



Urbanization
dramatically altered
the water balances

L. Hao et al.

This discussion paper is/has been under review for the journal Hydrology and Earth System Sciences (HESS). Please refer to the corresponding final paper in HESS if available.

Urbanization dramatically altered the water balances of a paddy field dominated basin in Southern China

L. Hao¹, G. Sun², Y. Liu³, J. Wan⁴, M. Qin¹, H. Qian¹, C. Liu⁵, R. John⁶, P. Fan⁷, and J. Chen⁶

¹International Center for Ecology, Meteorology, and Environment (IceMe), Jiangsu Key Laboratory of Agricultural Meteorology, Nanjing University of Information Science and Technology (NUIST), Nanjing 210044, China

²Eastern Forest Environmental Threat Assessment Center, Southern Research Station, USDA Forest Service, Raleigh, NC 27606, USA

³Center for Forest Disturbance Science, Southern Research Station, USDA Forest Service, Athens, GA 30602, USA

⁴China Institute of Water Resources and Hydropower Research, Beijing 100048, China

⁵State Key Laboratory for Information Engineering in Surveying, Mapping and Remote Sensing, Wuhan University, Wuhan 430079, China

⁶Center for Global Change and Earth Observations (CGCEO), and Department of Geography, Michigan State University, East Lansing, MI 48823, USA

⁷School of Planning, Design, and Construction (SPDC) and Center for Global Change and Earth Observations (CGCEO), Michigan State University, East Lansing, MI 48823, USA

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Received: 14 January 2015 – Accepted: 1 February 2015 – Published: 13 February 2015

Correspondence to: G. Sun (gesun@ncsu.edu)

Published by Copernicus Publications on behalf of the European Geosciences Union.

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)



[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



Abstract

Rice paddy fields provide important ecosystem services (e.g., food production, water retention, carbon sequestration) to a large population globally. However, these benefits are declining as a result of rapid environmental and socioeconomic transformations characterized by population growth, urbanization, and climate change in many Asian countries. This case study examined the responses of streamflow and watershed water balances to the decline of rice paddy fields due to urbanization in the Qinhuai River Basin in southern China where massive industrialization has occurred in the region during the past three decades. We found that streamflow increased by 58 % and evapotranspiration (ET) decreased by 23 % during 1986–2013 as a result of an increase in urban areas of three folds and reduction of rice paddy field by 27 %. Both highflows and lowflows increased significantly by about 28 % from 2002 to 2013. The increases in streamflow were consistent with the decreases in ET and leaf area index monitored by independent remote sensing MODIS data. The reduction in ET and increase in streamflow was attributed to the large cropland conversion that overwhelmed the effects of regional climate warming and climate variability. Converting traditional rice paddy fields to urban use dramatically altered land surface conditions from a water-dominated to a human-dominated landscape, and thus was considered as one of the extreme types of contemporary hydrologic disturbances. The ongoing large-scale urbanization in the rice paddy-dominated regions in the humid southern China, and East Asia, will likely elevate stormflow volume, aggravate flood risks, and intensify urban heat island effects. Understanding the linkage between land use change and changes in hydrological processes is essential for better management of urbanizing watersheds.

1 Introduction

Urbanization is a global phenomenon that poses profound threats to the local environment, society, and culture (Foley et al., 2005; McDonald et al., 2011). The most obvious

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



5 direct consequence of urbanization is the altered hydrology and water balances that control the flows of energy and matter in watershed ecosystems (Paul and Meyer, 2001; Sun and Lockaby, 2012). It is widely known that urbanization elevates peakflow rates (Brath et al., 2006; Du et al., 2012; Sun and Lockaby, 2012) as a result of increased impervious surfaces that promote quick surface runoff (Dietz and Clausen, 2008; Miller et al., 2014). However, the hydrologic response to urbanization is extremely variable (Jacobson, 2011; Caldwell et al., 2012) due to climatic differences and land use change patterns across a watershed (Sun and Lockaby, 2012). Empirical data are still lacking about changes in water balances and watershed hydrologic characteristics other than stormflows, such as total flows, lowflows, and evapotranspiration (ET) (Dow and De-Walle, 2000; Boggs and Sun, 2011) in different physiographic settings (Barron et al., 2013). Previous studies rely heavily on simulation models (Kang et al., 1998; Kim et al., 2005, 2014; Sakaguchi et al., 2014). Controversies on the magnitude and underlying mechanisms of hydrologic responses to land conversion during urbanization remain in the literature (Wang et al., 2009; He et al., 2009; Sun and Lockaby, 2012). In particular, data are scarce on the effects of converting paddy fields to other land uses, resulting in conflicting conclusions. For example, a simulation study in Taiwan suggested that rice paddy fields generated 55 % lower total runoff and 33 % lower peakflows than dry farms (Wu et al., 1997). However, another simulation study that used the HEC-HMS model for a rice paddy dominated watershed in southern China found that an increase in impervious surface areas from 3 to 30 % increased the peakflow rate and storm volume (4–20 %), but had very limited impacts on total annual flow (< 6 %) (Wang et al., 2009; Du et al., 2011, 2012) and thus long term water balances.

