

1 **Title Page**

2 Effects of land use/land cover and climate changes on surface runoff in a
3 semi-humid and semi-arid transition zone in Northwest China

4 Jing Yin ¹, Fan He ², YuJiu Xiong ^{3,4,*}, GuoYu Qiu ^{5,*}

5 ¹ Research Center for Sustainable Hydropower Development, China Institute of
6 Water Resources and Hydropower Research, Beijing 100038, China.

7 ² State Key Laboratory of Simulation and Regulation of Water Cycle in River
8 Basin, China Institute of Water Resources and Hydropower Research, Beijing 100038,
9 China.

10 ³ Department of Water Resource and Environments, School of Geography and
11 Planning, Sun Yat-Sen University, Guangzhou 510275, Guangdong, China.

12 ⁴ Department of Land, Air and Water Resources, University of California at Davis.

13 ⁵ Shenzhen Engineering Laboratory for Water Desalinization with Renewable
14 Energy, School of Environment and Energy, Peking University, Shenzhen 518055,
15 Guangdong, China.

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First author Email address: yingjing@iwhr.com

* Corresponding author: YuJiu Xiong, Email address: xiongyuj@mail.sysu.edu.cn.
Tel./Fax: +86 20 84114575.

Co-corresponding author: GuoYu Qiu, Email address: qiugy@pkusz.edu.cn. Tel./Fax:
+86 755 26033309.

17 **Abstract**

18 Water resources, which are considerably affected by land use/land cover (LULC)
19 and climate changes, are a key limiting factor in highly vulnerable ecosystems in arid
20 and semi-arid regions. The impacts of LULC and climate changes on water resources
21 must be assessed in these areas. However, conflicting results regarding the effects of
22 LULC and climate changes on runoff have been reported in relatively large basins, such
23 as the Jinghe River Basin (JRB), which is a typical catchment ($> 45000 \text{ km}^2$) located in
24 a semi-humid and arid transition zone on the central Loess Plateau, Northwest China. In
25 this study, we focused on quantifying both the combined and isolated impacts of LULC
26 and climate changes on surface runoff. We hypothesized that under climatic warming
27 and drying conditions, LULC changes, which are primarily caused by intensive human
28 activities such as the Grain for Green Program, will considerably alter runoff in the JRB.
29 The Soil and Water Assessment Tool (SWAT) was adopted to perform simulations. The
30 simulated results indicated that although runoff increased very little between the 1970s
31 and the 2000s due to the combined effects of LULC and climate changes, LULC and
32 climate changes affected surface runoff differently in each decade, e.g., runoff increased
33 with increased precipitation between the 1970s and the 1980s (precipitation contributed
34 to 88% of the runoff increase). Thereafter, runoff decreased and was increasingly
35 influenced by LULC changes, which contributed to 44% of the runoff changes between
36 the 1980s and 1990s and 71% of the runoff changes between the 1990s and 2000s. Our
37 findings revealed that large-scale LULC under the Grain for Green Program has had an
38 important effect on the hydrological cycle since the late 1990s. Additionally, the

39 conflicting findings regarding the effects of LULC and climate changes on runoff in
40 relatively large basins are likely caused by uncertainties in hydrological simulations.

41 **Keywords:** SWAT; climate change; land use/land cover; streamflow; Jinghe River
42 Basin.

43

44 **1 Introduction**

45 Both climate and land use/land cover (LULC) changes are key factors that can
46 modify flow regimes and water availability (Oki and Kanae, 2006; Piao et al., 2007;
47 Sherwood and Fu, 2014; Wang et al., 2014a). Since the 20th century, climate variability
48 is believed to have led to changes in global precipitation patterns (IPCC 2007), thereby
49 changing the global water cycle and resulting in the temporal and spatial redistribution
50 of water resources (Milly et al., 2005; Murray et al., 2012). LULC changes are primarily
51 caused by human activities (Foley et al., 2005; Liu and Li, 2008) and affect the
52 partitioning of water among various hydrological pathways, including interception,
53 evapotranspiration, infiltration, and runoff (Sterling et al., 2012). The influences of
54 climate and LULC changes on hydrological processes and water resources **will likely**
55 **continue to increase**, especially in arid and semi-arid regions characterized as vulnerable
56 (Fu, 2003; Vorosmarty et al., 2010).

57 The impacts of LULC and climate changes on runoff can generally be identified by
58 using hydrological models (Praskievicz and Chang, 2009). These models provide
59 valuable frameworks for investigating the changes among various hydrological
60 pathways that are caused by climate and human activities (Leavesley, 1994; Jiang et al.,
61 2007; Wang et al., 2010). Distributed hydrological models, which use input parameters
62 that directly represent land surface characteristics, have been applied to assess the
63 impacts of LULC and climate changes on runoff in water resource management areas
64 (Yang et al., 2008; Yang et al., 2014; Chen et al., 2015). **The Soil and Water Assessment**
65 **Tool (SWAT), a robust, interdisciplinary, and distributed river basin model, is**

66 commonly used to assess the effects of management practices and land disturbances on
67 water quantity and quality (Gassman et al., 2007). The hydrological responses to LULC
68 and climate changes are often investigated through scenario simulations using the
69 SWAT model.

70 Although substantial progress has been made in assessing the impacts of LULC
71 and climate changes on water resources (Krysanova and Arnold, 2008; Vigerstol and
72 Aukema, 2011; Krysanova and White, 2015), most studies have focused on individual
73 factors (i.e., either LULC or climate); thus, the combined effects of LULC and climate
74 changes are not well understood because their contributions are difficult to separate and
75 vary regionally (Fu et al., 2007; D'Agostino et al., 2010; Wang et al., 2014a). For
76 example, some studies have suggested that surface runoff is affected more by climate
77 change (increased precipitation) than by LULC changes (Guo et al., 2008; Fan and
78 Shibata, 2015), and other studies have found that urbanization contributes more to
79 increased runoff than precipitation (Olivera and Defee, 2007). According to Krysanova
80 and White (2015), less than 30 papers were published between 2005 and 2014 on topics
81 related to the combined effects of LULC and climate changes and the SWAT model,
82 whereas 210 and 109 papers presented studies of climate and LULC changes,
83 respectively. However, water resource management requires an in-depth understanding
84 of the isolated and integrated effects of LULC and climate changes on runoff (Chawla
85 and Mujumdar, 2015).

86 Notable evidence of drying trends exists in semi-arid and semi-humid regions (Ma
87 and Fu, 2006; Li et al. 2007; Li et al. 2010; Li et al. 2011). These regions have

88 experienced serious water shortages in addition to intensive human activity and climate
89 change (Wang and Cheng, 2000; Ma and Fu, 2003). In this case, the effects of LULC
90 and climate changes on runoff are considerably more sensitive, and a dry climate can
91 result in serious environmental degradation and water crises (Ma et al., 2008; Jiang et
92 al., 2011; Leng et al. 2015). The Jinghe River Basin (JRB), which is located on the
93 central Loess Plateau, is a typical catchment located in a semi-humid and semi-arid
94 transition zone in Northwest China. The agricultural activities in this basin play an
95 important role in Northwest China (Zhao et al., 2014). However, the relative importance
96 of agriculture in the basin has caused ecological problems associated with social
97 development. For example, local water resources cannot maintain the rapid
98 socio-economic growth in the region (Wei et al., 2012), and the river system has
99 become unhealthy (Wu et al., 2014). Water and environmental management in the
100 region requires improved knowledge of the hydrological impacts of LULC and climate
101 changes. The effects of LULC and climate changes on the water cycle and water
102 resources must be assessed in these critical regions (Zhang et al., 2008; Li et al., 2009;
103 Qiu et al., 2011; Qiu et al., 2012; Peng et al., 2013).

