series recorded over the last 15 years with the piezometer network of the Pliocene aquifer.

10144

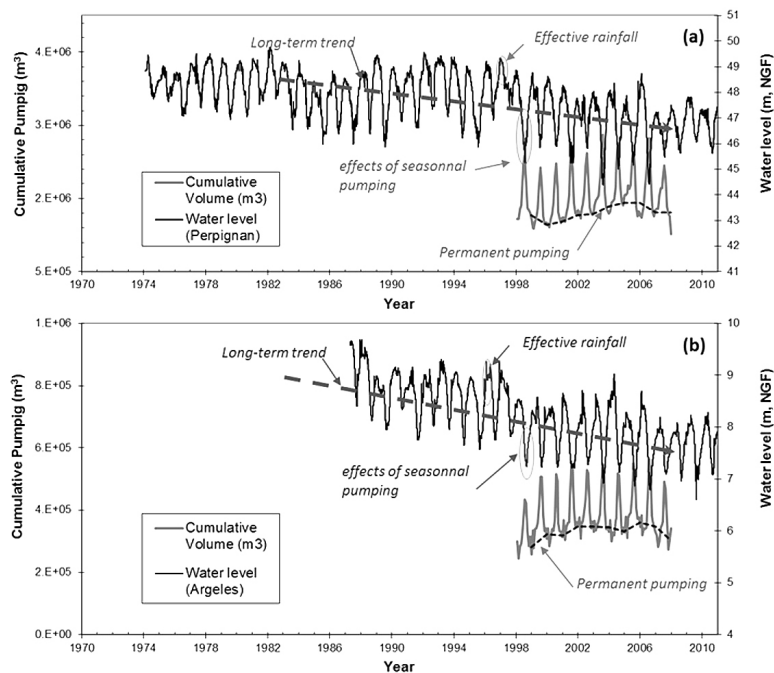


Figure 3. Daily piezometric data (ma.s.l.) at Perpignan (a) and Argelès (b) (see Fig. 1 for location) showing the long-term depletion trend, the effect of recharge by net rainfall, and the effect of summer water withdrawals. Pumping cumulated at the yearly time scale for the 1998–2007 period is also shown. The thin dashed line represents the permanent pumping equivalent of pumping during the winter period (October to March).

10145

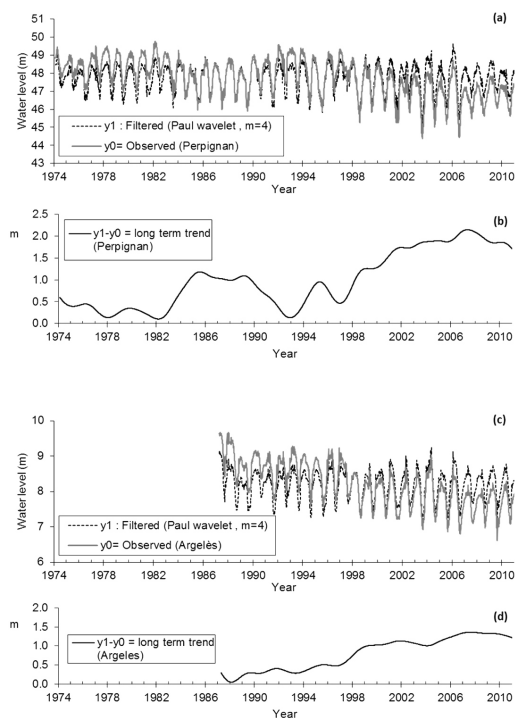


Figure 4. Wavelet-filtered time series of water levels (Perpignan and Argelès) computed with the Paul wavelet.