25 The highly populated Yangtze River Delta (YRD) region covers about 2 % of China's land mass, but provides over 18 % of China's Gross Domestic Product (Gu et al., 2011). The population increased by almost 13 % in the past decade to 156 million in 2013, and has become China's most industrialized region and one of the global "hot spots" of economic and social development. As the "homes of the fish and rice", southern China's landscapes have been dominated by rice paddy fields for thousands of years. The

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)

[⏪](#)

[⏩](#)

[◀](#)

[▶](#)

[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



original coastal wetlands have long been ditched, drained, and cultivated for growing rice and other crops. Rice paddy fields are major sources of food production and offer many other ecosystem services similar to wetlands including flood retention, ground-water recharge (He et al., 2009), nutrient cycling, and sequestration of greenhouse gases (Tsai, 2002). One study on 10 typical rice paddies in China concluded that their ecosystem service values exceed their economic values by three folds (Xiao et al., 2011). The rapid urbanization and population rise under a warming climate in the YRD region has caused serious environmental and resource concerns such as overdrawing and pollution of groundwater, flooding, land subsidence, and urban heat islands (He et al., 2007; Gu et al., 2011). The majority of the existing studies on paddy fields have focused on grain yield and irrigation with little research on the hydrologic response to urbanization in paddy field-dominated landscapes (Du et al., 2011, 2012; Kim et al., 2014).

Converting paddies to urban land use have many cascading effects on the local environments (Fig. 1). In particular, because rice paddy fields are rarely under water stress, the water loss or actual ET is close to the potential ET (Wu et al., 1997) and has been recognized as their cooling functions in regulating local climate (Xiao et al., 2011). In contrast, urban land use is generally characterized by low vegetation coverage with low ET and high runoff (Sun and Lockaby, 2012). Therefore, we hypothesized that converting rice paddy fields to urban areas represents the maximum ET reduction possible among all common land cover change scenarios, potentially resulting in disproportionately higher impacts on water balances than other land conversion scenarios (e.g., converting dryland to urban uses). Along with the increase in impervious surface areas that are well known to increase stormflow, ET reduction during urbanization is likely to cause large impacts on the local micro-climate, streamflow, and water quality on paddy field dominated watersheds (Fig. 1).

The overall goal of this study was understand the processes underlining the hydrologic impacts of converting rice paddy fields to urban uses. The specific objectives of this study were: (1) examine how urbanization in the past decade (2000–2013) has

increased from 3 to 8 % from 1988 to 2003, but jumped to 25 % in 2012, and the area of rice paddy fields decreased from 45 % of the total land area in 2001 to 37 % in 2012 (Du et al., 2011, 2012; Chen and Du, 2014) (Fig. 2). Jiangsu Province Rural Statistics also reported that rice planting area in the QRB shrank more than 25 % from 995 km² in 2000 to 745 km² in 2010.

2.2 Land use, climate, streamflow, potential ET, ET, and Leaf Area Index (LAI)

We compiled land use and land cover data acquired for the period from 1988 to 2012 from multiple sources, including published thesis and journal papers for the study basin (Du et al., 2012; Chen and Du, 2014), and publically release land use databases for China such as the 30m Landsat data for 2000 created by the Chinese Academy of Sciences (Liu et al., 2009), and for 2010 we used the new Finer Resolution Observation and Monitoring of Global Land Cover that was created by Tsinghua University using Landsat TM and ETM+ data (Gong et al., 2013). The daily meteorological data (Precipitation, Radiation, Temperature, Wind Speed, and Humidity) for estimating potential ET were acquired from four standard climatic stations maintained by the local meteorological bureau across the QRB (Fig. 2). Streamflow data with varying temporal resolutions were compiled from hydrologic records for two hydrological stations, the Wuding Sluice Gate (Wuding Station thereafter) and the Inner Qinhuai Sluice Gate (Inner Qinhuai Station thereafter), which controlled the outflows from the Qinhuai River and backflows from the Yangtze River (Fig. 2). The daily streamflow data (2002–2003; 2006–2013) for the “flooding periods” from May to October recorded at the Wuding Station were used to characterize highflows and lowflows. The total annual streamflow discharged to the Yangtze River was the sum of flows measured at the Wuding Station and the Inner Qinhuai Station. Total annual streamflow data for the period of 1986–2006 were reported in Du et al. (2011) and we collected daily and monthly streamflow data for other periods.