104 Because the JRB transports the largest volume of sediment from the Loess Plateau
105 to the Yellow River, hydrological studies of the basin have primarily assessed the
106 impacts of soil and water conservation measures on surface runoff and sediment
107 transport (e.g., Feng et al., 2012; He et al., 2015; Peng et al., 2015a, 2015b; Wang et al.,
108 2016). Relatively few studies have been conducted regarding the effects of LULC and
109 climate changes on runoff. Studies of the Weihe River Basin (Zuo et al., 2014) and

110 Loess Plateau (Liang et al., 2015), which included the JRB as a sub-basin, have
111 identified the response of runoff to climate change and human activities by using a
112 climate elasticity model based on the Budyko framework. Zuo et al. (2014) found that
113 runoff in the JRB decreased by 17.79 mm between 1997 and 2009, with human
114 activities and climate change accounting for 51% and 39% of this decrease, respectively.
115 Liang et al. (2015) showed that streamflow decreased substantially from 1961 to 2009,
116 and the contribution of climate change (65%) to streamflow reduction was much larger
117 than that of ecological restoration measures (35%) in the JRB. Another study based on
118 the relationship between precipitation and runoff from 1966 to 1970 showed that runoff
119 mainly decreased due to precipitation before the 2000s and due to human activity
120 became dominant thereafter (with a contribution greater than 76%) (Zhang et al., 2011).
121 The different results reported by Zuo et al. (2014) and Liang et al. (2015) suggest that
122 assessing the impacts of LULC and climate changes on runoff in relatively large basins
123 (over 1000 km²) is difficult (Chawla and Mujumdar, 2015; Peng et al., 2015b) due to
124 their complex effects on streamflow (Fu et al., 2007) and the variable boundary
125 conditions (Chen et al., 2011; Niraula et al., 2015).

126 Therefore, the objectives of this study were as follows: 1) to assess the surface
127 runoff variability influenced by LULC and climate changes in recent decades in the JRB
128 by using the SWAT model, which differs from the climate elasticity model based on the
129 Budyko framework; 2) to quantify the combined and isolated impacts of LULC change
130 and climate variability on surface runoff in the basin from 1971 to 2005 by using
131 scenario simulations after calibrating and validating the SWAT model at monthly and

132 yearly time scales; 3) to discuss how LULC and climate changes affect surface runoff;
133 and 4) to discuss the simulation uncertainty in the context of SWAT modelling due to
134 parameterizations and provide potential explanations for the conflicting results
135 regarding the effects of LULC and climate changes on runoff in relatively large basins.

136 **2 Methods and materials**

137 **2.1 Study area**

138 The JRB, which covers an area of approximately 45421 km², is located at 106°14' –
139 108°42' E and 34°46' – 37°19' N on the central Loess Plateau in Northwest China (Fig.
140 1). The main stream of the Jinghe River, with a length of 450 km, originates in the
141 Liupan Mountains in the Ningxia Autonomous Region and flows across Gansu and
142 Shanxi Provinces before draining into the Weihe River. The outlet gauging station,
143 Zhangjiashan, has a control area of approximately 43216 km². The study area is
144 characterized by hills and syncline valleys, with the Liupan Mountains to the west and
145 the Ziwu Mountains to the east. The elevation decreases from 2900 m to 360 m above
146 sea level. The climate varies from sub-humid to semi-arid, with mean annual
147 precipitation, temperature, and pan evaporation values of 390–560 mm, 8–13 °C, and
148 1000–1300 mm, respectively. Precipitation mainly occurs between July and September,
149 accounting for 50–70% of the total annual rainfall.

150 **2.2 Runoff change simulation**

151 Under the assumption that runoff is affected only by LULC and climate changes, the
152 effects of LULC and climate changes on surface runoff were evaluated using SWAT.
153 Before the simulations, the SWAT model was calibrated and validated as described

154 below.

155 **2.2.1 SWAT model and data collection**

156 SWAT, a semi-distributed hydrological model, was developed to assess the impacts of
157 land management and climate on water, nutrient, and pesticide transport at the basin
158 scale (Arnold et al., 1998; Neitsch et al., 2005). SWAT simulates hydrological processes
159 such as surface runoff at the daily time scale based on information regarding weather,
160 topography, soil properties, vegetation, and land management practices. In SWAT, the
161 study basin is divided into sub-basins, and each sub-basin is further subdivided into
162 hydrological response units (HRUs) with homogeneous characteristics (e.g., topography,
163 soil, and land use). Hydrological components are then calculated in the HRUs based on
164 the water balance equation.

165 In this study, SWAT is operated via an interface in ArcView GIS (Di Luzio et al.,
166 2002). Therefore, the required data are either raster or vector data sets, including a
167 digital elevation model (DEM), soil properties, vegetation, LULC, meteorological
168 observations, and discharge observations at Zhangjiashan gauging station.

169 (1) DEM

170 The Shuttle Radar Topography Mission (SRTM) 90-m DEM (Jarvis et al., 2008)
171 was used in this study.

172 (2) Soil data

173 Soil property information was obtained from the soil map of China at a scale of
174 1:1000000. The map was provided by the Chinese Natural Resources Database.
175 Huangmiantu, which covers 75.10% of the basin area, is the major soil type in the area

176 according to the Genetic Soil Classification of China. The other seven soil types are
177 Heilutu (13.27%), Chongjitu (4.30%), Huihetu (3.23%), Hetu (2.41%), Hongniantu
178 (1.10%), Cugutu (0.35%), and Shandicaodiantu (0.24%).

179 (3) Vegetation and LULC data

180 LULC data from four periods were retrieved from Landsat images by supervised
181 classification, i.e., Multispectral Scanner (MSS) images (60 m resolution) from 1979,
182 Thematic Mapper (TM) images (30 m resolution) from 1989, and Enhanced Thematic
183 Mapper Plus (ETM+) images (30 m resolution) from 1999 and 2006. Each LULC
184 dataset represents the land use patterns for one decade (e.g., LULC data obtained from
185 1979 represents the land use patterns in the 1970s). Land use was classified into seven
186 categories: forest, dense grassland, sparse grassland, cropland, water, and barren areas.
187 Then, the accuracy of the classification was verified, yielding a minimum Kappa
188 coefficient of 0.73 (Xie et al., 2009).

189 (4) Meteorological data

190 Daily precipitation was collected from 16 rainfall stations (Fig. 1), whereas the
191 daily minimum and maximum temperatures, wind speed, and relative humidity data
192 required by the SWAT model were collected from 12 meteorological stations between
193 1970 and 2005. These data were interpolated to DEM grids using the SWAT model's
194 built-in weather generator, which describes the weather conditions in the model
195 simulations.

196 (5) Surface runoff

197 Daily runoff data measured at the Zhangjiashan gauging station between 1970 and

198 1990 were collected from the State Hydrological Statistical Yearbook. These data were
199 compared to the modelled surface flow during model calibration and validation.

200 2.2.2 Model calibration and validation

201 The SWAT model of the basin was first calibrated for the period of 1971 to 1997 and
202 was then validated for the period of 1981 to 1990. Based on published results (e.g., Li et
203 al., 2009) and our previous research results (Qiu et al., 2011), the simulation was the
204 most sensitive to the following six parameters: runoff curve number (CN₂), soil
205 evaporation compensation factor (ESCO), the available water capacity of the soil layer
206 (SOL_AWC), channel conductivity (CH_K₂), the baseflow alpha factor (ALPHA_BF),
207 and the surface runoff coefficient (SURLAG). Therefore, these six parameters were
208 calibrated in the SWAT model (Table 1) (Qiu et al., 2011). Model performance was
209 assessed qualitatively using visual time series plots and quantitatively using the
210 coefficient of determination (R²) and the Nash-Sutcliffe efficiency coefficient (*Ens*) (Eq.
211 (1)) (Moriasi et al., 2007).

$$212 \quad Ens = 1 - \frac{\sum_{i=1}^n (Q_{obs} - Q_{sim})^2}{\sum_{i=1}^n (Q_{obs} - Q_{obs_m})^2} \quad (1)$$

213 where Q_{obs} and Q_{sim} are the observed and modelled runoff, respectively; Q_{obs_m} is the
214 mean value of observed runoff; and n is the number of data records. When *Ens*
215 approaches 1, the model simulates the measured data more accurately, whereas a
216 negative *Ens* indicates that the model performance is poor. In this study, a criterion
217 proposed by Moriasi et al. (2007), the Nash-Sutcliffe coefficient, was adopted to
218 evaluate the simulation (Table 2).