10146

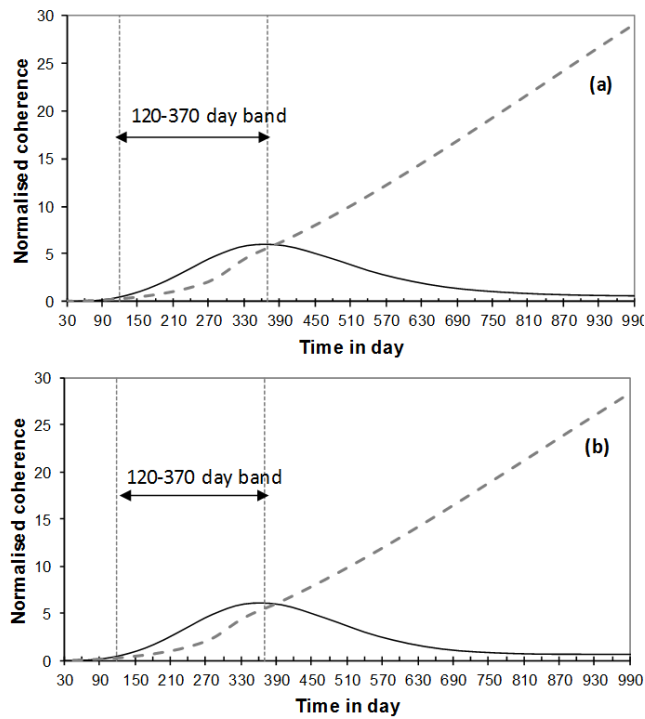


Figure 5. Fourier squared coherence calculated with Paul wavelet ($m = 4$) between pumping data (input) and water level (output) for the Perpignan piezometer **(a)** and Argelès piezometer **(b)**. The dashed line is the 95 % confidence level.

10147

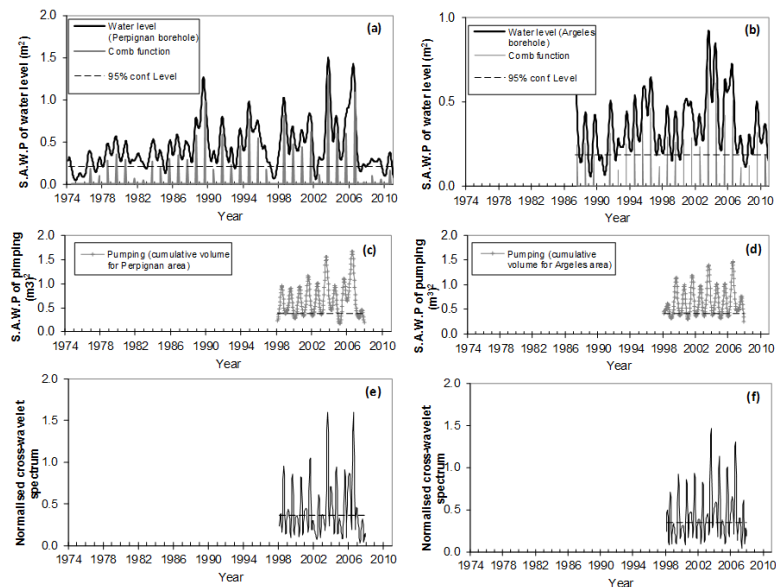


Figure 6. **(a, b)** Scale–Averaged Wavelet Power (SAWP) over the 120–371 day band for water levels measured in the Perpignan **(a)** and Argelès **(b)** piezometers using the Paul wavelet ($m = 4$). The dashed line is the 95 % confidence level. The thin solid line is the SAWP corrected to the 95 % confidence level, calculated for the month of August. This function is called the “comb function”. **(c, d)**: SAWP over the 120–371 day band for pumping in the Perpignan area **(c)** and Argelès area **(d)** using the Paul wavelet ($m = 4$). The dashed line is the 95 % confidence level. **(e, f)**: cross SAWP over the 120–371 day band with pumping data (input) and water level (output) calculated with the Paul wavelet ($m = 4$) for Perpignan **(e)** and Argelès **(f)**. The dashed line is the 95 % confidence level.

10148

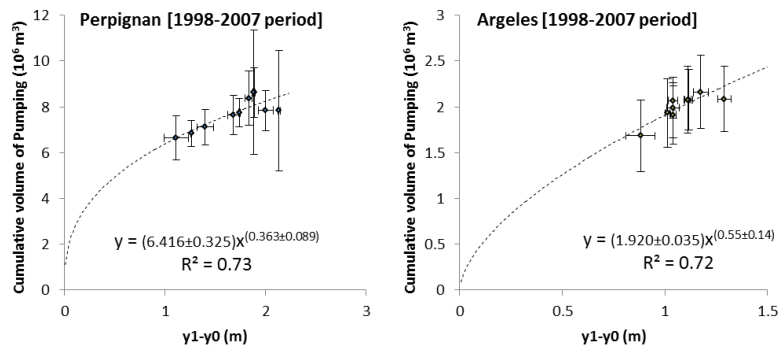


Figure 7. Changes in the volume of PP (1998–2007 period) based on the long-term trend.