Potential ET (PET) represents the maximum ecosystem evapotranspiration when soil water is not limited, such as the case of paddy fields. PET represents a compre-

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



hensive index of availability of atmospheric evaporative energy that is controlled by radiation, temperature, humidity, and wind speed. Daily PET rates were calculated using the standard FAO 56 method and were averaged across the four climatic stations for the period 2000–2013 (Allen et al., 1994). The improved MOD16 datasets provide consistent estimates of global actual ET at an eight-day and 1-km² resolution (Mu et al., 2011). Yuan et al. (2011) reprocessed the MODIS leaf areas index (LAI) datasets using the modified temporal spatial filter (mTSF) and time-series analysis with the TIMESAT software (Jonsson and Eklundh, 2004) and provided reliable continuous LAI estimates from 2000 to 2013. Monthly PET and MODIS ET rates were presented in Fig. 4 along with other climatic variables to contrast seasonal ET, PET, and P that are control seasonal streamflow.

2.3 Change detection

Three statistical methods were used to analyze the temporal changes in the hydroclimatic data series: (1) The Mann–Kendall test (Mann, 1945; Kendall, 1975) for the non-linear trend at significance levels of $\alpha = 0.001, 0.01, 0.05,$ and $0.10,$ (2) the non-parametric Sen’s method for estimating the slope of an existing trend (Gilbert, 1987), and (3) Dynamic Harmonic Regression (DHR) method for determining the change rates for meteorological, hydrological, and LAI time series based on the Captain Toolbox (Taylor et al., 2007). The Flow Duration Curve (FDC) (Vogel and Fennessey, 1993) and the Double Mass Curve (DMC) methods (Wei and Zhang, 2010) were used to determine changes of streamflow frequency, distribution, and timing of changes in annual streamflow as a result of urbanization. The Baseflow Index (BFI) program was used to separate the baseflow and stormflow from measured daily streamflow (Wahl and Wahl, 1995). These hydrologic detection methods have been widely used for understanding the effects of land use change and climate change on streamflow (Ma et al., 2010; Tang et al., 2011; Wei and Zhang, 2012).

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



3 Results

3.1 Land conversion and change in LAI

During 1988–2012, the QRB has gone through dramatic land cover changes characterized by an increase in impervious areas and a decrease in paddy fields (Du et al., 2011, 2012; Chen and Du, 2014) (see insert in Fig. 2). The land cover change matrix showed that, from 2000 to 2010, the area of urban lands increased 482 km² or 307 % at the expense of cropland (–408 km², or –19 %) and forest lands (–133 km², or –50 %) (Table 2). Since dryland changed relatively small from 1988 to 2010 (insert Fig. 2), majority of detected reduction in cropland area came from the changes in paddy fields.

MODIS data indicated that both mean annual and peak growing season watershed level LAI decreased significantly ($p < 0.05$) with Z statistic = –2.08 and Z statistic = 2.41, respectively (Table 1) (Fig. 5). Since the major decrease in land use was paddy rice (Du et al., 2011), we believe that the decline of LAI was mainly caused by land conversion of paddy field to urban uses. The decrease trend of LAI followed a similar pattern as ET during 2000–2013.

3.2 Trend in Climate and MODIS ET

The M-K test showed that the growing season precipitation had a weak increasing trend, but annual total precipitation had an insignificant decreasing trend during 2000–2013 (Table 1). The mean annual air temperature showed an insignificant change, but with an weak increase of 0.07 °C yr^{–1} in the peak growing season from July to August (Table 1). In contrast, the mean temperature had a significant increase during the long period of 1961–2013 (Z statistic = 2.21, $p < 0.05$) (Fig. 3). Both annual and growing season PET rose significantly by 7.52 mm yr^{–1} (Z statistic = 2.52, $p < 0.05$) and 5.06 mm yr^{–1} (Z statistic = 2.41, $p < 0.05$), respectively, an opposite trend of the actual ET (Table 1). The DHR method also identified a rising trend for annual PET.