219 The SWAT model was calibrated and validated based on annual and monthly river
220 discharges measured at the outlet gauging station shown in Fig. 1.

221 **2.2.3 Simulation scenarios**

222 In this study, the effects of LULC and climate changes on surface runoff were evaluated
223 by comparing the SWAT outputs of ten scenarios. Each scenario represented one decade,
224 and each simulation required an LULC map and a meteorological data set (Table 3). If
225 the LULC map and the meteorological data were within the same decade (i.e., the 1970s,
226 1980s, 1990s, or 2000s), the simulation results represented "real runoff" or a "baseline"
227 affected by the combination of LULC and climate changes. **Alternatively, varying one**
228 **driving factor while holding others constant simulated the effects of the variable factor**
229 **on runoff** (Li et al., 2009). For example, to assess the response of streamflow to
230 combined LULC and climate changes in the 1970s and 1980s, the simulation of the
231 1970s (1970–1979) ($Q_{base, i}$), which is used as a reference period or baseline, should be
232 based on the current LULC (year 1979) and current climate (years 1970–1979). The
233 simulation of the 1980s (1980–1989) ($Q_{base, i+1}$) should be based on future LULC (year
234 1989) and future climate (years 1980–1989). The difference between the first and
235 second simulations represents the combined effects of LULC and climate changes on
236 streamflow. Regarding LULC changes, the third simulation ($Q_{sim, cL, i}$) was based on the
237 current climate (years 1970–1979) and the LULC in the next period, or the future LULC
238 (in this example, 1989). The difference between the first and third simulations is the
239 effect of the LULC change on streamflow. Similarly, the difference between the first
240 simulation and the fourth simulation ($Q_{sim, cc, i}$) based on the current LULC (year 1979)

241 and future climate (in this example, 1980–1989) represents the impact of climate change
 242 on streamflow. The combined effects of LULC and climate changes on streamflow
 243 ($\Delta R_{comb}\%$) and the isolated effects of LULC ($\Delta R_{iso, cL}\%$) and climate ($\Delta R_{iso, cc}\%$) can be
 244 assessed using Eqs. (2) to (4).

$$245 \quad \Delta R_{comb}\% = \left(\frac{Q_{base,i+1} - Q_{base,i}}{Q_{base,i}} \right) \times 100 \quad (2)$$

$$246 \quad \Delta R_{iso, cL}\% = \left(\frac{Q_{sim, cL,i} - Q_{base,i}}{Q_{base,i}} \right) \times 100 \quad (3)$$

$$247 \quad \Delta R_{iso, cc}\% = \left(\frac{Q_{sim, cc,i} - Q_{base,i}}{Q_{base,i}} \right) \times 100 \quad (4)$$

248 **3 Results**

249 **3.1 Climate change**

250 Variations in precipitation, dryness index (E_0/P , defined as the ratio of annual potential
 251 evapotranspiration calculated using the Penman–Monteith method to annual
 252 precipitation), and air temperature were evaluated over four decades based on
 253 meteorological data from 1970 to 2009 (Fig. 2). Precipitation decreased by 3.4% from
 254 the 1970s to the 2000s. However, precipitation in the 1980s was slightly higher than that
 255 in the 1970s. The decreasing trend in precipitation was substantial from the 1980s to the
 256 1990s, reaching 4.1%. Thereafter, the decrease in precipitation was less than that from
 257 1980 to 1999. During the entire period (from the 1970s to the 2000s), the temperature
 258 increased by 13.6% (1.18 °C), including an abrupt increase of 0.7 °C from the 1980s to
 259 the 1990s. Although the dryness index exhibited little change (increasing by 1.8%), a
 260 large dryness index (>1.9) indicates that the climate became drier. These results indicate
 261 that the climate in the JRB changed dramatically over the last four decades, as

262 characterized by decreased precipitation and increased temperature and dryness index
263 values. Both warming and drying trends are evident in the JRB. These results agree with
264 the results of other studies that reflect a broader phenomenon known as “climatic
265 warming and drying” in northern China (Ma and Fu, 2003; Huang et al., 2012).

266 **3.2 LULC change**

267 Figure 3 shows the variations in LULC distributions over the last four decades. The
268 dominant land-use types are sparse grassland (with a vegetation coverage of $< 20\%$) and
269 cropland, which encompass a total of $> 61\%$ of the area over the four decades. However,
270 the percentage of sparse grassland was slightly higher than that of cropland, and the
271 margin varied from 2.96% to 9.80%. The remaining types include dense grassland (with
272 a vegetation coverage of $\geq 20\%$), forest, barren areas, urban and built-up areas, and
273 water, with mean ratios of 17.57%, 13.71%, 6.35%, 0.31%, and 0.29%, respectively.
274 The vegetation with low coverage that is predominant in the study basin corresponds
275 with the regional climate, and the relatively high percentage of cropland indicates the
276 importance of agriculture in this area.

277 The statistical results illustrated by the four LULC maps over the last four decades
278 indicate that vegetation (including grassland and forest) decreased by 11% between the
279 1970s and the 1990s and increased by 6% thereafter. The areas of cropland and urban
280 and built-up areas increased by 4.03% and 0.95%, respectively, over time. The area of
281 water fluctuated slightly, increasing by 0.09%. The area of barren land increased from
282 3.09% to 12.35% before the 1990s but then decreased to 3.02% in the 2000s. The
283 LULC changes potentially resulted from two major factors: social development and

284 population growth. These factors have increased since the 1980s, leading to the
285 expansion of urban and agricultural activities as well as unreasonable land utilization,
286 reclamation of vulnerable land, and vegetation removal. Therefore, the areas of urban
287 and barren land increased while the area of vegetation decreased. However, the
288 decreasing trend in vegetation changed due to a nationwide environmental conservation
289 programme initiated in 1999 by the Chinese government, the **Grain for Green Program**
290 **(GGP)** (Xu et al., 2004). The main goal of the **GGP** was to reduce soil erosion and
291 improve the eco-environmental status of western and northern China (Xu et al., 2004).
292 Noticeable evidence of ecological restoration was observed on the Loess Plateau after
293 the **GGP** was implemented (Chang et al., 2011; Sun et al., 2015). In addition to
294 preventing soil erosion, the **GGP** improved the soil physical and chemical properties
295 (Deng et al., 2014; Song et al., 2014) and facilitated vegetation restoration. The results
296 indicate that vegetation increased since the late 1990s, and these results agree with the
297 results of other studies (e.g., Liang et al., 2015; Wang et al., 2016).

298 **3.3 Performance of the SWAT model**

299 The SWAT model performed well in both the calibration and validation periods,
300 accurately simulating the outlet flows according to the model performance criteria (R^2
301 and *Ens*) after the six sensitive parameters were optimized. During the calibration
302 period (1971–1980), the time series plots of simulations and observations were similar
303 at both the annual (Fig. 4 (a)) and monthly scales (Fig. 5 (a)), although overestimation
304 was observed in the simulated streamflow. Point-by-point comparisons between the
305 simulations and observations further showed that most of the paired streamflow values

306 were distributed near the 1:1 line, with mean R^2 values of 0.90 (Fig. 4 (b)) and 0.84 (Fig.
307 5 (b)) at the annual and monthly scales, respectively (Qiu et al., 2011). In addition, the
308 results of a statistical analysis indicated that the mean *Ens* values were 0.76 and 0.72 at
309 the annual and monthly scales, respectively (Table 4). Similarly, although the SWAT
310 model did not perform as well during the validation period (1981–1990) relative to the
311 calibration period, the performance was still adequate, with *Ens* (R^2) values of 0.73
312 (0.83) and 0.69 (0.77) at the annual and monthly scales, respectively (Table 4, Figs. 6
313 and 7).

314 Although the *Ens* performance statistic associated with SWAT runoff modelling
315 can be larger than 0.8 in small or humid basins (e.g., Luo et al., 2008; Qiao et al., 2015;
316 Wu et al., 2016), *Ens* is typically less than 0.7 in relatively large river basins in arid to
317 semi-arid regions (e.g., Xu et al., 2011; Notter et al., 2013; Zhang et al., 2015; Liu et al.,
318 2016; Zhao et al., 2016). The *Ens* values in this study were generally good in the
319 calibration and validation periods and were comparable to those reported in other
320 studies in arid to semi-arid river basins. The results suggested that the SWAT model
321 performed well and was applicable to the study basin.