10149

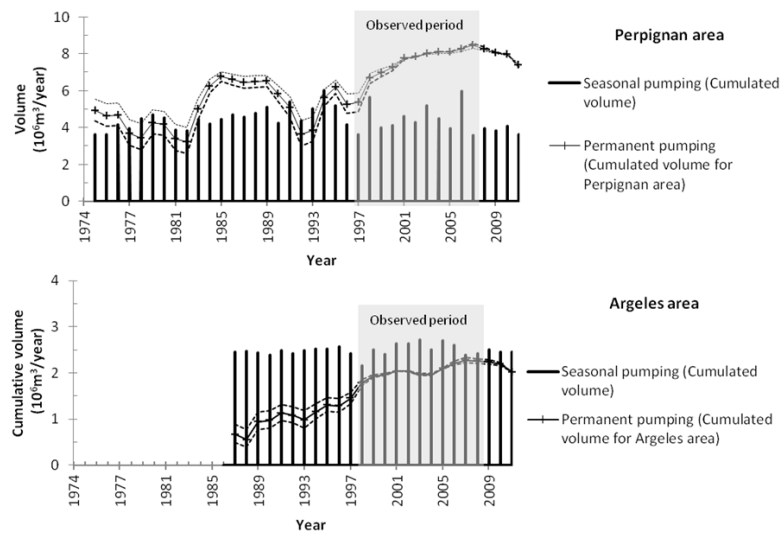


Figure 8. Variation of pumped volumes observed between 1998 and 2007 and reconstituted volumes for the preceding period. The permanent pumping was estimated from the long-term trend deduced from the Wavelet-filtered time series. The seasonal pumping volume (comb function) was estimated from the maximum SAWP (corrected to the 95 % confidence level) calculated for the month of August.

10150

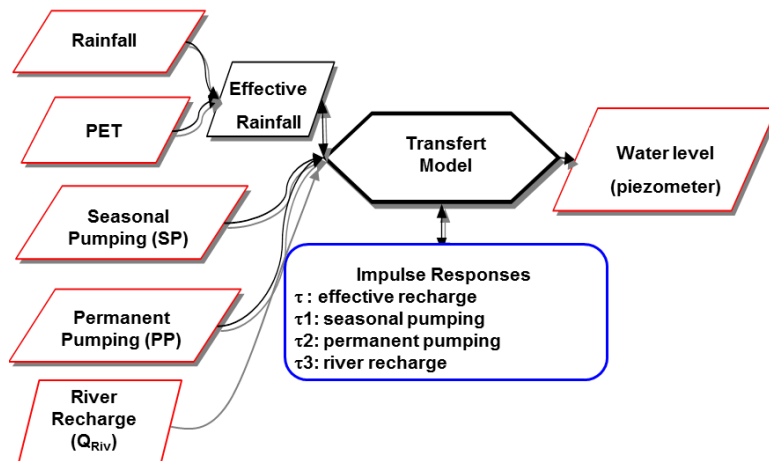


Figure 9. Architecture of the model used to simulate the water levels in the Perpignan and Argelès piezometers located on the PQ aquifer.