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



3.4 Changes in annual watershed water balances

The DMC analysis identified a clear “break point” of total annual streamflow (Q) around 2003 (Fig. 9). The slopes of the regression lines between accumulated precipitation and streamflow increased from 0.26 to 0.50. Mean annual streamflow significantly increased from 353 to 556 mm from period 1 (1986–2002) to period 2 (2003–2013) (Fig. 10). This represented an increase of runoff coefficient (Q/P) from 0.32 to 0.49, a 53 % increase. The trend of annual streamflow was influenced heavily by year 1991, a huge flooding event occurred in the Yangtze River Basin. When this year was removed, R^2 increased from 0.1 to 0.34 and p value increased to a highly significant level ($p = 0.002$). In the meantime, annual ET as estimated by $P-Q$, decreased significantly from 752 to 578 mm from period 1 to period 2, representing a decline in ET by 23 % or ET/P ratio by 25 %.

4 Discussion

4.1 Increased streamflow explained by the decreases in ET and LAI

The total streamflow (Fig. 10), and highflows and lowflows (Figs. 7 and 8) in the QRB have substantially increased during 2000–2013 while both LAI and ET have decreased (Figs. 5 and 10). Based on the watershed balance theory and all evidence of FDC analysis, we attributed the dramatic increase in streamflow to the decrease in LAI, not climate (PET or P), for the following three complementary reasons.

First, LAI is a major controlling factor for ET, especially during the growing season (Sun et al., 2011a, b; Sun and Lockaby, 2012) and in humid, energy-limited southern China in particular (Liu et al., 2013). The strong relationship between MODIS ET and LAI (Figs. 5) supported our hypothesis that urbanization dramatically reduced ET due to the reduction of LAI, thus explained the observed increase in streamflow.

HESSD

12, 1941–1972, 2015

**Urbanization
dramatically altered
the water balances**

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Second, regional annual ET is generally controlled by PET, P, and land surface conditions (Sun et al., 2005). A decrease in ET is normally caused by an increase in P and PET. Our data suggested that the decrease in ET was not caused by PET or P because annual and growing season PET significantly increased and precipitation did not change significantly. In fact, a negative correlation was found between ET and PET departures (Fig. 5). The DMC method that eliminated precipitation effects on streamflow also suggested the QRB had a shift of annual streamflow upward around 2003.

Third, the large decrease in LAI as detected by remote sensing corresponded closely to the dramatic conversion of rice paddy fields and increase in total impervious area (TIA) during the urbanization campaign in the QRB since the early 2000s. Previous studies in the United States suggest that stream flow and water quality regimes are degraded when the TIA exceeds 10–20% of total watershed area (Arnold and Gibbons, 1996; Bledsoe and Watson, 2001). Our study result was consistent with the finding of the threshold response in the literature, perhaps in the lower end of the spectrum (< 10%). The detected decrease in LAI due to shrinking rice paddy areas has overwhelmed the impacts of climate change (i.e. rise of PET) on ET, highlighting the importance of land use change in evaluating environmental change in the study region.

4.2 Regional hydrologic and environmental implications

Our findings complement an earlier study for the same basin that concluded that the elevated highflows were due to an increase in impervious surface area (Du et al., 2011, 2012). Our new analysis suggested that in addition to the increase in impervious surface areas other factors such as reduced ET could be the main causes that contributed to the observed increase in total flow and lowflows in the study basin. The present study advanced the understanding of the processes of hydrologic disturbances and has important environmental implications for paddy field-dominated regions in China and elsewhere in East Asia.

First, we confirmed our hypothesis that converting water stress-free paddy fields to relatively “dry” urban uses or impervious surfaces dramatically reduced ET (Fig. 1).

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)



[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



Thus, converting wetlands, such as paddy fields, to impervious or built-up areas is expected to have a much higher magnitude of hydrologic impacts than that for converting dry croplands or forests to urban land uses. The ET estimates based on two independent methods, watershed water balance and remote sensing, all show large decreases in ET.

Second, the populated study region is prone to floods and droughts due to the nature of a strong summer monsoon climate (Gu et al., 2011). Urbanization is likely to exacerbate the flood risks during the monsoon season as a result of decreased ET, an increased impervious surface area, and decreased retention capacity of rice paddy fields (Kang et al., 1998; Kim et al., 2014). In addition, an increase in stormflow has important concerns on stream channel stability, soil erosion, and reactivation of streambed sediment and pollutants (Sun and Locakby, 2012). This is of particular concern given the increasing trend of typhoon activities in southern China under climate change (Gu et al., 2011).