322 **3.4 Simulated surface runoff**

323 The annual runoff simulated by SWAT under different scenarios is shown in Table 3.
324 Generally, runoff increased minimally between the 1970s and the 2000s at a rate of 1.51
325 $\text{m}^3 \text{s}^{-1}$ (simulations S1 and S10) due to the combined effects of LULC and climate
326 changes (Fig. 8). However, runoff changed differently in different decades. For example,
327 runoff increased by 35.4% ($29.75 \text{ m}^3 \text{ s}^{-1}$) from the 1970s to the 1980s (simulations S1

328 and S4) but decreased thereafter. Notably, the simulated runoff in the 1990s was 12.59
329 $\text{m}^3 \text{s}^{-1}$ less than that in the 1980s (simulations S4 and S7), and runoff decreased by
330 15.5% ($15.65 \text{ m}^3 \text{ s}^{-1}$) from the 1990s to the 2000s (simulations S7 and S10) (Table 3).

331 **4 Discussion**

332 **4.1 Impacts of LULC and climate changes on surface runoff**

333 The hydrological effects were analysed using the simulated runoff data rather than the
334 observed data. The combined effects of LULC and climate changes on surface runoff
335 are presented in section 3.4. The simulated runoff increased between the 1970s and the
336 1980s, while precipitation increased from 521 mm to 527 mm during the same period.
337 Thereafter, runoff decreased as precipitation decreased. However, runoff decreased by
338 11.1% from the 1980s to the 1990s but decreased by 15.5% from the 1990s to the 2000s.
339 These results indicate that, although precipitation can considerably affect runoff
340 simulation, variations in runoff and precipitation were nonlinear due to the combined
341 effects.

342 The isolated impacts of LULC and climate changes on surface runoff can be
343 analysed by comparing two sets of simulations. The differences between S1 and S2 (as
344 well as between S4 and S5 and S7 and S8) reflect the impacts of climate change on
345 runoff. Accordingly, the differences between S1 and S3 (as well as between S4 and S6
346 and S7 and S9) reflect the impacts of climate change on runoff.

347 **4.1.1 Isolated impacts of LULC change**

348 During the first two decades, LULC changes increased runoff by $2.30 \text{ m}^3 \text{ s}^{-1}$ and
349 accounted for 7.73% of the total change ($29.75 \text{ m}^3 \text{ s}^{-1}$). Thereafter, LULC change

350 decreased runoff by $6.83 \text{ m}^3 \text{ s}^{-1}$, which accounted for 54.25% of the total change in
351 runoff ($12.59 \text{ m}^3 \text{ s}^{-1}$) from the 1980s to the 1990s. The impacts of LULC changes on
352 runoff increased during the last two decades because the contribution of LULC changes
353 to runoff increased to 70.67% from the 1990s to the 2000s (Fig. 9).

354 The results in section 3.2 show that the LULC changed slightly from the 1970s to
355 the 1980s. For example, the area of cropland marginally increased by 0.76%, and the
356 vegetative area decreased by 3.19%. This small LULC change indicates that human
357 activities minimally influenced runoff during the first two decades because the LULC
358 changes only accounted for 7.73% of the increase in runoff. However, the LULC
359 changed considerably with social development and population growth beginning in the
360 1980s. The vegetative area decreased by 7.81% from the 1980s to the 1990s, and the
361 percentages of cropland, barren areas, and urban and built-up areas increased by 2.39%,
362 5.43%, and 0.11%, respectively. LULC changes associated with increased human
363 activities accounted for 54.25% of the increase in surface runoff. Furthermore, the GGP,
364 which was initiated in the late 1990s, mitigated the decreasing trend in vegetation.
365 Although cropland and urban and built-up areas still expanded by 2.40% and 0.82%,
366 respectively, from the 1990s to the 2000s, vegetation increased by 6.00%, and barren
367 areas decreased by 9.33%. Therefore, LULC change exhibited a relatively large
368 influence on the surface runoff change, contributing to 70.67% of the surface runoff in
369 the last two decades.

370 In addition, the spatial distributions of different land use types influence the
371 generation of runoff. As reported in our previous publication (Qiu et al., 2011), the soil

372 moisture content and evapotranspiration were modified by LULC changes (i.e., the
373 GGP) after the GGP in the JRB, which led to changes in surface runoff. However, the
374 modification was different. Fig. 10 shows that, after the GGP, the soil moisture content
375 increased in the three selected sub-basins from the upstream to downstream regions,
376 while the runoff and evapotranspiration decreased. When considering the upstream area
377 as an example, barren land, with an initial percentage of 15.90%, and partial farmland,
378 with an initial percentage of 6.56%, were converted to grassland due to the GGP, which
379 improved water filtration and increased the soil moisture (Fig. 10 (a)). The simulation in
380 Fig. 10 shows that the soil moisture content increased by 163.66%, 208.23%, and
381 262.66% in the sub-basins from the upstream to downstream, whereas the surface runoff
382 (evapotranspiration) decreased by -37.53%, -38.55%, and -49.01% (-1.21%, -3.06%,
383 and -25.90%), respectively. These results indicate that the impacts of LULC changes on
384 flow regimes were larger in the downstream areas of the basin than in the upstream
385 areas.

386 Although climate variables were held constant when simulating LULC changes,
387 the isolated influences of LULC changes on runoff did not exclude the impacts of
388 precipitation variations because the climate (including precipitation) varied in each
389 decade (Table 3). Nonetheless, the above results indicate that LULC changes
390 contributed considerably to decreased runoff, as reported in previous studies (e.g.,
391 Zhang et al., 2011; Zuo et al., 2014; Wang et al., 2014b; Wang et al., 2016). Additionally,
392 the results suggest that vegetation restoration due to the GGP reduced surface runoff,
393 which agrees with the results of other studies (e.g., Li et al., 2009; Nunes et al., 2011).

394 **4.1.2 Isolated impacts of climate change**

395 Unlike the contributions of LULC changes, the influences of climate change decreased
396 in recent decades (Fig. 9). Climate change increased runoff by $26.07 \text{ m}^3 \text{ s}^{-1}$ from the
397 1970s to the 1980s, accounting for approximately 87.63% of the increased total runoff
398 during that period. Since the 1980s, surface runoff decreased, and the contributions of
399 climate change to decreased runoff were 55.92% and 42.11% from the 1980s to the
400 1990s and from the 1990s to the 2000s, respectively. The influence of climate change on
401 runoff agrees with climatic warming and drying trends. Decreasing precipitation will
402 potentially lead to less runoff, whereas increasing temperatures will result in increased
403 evaporation.

404 In summary, LULC and climate changes accounted for 7.73% and 87.63% of the
405 total runoff increase ($29.75 \text{ m}^3 \text{ s}^{-1}$) in the 1970s and 1980s, respectively. The isolated
406 influences of LULC and climate changes on runoff were nearly the same from 1980 to
407 1999 (54.25% and 55.92%, respectively) compared to the total decrease in runoff. In the
408 last two decades, the percentage of the total runoff decrease that was caused by LULC
409 changes (70.67%) was greater than that caused by climate change (42.11%).

410 Although uncertainties exist in the simulations (see section 4.2 for details), the
411 above results indicate that the contribution of climate variability decreased over the last
412 four decades, while the contribution of LULC change increased. Unlike the results
413 reported by Liang et al. (2015), the findings in this study suggested that runoff
414 fluctuations are influenced less by climate change and more by human activities. The
415 results also indicate that the impacts of human activities on runoff have gradually

416 increased in the JRB, which agrees with the results of other studies (Zhang et al., 2011;
417 Zuo et al., 2014; Wang et al., 2016).