10151

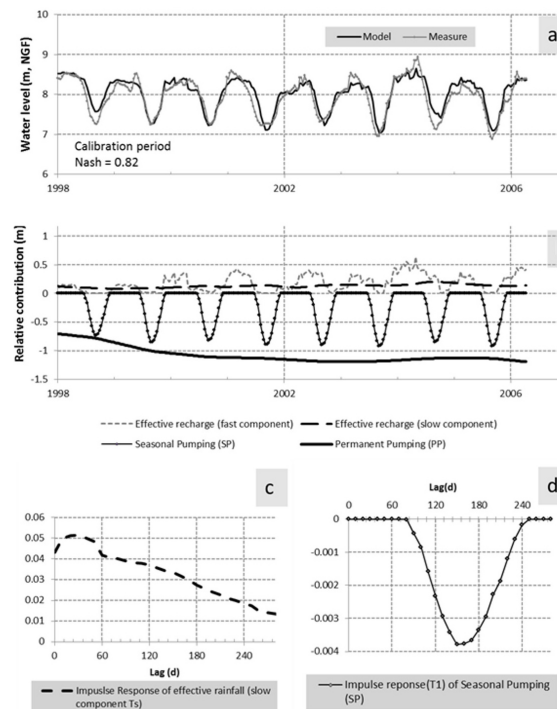


Figure 10. Inverse modeling of Pliocene confined aquifer (Argelès Piezometer) with a 10 day step. The measured and simulated piezometry is shown (a) as well as the relative contribution of the components (b), the impulse response of recharge (c) and of the SP (d).

10152

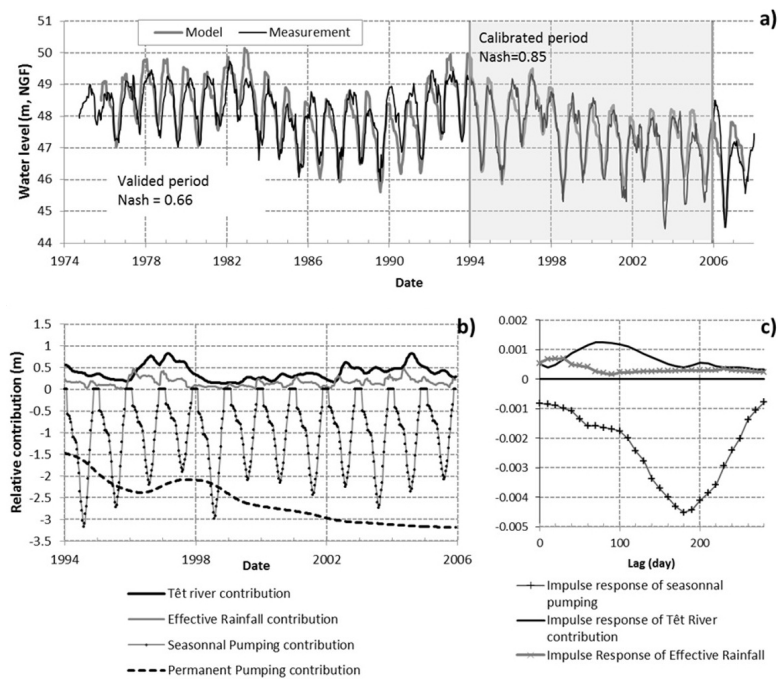


Figure 11. Inverse modeling of Pliocene confined aquifer (Perpignan Piezometer) with a 10 day step. The measured and simulated piezometry is shown (a) as well as the relative contributions of the recharge, pumping, and Têt river components (b) and their impulse responses (c).

10153

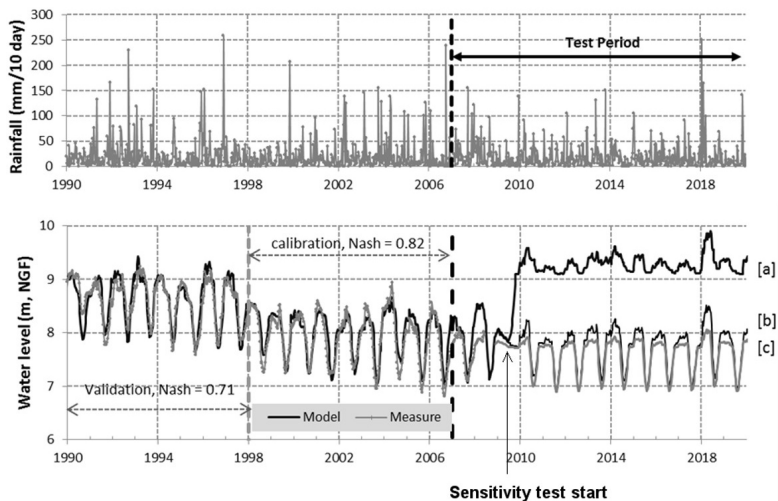


Figure 12. Sensitivity test of the model constructed for the Argelès piezometer for various conditions: variation of the piezometry calculated by the model in the absence of pumping [a] and for two precipitation situations, taking pumping into account: [b] = 50 % of the computed rain for the period 2008–2017 and [c] = 10 % of the computed rain for the period 2008–2017 (SP and PP pumping conditions identical to those of 2008).