Third, the increasing trend in baseflow found in this study was in contradiction to the popular literature that suggested otherwise (Ott and Uhlenbrook, 2004; Kim et al., 2005; Price et al., 2011). The large reduction in ET from paddy fields might have overwhelmed the reduction of groundwater recharge from the increased impervious surfaces. The QRB is still dominated by croplands (65% of land area in 2010) and the dramatic reduction in water loss from rice cultivation and irrigation needs likely elevated groundwater recharge from uplands or stream channels overall. Reductions in forest land coverage, thus reduction in ET, have showed to increase baseflow in the humid piedmont region in North Carolina (Boggs and Sun, 2011) and northeastern US Lull and Sopper (1969). Boggs and Sun (2011) argue that the effects of vegetation removal on streamflow are most pronounced during the growing seasons when the contrast of ET between a vegetated surface and an urbanized surface is the greatest. Therefore, it is plausible that replacing paddy fields with high ET with urban land uses (grass or impermeable surface) with low ET may result in similar effect as forest removal during

for assessing the ecosystem service function (e.g., climate cooling, flood retention) of rice paddy fields must include proper algorithms describing the hydrological processes including ET that links water and energy balances.

Rice cultivations have been practices for thousands of years around the world. However, converting rice paddy fields to other uses in southern China and East Asia has been on the rise under a changing climate and demographics. Our study indicates that urbanization will likely increase the risk of flooding, heat islands, and social vulnerability due to the loss of ecosystem services of rice paddies. To minimize and mitigate the hydrologic and environmental impacts of converting paddy fields down streams while maintaining resource sustainability requires an integrated watershed management approach that involves careful urban planning (Dunne and Leopold, 1978), landscape design (Dietz and Clausen, 2008), and irrigation management (Park et al., 2009).

Acknowledgements. This study was supported by the Natural Science Foundation of China (71373130), the Jiangsu Key Laboratory of Agricultural Meteorology Fund (No. KYQ1201), the National Key Basic Research Program of China (2013CB430200, 2013CB430206), and IceMe of NUIST. Partial support was from the Southern Research Station, USDA Forest Service.

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Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)

[⏪](#)

[⏩](#)

[◀](#)

[▶](#)

[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



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Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Yao, W. B., Zhang, H., Zhu, P., Zhao, Z. Y., Zhang, H. Y., Zheng, Y. M., Ji, L. Y., Zhang, Y. W., Chen, H., Yan, A., Guo, J. H., Yu, L., Wang, L., Liu, X. J., Shi, T. T., Zhu, M. H., Chen, Y. L., Yang, G. W., Tang, P., Xu, B., Ciri, C., Clinton, N., Zhu, Z. L., Chen, J., and Chen, J.: Finer resolution observation and monitoring of global land cover: first mapping results with Landsat TM and ETM+ data, *Int. J. Remote Sens.*, 34, 2607–2654, 2013.

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Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[⏪](#)[⏩](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

Table 2. The conversion matrix for area changes in land use and land cover during the period of 2000 to 2010 in the Qinhuai River Basin. Cropland refers to paddy field and dry crop lands.

2000 (km ²)	2010 (km ²)				
	Cropland	Forest	Water	Urban	Σ (2000)
Cropland	1516	37	86	478	2116
Forest	128	98	8	37	271
Water	28	3	33	10	73
Urban	37	1	5	115	157
Σ (2010)	1709	137	132	639	2617
Area change from 2000 to 2010 (km ²)	-408	-133	59	482	-
Area change from 2000 to 2010 (%)	-19	-50	81	307	-

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)

[⏪](#)

[⏩](#)

[◀](#)

[▶](#)