418 **4.2 Uncertainty in SWAT model simulations**

419 Uncertainty in model simulations, which is mainly caused by model structure (e.g.,
420 algorithm limitations) and model parameterizations, is a major challenge when
421 assessing the impacts of LULC and climate changes on runoff in relatively large basins.
422 In this study, the SWAT model performed well, with a Nash-Sutcliffe efficiency
423 coefficient and coefficient of determination of 0.76 and 0.90, respectively, for annual
424 runoff during the calibration period, as well as values of 0.73 and 0.83, respectively,
425 during the validation period. However, under the assumption that runoff is affected only
426 by LULC or climate changes, the simulated runoff associated with changes in only one
427 driving factor was slightly different than the simulated runoff obtained when
428 considering the combined effects of both factors due to the uncertainty in representing
429 LULC and climate change interactions in the SWAT model. For example, $28.37 \text{ m}^3 \text{ s}^{-1}$,
430 which was the combined runoff rate in S2 and S3, was not equal to the "real or baseline
431 runoff" of $29.75 \text{ m}^3 \text{ s}^{-1}$ in S4.

432 Qiao et al. (2015) reported that the SWAT model performed much better in small
433 watersheds (2–5 ha) than in a larger watershed (78 km^2) because the meteorological
434 inputs (e.g., precipitation) do not represent the spatial variability in a given parameter
435 over larger basins because ground-based observations are limiting. To reduce the
436 uncertainty and improve the accuracy of the hydrological model and forecasting results
437 for relatively large basins, **the uncertainty associated with model parameterization is**

438 discussed below and potential solutions are proposed for future studies.

439 In this study, the basin area exceeded 45000 km². However, only 16 rainfall
440 stations were available, among which six stations were outside the study basin. The
441 station density was 0.35 stations per 1000 km². Xu et al. (2013) found that model
442 simulations are influenced by rainfall station densities below 0.4 per 1000 km². Under
443 such conditions, runoff simulations may contain uncertainties due to poor representation
444 of spatial precipitation variability, which is crucial in determining the runoff hydrograph
445 (Singh, 1997). Previous studies (e.g., Chu et al., 2011; Masih et al., 2011; Shope and
446 Maharjan, 2015) have suggested that the density of rainfall measurement stations has a
447 significant impact on SWAT simulations and that reducing the precipitation uncertainty
448 can improve the accuracy of simulated streamflows. Although the SWAT model
449 performed well in this study and the uncertainty in the simulations associated with
450 precipitation was similar to the uncertainties observed in other studies, peak flow
451 overestimation was observed in the simulated runoff (Figs. 4 to 7). To reduce
452 uncertainty, precipitation from stations should be processed (e.g., via interpolation)
453 before conducting runoff simulations, thereby improving the precision and spatial
454 representativeness, especially in relatively large basins without reliable and precise areal
455 rainfall data.

456 In addition, the coarse vegetation information provided by the LULC data in this
457 study can lead to uncertainty in the simulations because vegetation distinction is
458 required in SWAT modelling. Although the LULC data had a relatively high resolution
459 of 30 m, we can only provide a general vegetation categorization, such as forest, due to

460 the data limitations. Recent results (e.g., Pierini et al., 2014; Qiao et al., 2015) have
461 shown that detailed biophysical parameters of vegetation species can improve the
462 performance of distributed, physically based models such as SWAT and reduce model
463 uncertainty. In China, detailed and reliable data related to vegetation species are
464 uncommon. Reliable maps of vegetation species (as well as other geographic maps) at
465 high spatial resolutions (e.g., <1000 m) are an urgently needed to provide detailed and
466 heterogeneous information for accurate biophysical and hydrological parameterization.

467 **5 Conclusions**

468 In this study, the SWAT model was used to simulate the effects of LULC and climate
469 changes on surface runoff. The satisfactory performance of the SWAT model was
470 confirmed by the Nash-Sutcliffe coefficient and coefficient of determination values of
471 annual runoff of 0.76 and 0.90, respectively, during the calibration period and 0.73 and
472 0.83, respectively, during the validation period. Simulations showed that the combined
473 effects of LULC and climate changes increased surface runoff by $29.75 \text{ m}^3 \text{ s}^{-1}$ during
474 the 1970s and the 1980s, whereas LULC and climate changes both decreased runoff by
475 $28.24 \text{ m}^3 \text{ s}^{-1}$ during the 1980s and the 2000s. Further analysis suggested that different
476 driving factors had different influences on surface runoff.

477 The isolated results indicated that the impacts of LULC changes on the
478 hydrological cycle were gradual, and that LULC changes altered runoff to a similar or
479 greater extent than climate change, accounting for 70.67% of the streamflow reduction
480 since the late 1990s. This result suggests that LULC plays an important role in the
481 transition zone between semi-humid and semi-arid regions. As an indicator that is

482 closely related to human activities, the LULC in the study area underwent considerable
483 changes, especially the vegetation cover rate, which decreased by 16% from the 1970s
484 to the 1990s and increased by 6% between the 1990s and the 2000s due to the **Grain for**
485 **Green Program (GGP)**. In conclusion, the increased vegetation and land use changes
486 inevitably altered the hydrological cycle, and large-scale LULC changes under the **GGP**
487 considerably affected the hydrological cycle.

488 To reduce simulation uncertainty and improve the accuracy of hydrological
489 modelling and forecasting in relatively large basins, areal input parameters (e.g.,
490 precipitation and **vegetation species information**) **should be generated with reliable**
491 **precision and high spatial resolution.**

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747

748 **Table 1.** Calibrated values of the six parameters in SWAT

No.	Parameter name	Description	Range	Calibrated value
1	CN ₂	SCS runoff curve number for moisture condition II	-8--+8	-8
2	ESCO	Soil evaporation compensation factor	0-1	0.1
3	SQL_AWC	Available water capacity of the soil layer	0-1	0.05
4	CH_K ₂	Channel conductivity	0-150	0.35
5	ALPHA_BF	Baseflow alpha factor	0-1	0.01
6	SURLAG	Surface runoff coefficient	0-10	0.85

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Table 2. SWAT performance of runoff simulations according to the Nash–Sutcliffe coefficient (Moriassi et al., 2007).

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Simulation performance	Nash–Sutcliffe coefficient (<i>Ens</i>)
Very good	$0.75 < Ens \leq 1.00$
Good	$0.65 < Ens \leq 0.75$
Satisfactory	$0.50 < Ens \leq 0.65$
Unsatisfactory	$Ens \leq 0.50$

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Table 3. Simulated annual runoff by SWAT under different scenarios considering

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both LULC and climate.

	Scenarios	Climate	LULC	Simulation ($\text{m}^3 \text{s}^{-1}$)	Runoff change ($\text{m}^3 \text{s}^{-1}$)	Runoff change (%)
	LULC and					
S1	meteorological data from the 1970s	1970s	1970s	84.10	–	–
	Changing LULC while					
S2	holding climate constant	1970s	1980s	86.40	+2.30	+7.73
	Changing climate while					
S3	holding LULC constant	1980s	1970s	110.17	+26.07	+87.63
	LULC and					
S4	meteorological data from the 1980s	1980s	1980s	113.85	+29.75	–
	Changing LULC while					
S5	holding climate constant	1980s	1990s	107.02	-6.83	-54.25
	Changing climate while					
S6	holding LULC constant	1990s	1980s	108.61	-7.04	-55.92
	LULC and					
S7	meteorological data from the 1990s	1990s	1990s	101.26	-12.59	–
	Changing LULC while					
S8	holding climate constant	1990s	2000s	90.20	-11.06	-70.67
	Changing climate while					
S9	holding LULC constant	2000s	1990s	94.67	-6.59	-42.11
	LULC and					
S10	meteorological data from the 2000s	2000s	2000s	85.61	-15.65	–