10154

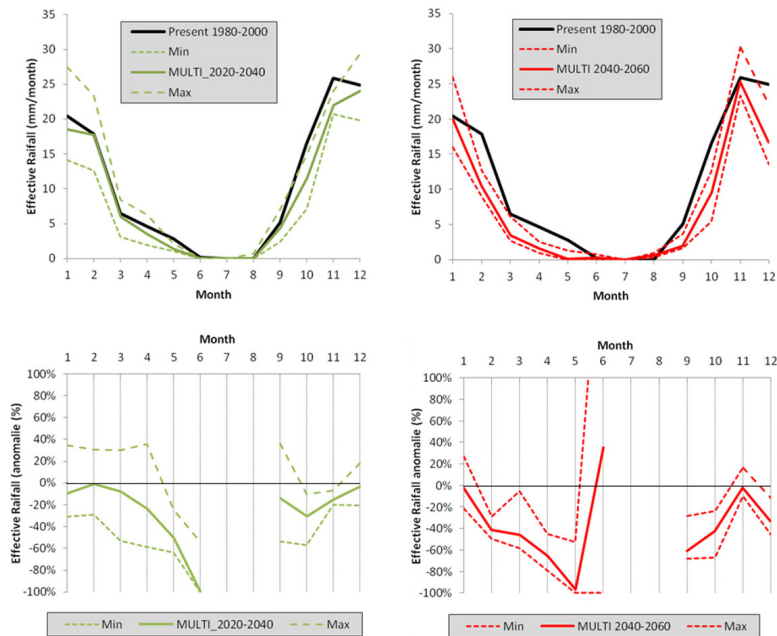


Figure 13. Variation of mean monthly values of effective rainfall for the reference period (1980–2000) and the short- (2020–2040) and medium- (2040–2060) term periods. The dashed lines represent the scenarios’ dispersions limited by the minimum and maximum values of monthly averages considering all scenarios. January = Month 1 and December = Month 12.

10155

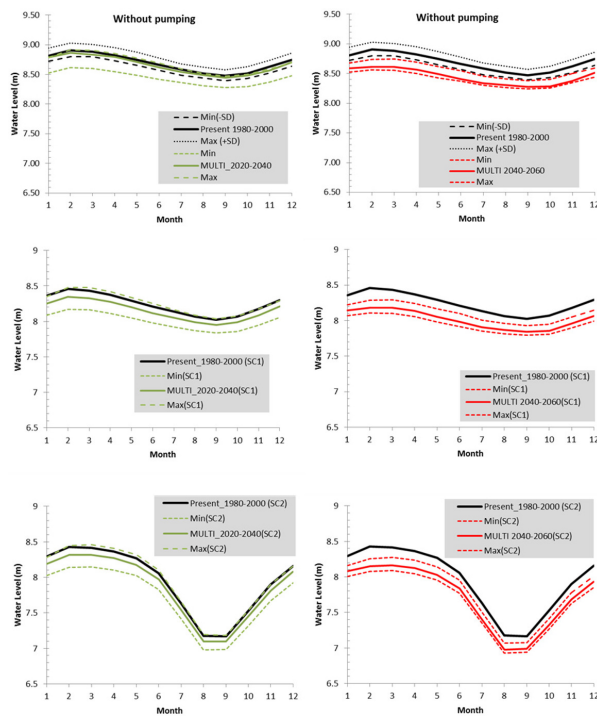


Figure 14. Variation of interannual monthly means of the piezometric levels simulated for the Argelès piezometer: reference period (1980–2000), short- (2020–2040) and medium- (2040–2060) term. The minimum and maximum envelope curves correspond to the extreme values across all scenarios. January = Month 1 and December = Month 12. Three different situations are simulated: (top) no pumping; (middle) with PP and without SP (SC1); and (bottom) with PP and SP (SC2).

10156