[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



Urbanization dramatically altered the water balances

L. Hao et al.

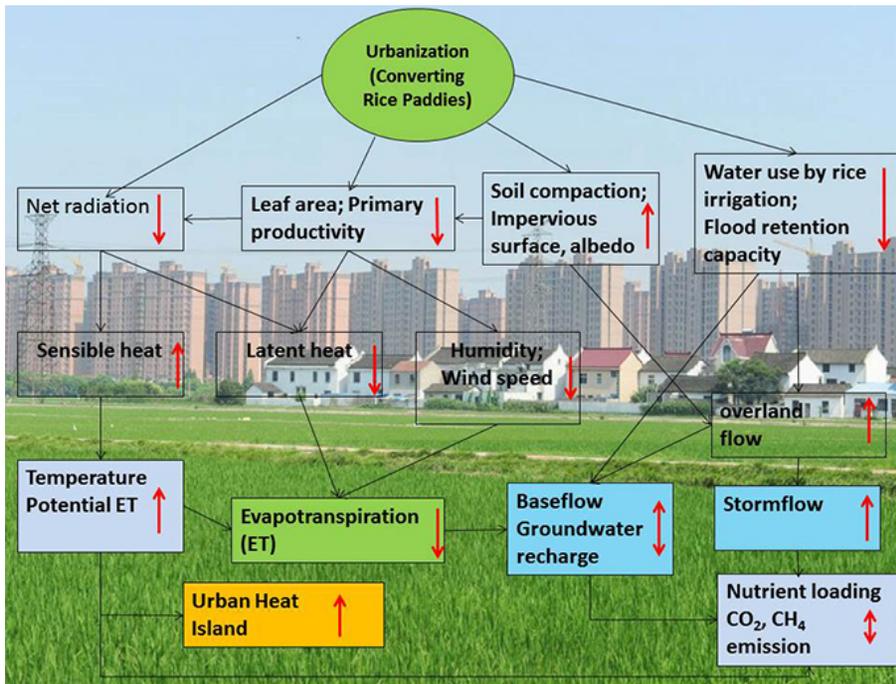


Figure 1. A conceptual model illustrating the potential hydrologic and environmental impacts of converting rice paddies to urban uses in the Yangtze River Delta region. Arrows represent directions (up or down or both) of change (Background photo credit: <http://blog.sciencenet.cn/blog-578415-712508.html>).

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



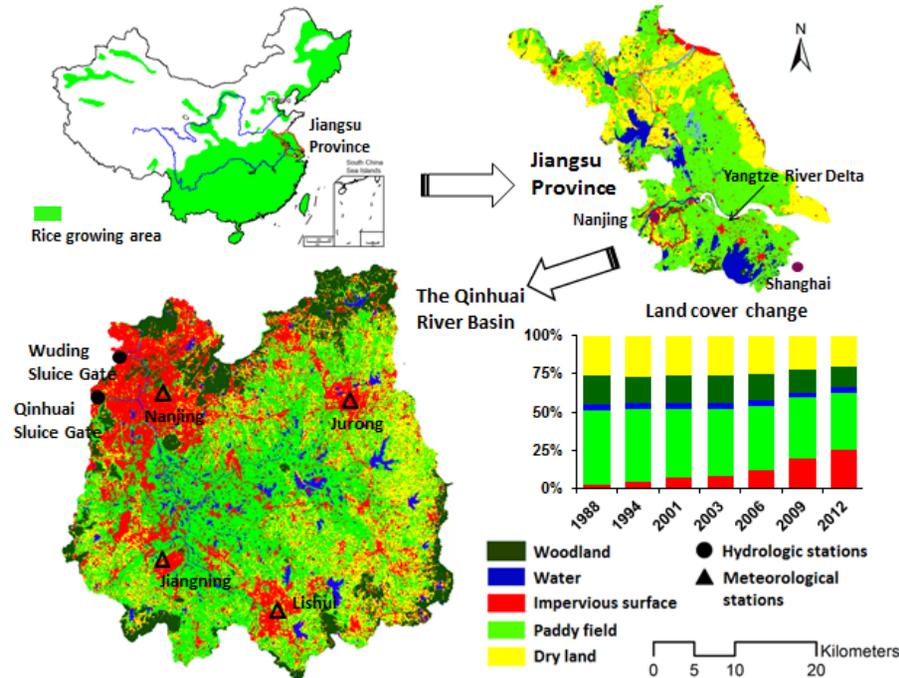


Figure 2. Watershed locations, instrumentation, and land use change from 1988 and 2012 (Du et al., 2012; Chen and Du, 2014) in the Qinhuai River Basin, Yangtze River Delta, southern China. The insert map shows land cover in the year 2012 generated from Landsat ETM+ images.

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#)

[Introduction](#)

[Conclusions](#)

[References](#)

[Tables](#)

[Figures](#)

[⏪](#)

[⏩](#)

[⏴](#)

[⏵](#)

[Back](#)

[Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



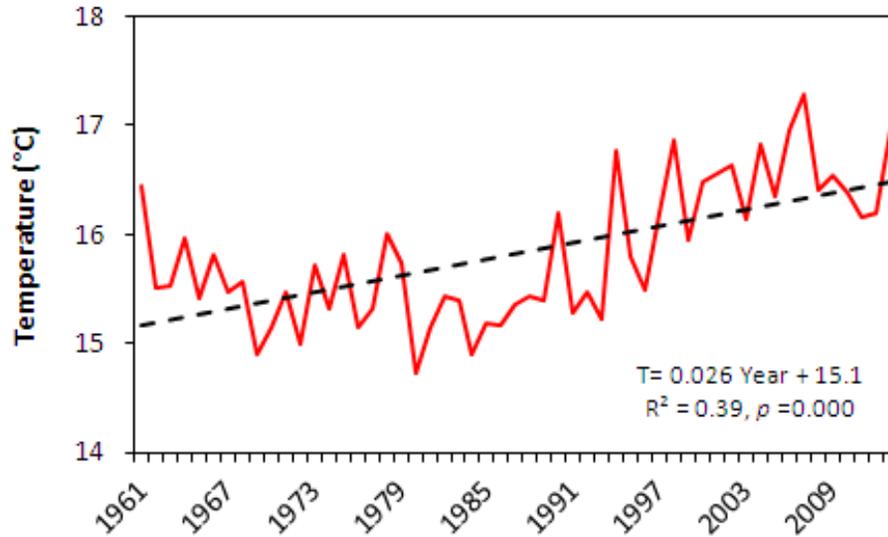


Figure 3. Mean annual air temperature change across four meteorological stations in the Qin-huai River Basin, southern China, during 1961–2013.

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#) | [Introduction](#)

[Conclusions](#) | [References](#)

[Tables](#) | [Figures](#)

[◀](#) | [▶](#)

[◀](#) | [▶](#)

[Back](#) | [Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



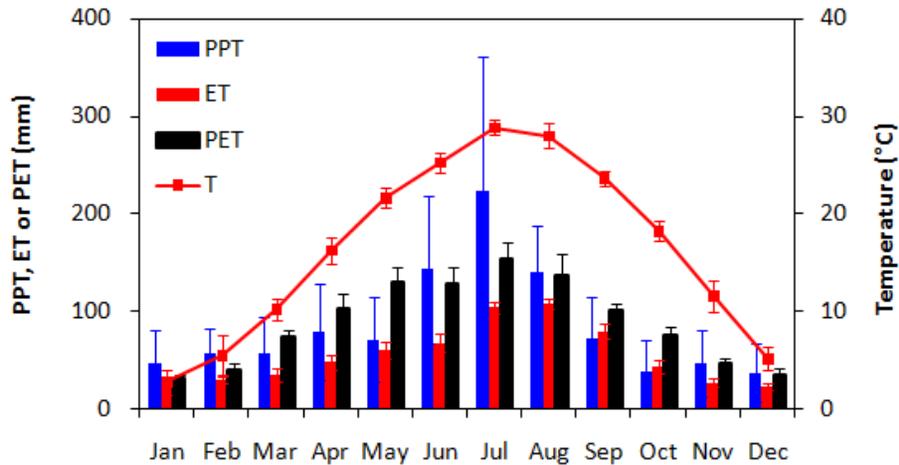


Figure 4. Mean monthly precipitation (P), MODIS ET, PET and temperature (T) (2000–2013), and the vertical lines are SD.

HESSD

12, 1941–1972, 2015

Urbanization dramatically altered the water balances

L. Hao et al.

[Title Page](#)

[Abstract](#) [Introduction](#)

[Conclusions](#) [References](#)

[Tables](#) [Figures](#)

[◀](#) [▶](#)

[◀](#) [▶](#)

[Back](#) [Close](#)

[Full Screen / Esc](#)

[Printer-friendly Version](#)

[Interactive Discussion](#)