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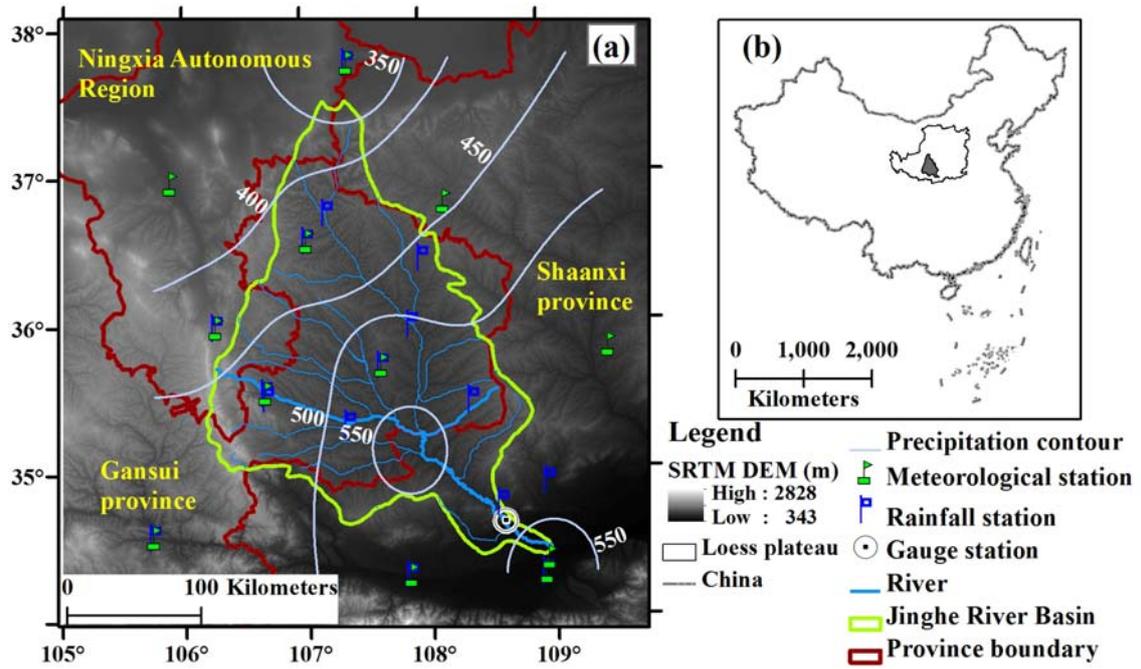
Table 4. Nash-Sutcliffe coefficient (*Ens*) statistics in the SWAT calibration and validation periods.

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Statistic	Calibration from		Validation from	
	1971–1980		1981–1990	
	monthly	yearly	monthly	yearly
<i>N</i>	120	10	120	10
Minimum	0.58	0.53	0.54	0.58
Maximum	0.95	0.98	0.81	0.9
Mean	0.72	0.76	0.69	0.73

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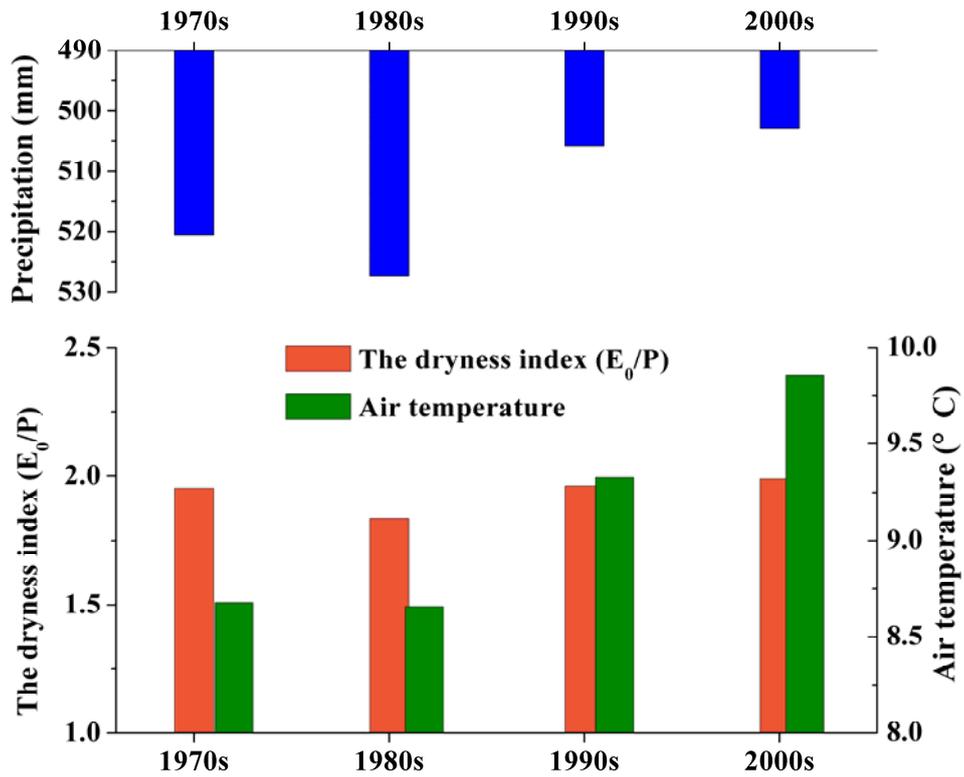
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Figure 1. Geographic information regarding the study area: (a) Location and SRTM DEM of the Jinghe River Basin and (b) schematic of the selected study area in China. Precipitation (mm) is averaged and interpolated from meteorological data between 1970 and 2010.



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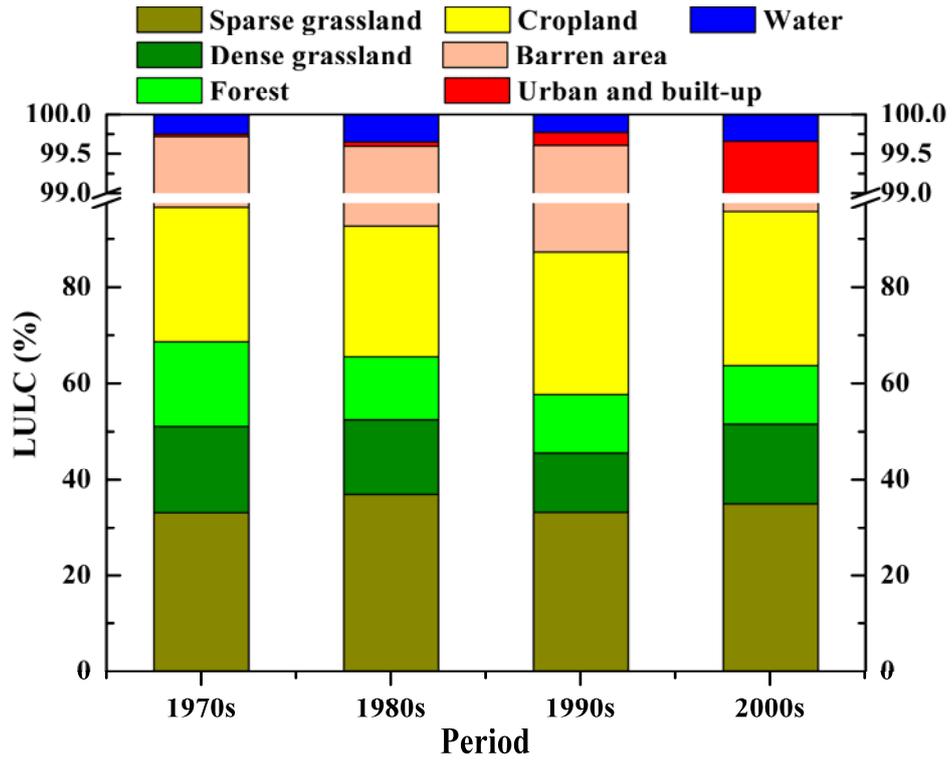
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Figure 2. Variation in decadal mean precipitation (top), dryness index, and air temperature (bottom) in the Jinghe River Basin from the 1970s to the 2000s. The dryness index was defined as the ratio of annual potential evapotranspiration (E_0) to annual precipitation (P).



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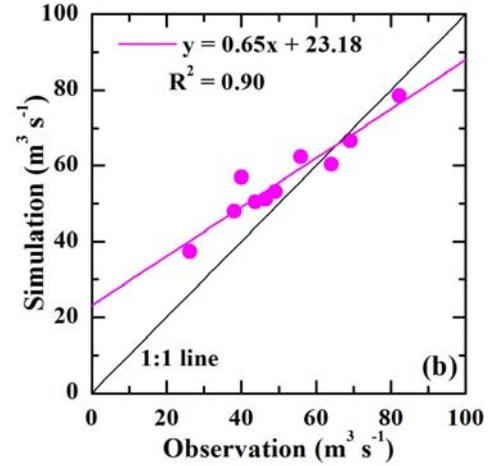
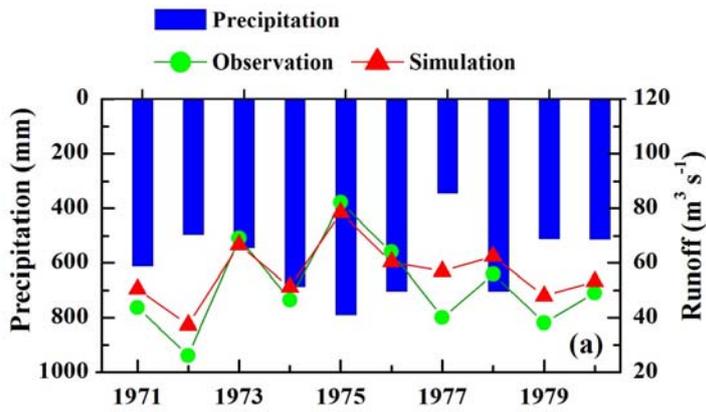
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Figure 3. LULC composition and its change in the Jinghe River Basin from the 1970s

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to the 2000s.

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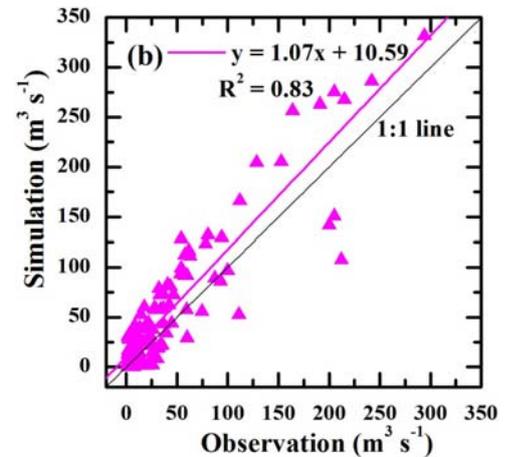
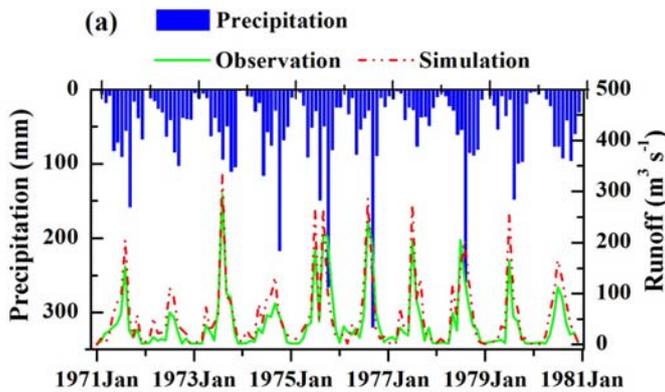
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Figure 4. Comparison of observed and simulated runoff at the yearly scale in the Jinghe River Basin during the calibration period from 1971 to 1980. Fig. 4(b) is redrawn from Qiu et al. (2011).



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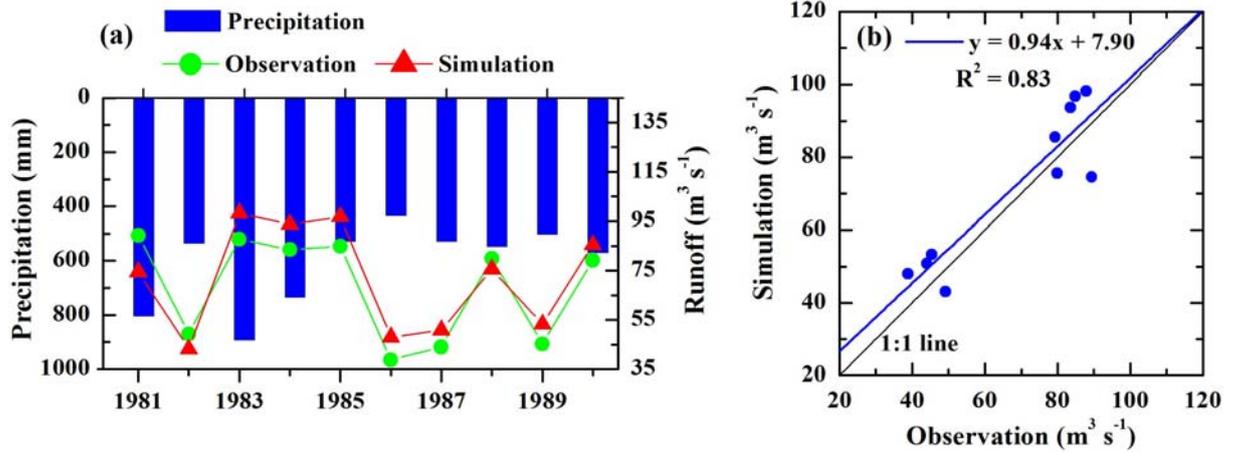
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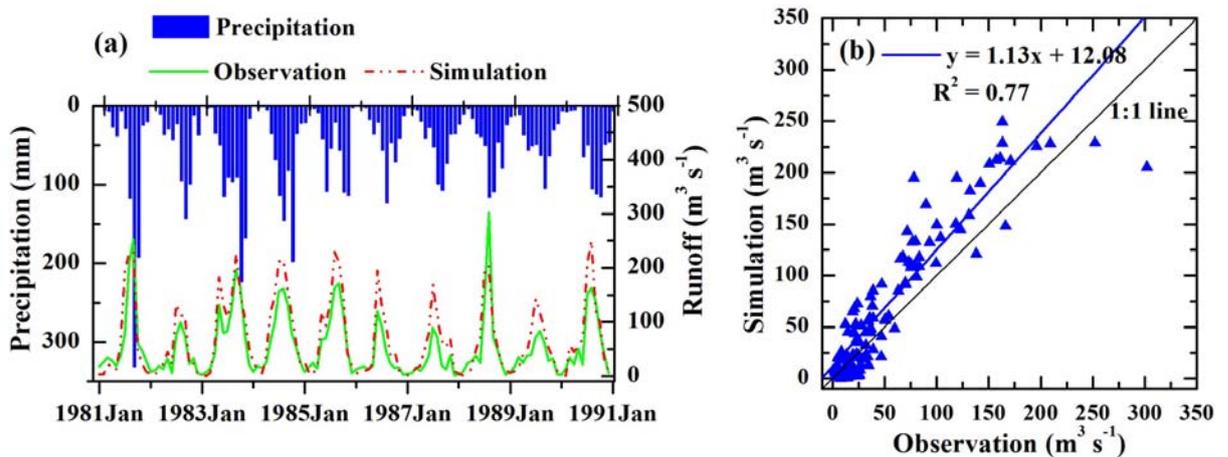
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Figure 5. Comparison of observed and simulated runoff at the monthly scale in the Jinghe River Basin during the calibration period from 1971 to 1980. Fig. 5(b) is redrawn from Qiu et al. (2011).



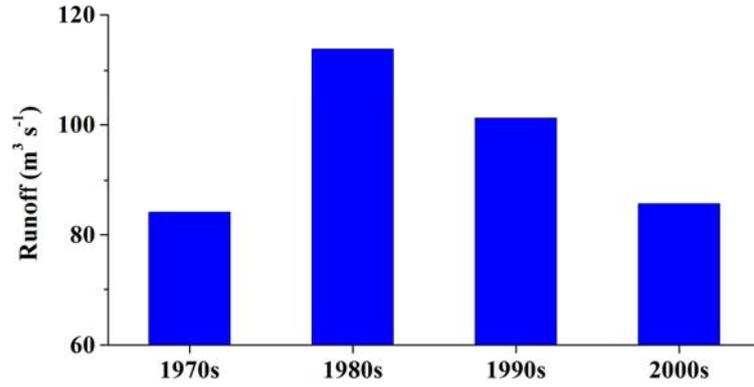
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Figure 6. Comparison of observed and simulated runoff at the yearly scale in the Jinghe River Basin during the validation from 1981 to 1990. Fig. 6(b) is redrawn from Qiu et al. (2011).