Urbanization dramatically altered the water balances

L. Hao et al.

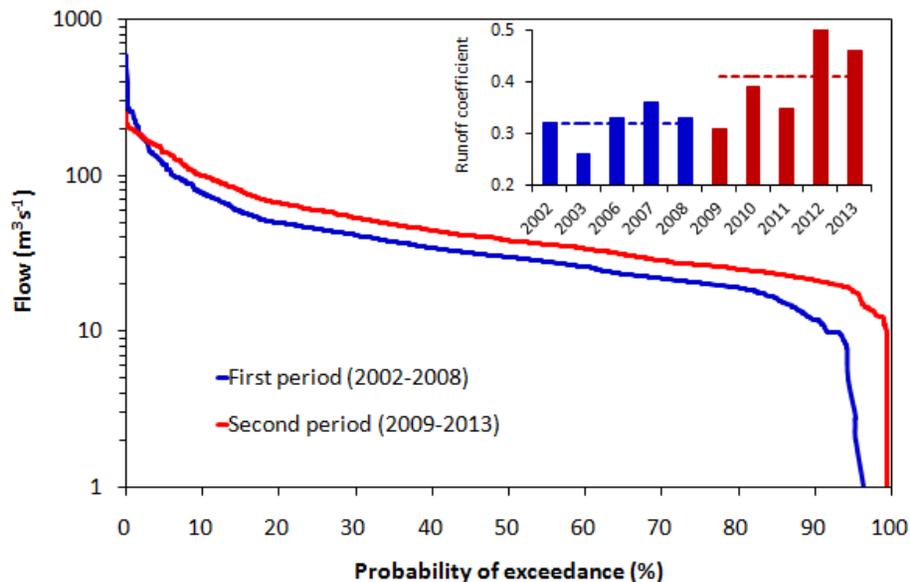


Figure 7. Flow Duration Curve for mean daily flow in the first period (2002–2003 and 2006–2008) and the second period (2009–2013) (May–October, Wuding Station) in the Qinhuai River Basin. Insert is runoff coefficient, the ratio of streamflow/precipitation for the period from May to October when the flow control gate was open.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

⏪

⏩

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion



**Urbanization
dramatically altered
the water balances**

L. Hao et al.

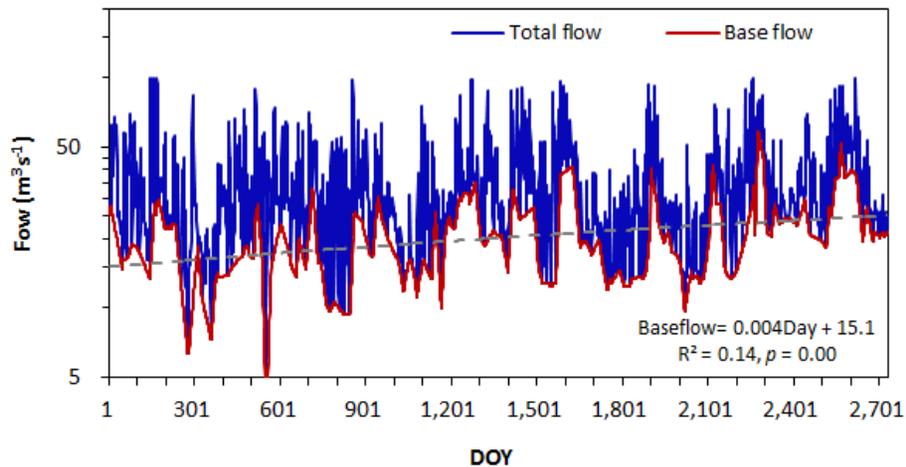


Figure 8. Trend of daily base flow separated from total stream flow measured at Wuding Station during 2006–2013. DOY is the day since 1 January 2006.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

Urbanization dramatically altered the water balances

L. Hao et al.

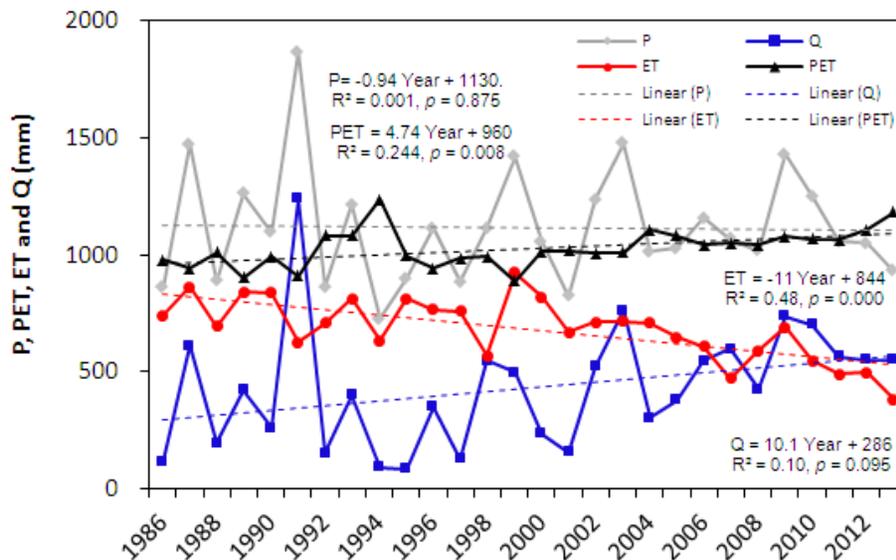


Figure 10. Trend of annual water balance and potential evapotranspiration (PET) for the Qinhuai River Basin from 1986–2013. ET was estimated as the difference between precipitation (P) and measured streamflow (Q).

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures



Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