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Figure 7. Comparison of observed and simulated runoff at the monthly scale in the Jinghe River Basin during the validation period from 1981 to 1990. Fig. 7(b) is redrawn from Qiu et al. (2011).



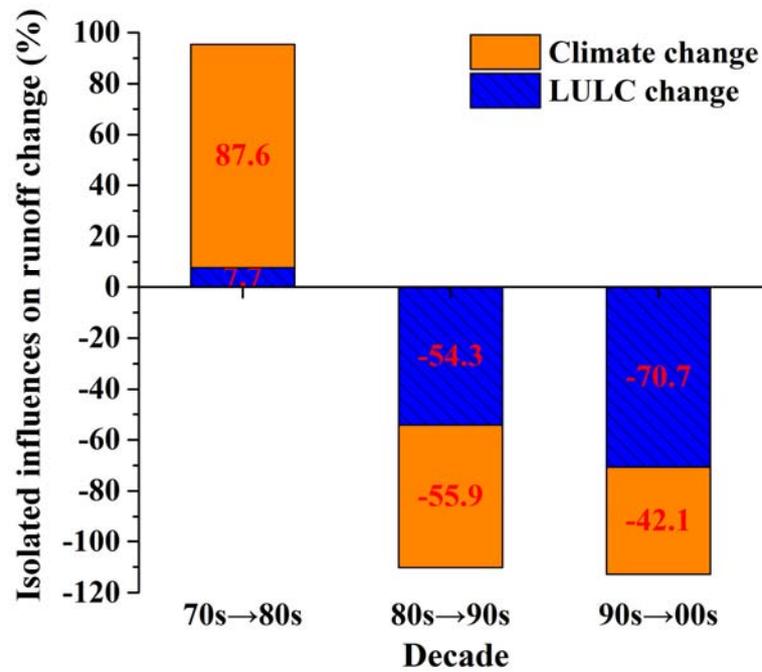
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Figure 8. Variation in mean annual surface runoff at the decadal scale in the Jinghe River Basin from the 1970s to the 2000s.

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Figure 9. Isolated impacts of LULC and climate changes on surface runoff. Positive

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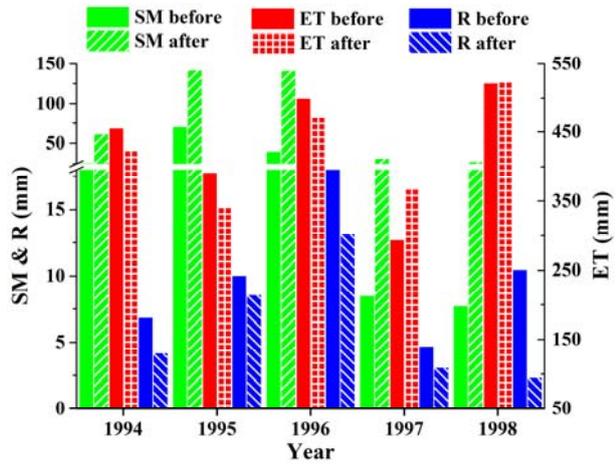
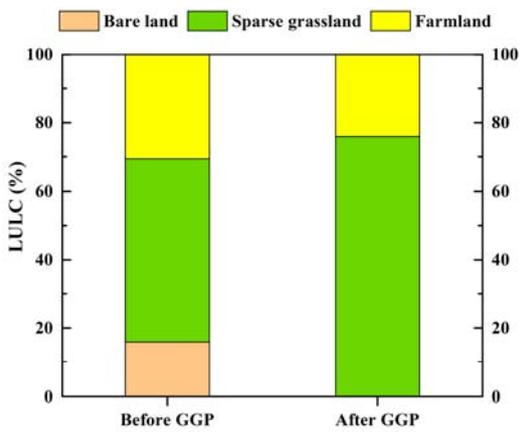
values indicate that runoff increased due to these factors, whereas negative values

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indicate that runoff decreased due to these factors. The summation of the isolated

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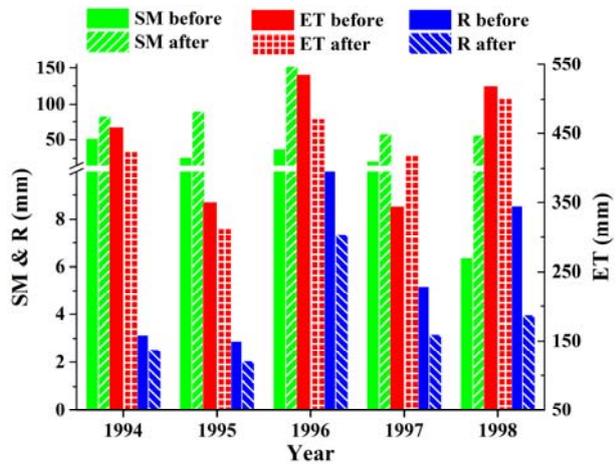
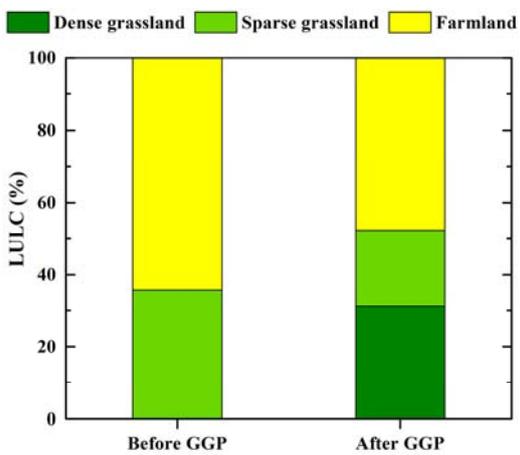
influences is not equal to 100% due to simulation uncertainty (see section 4.2 for details).



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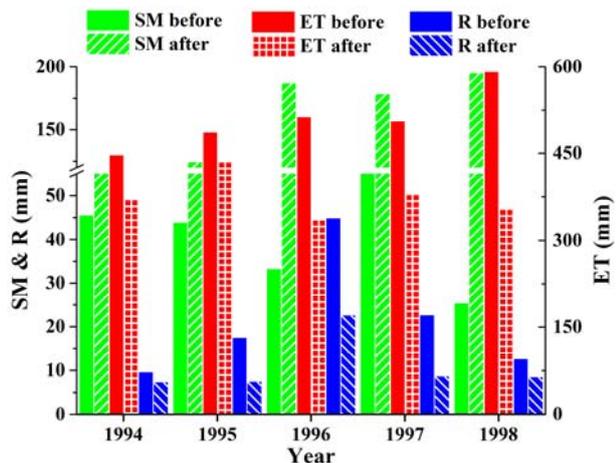
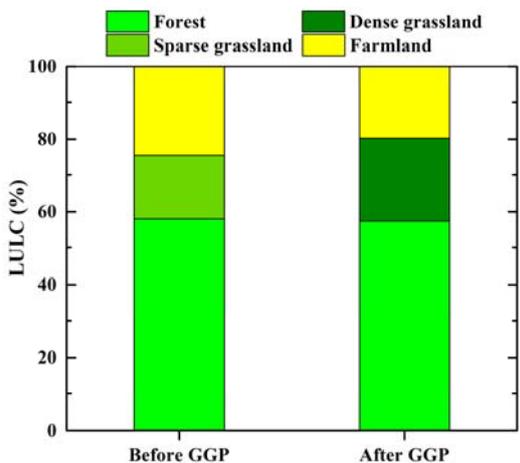
(a) upstream



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(b) midstream



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(c) downstream

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Figure 10. Impact of LULC changes on surface runoff in selected sub-basins distributed in the upstream, midstream, and downstream areas of the basin. The left column shows the land use types and corresponding ratios, and the right column shows the simulated changes of the soil moisture content (SM), evapotranspiration (ET), and

819 surface runoff (R) before and after the Grain for Green Program (GGP) scenarios while
820 holding climate constant.