

# 1 **Towards a simple representation of chalk hydrology in land**

## 2 **surface modelling**

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### 6 **Abstract**

7 Modelling and monitoring of hydrological processes in the unsaturated zone of chalk, a  
8 porous medium with fractures, is important to optimize water resources assessment and  
9 management practices in the United Kingdom (UK). However, incorporating the processes  
10 governing water movement through chalk unsaturated zone in a numerical model is  
11 complicated mainly due to the fractured nature of chalk that creates high-velocity preferential  
12 flow paths in the subsurface. In general, flow through chalk unsaturated zone is simulated  
13 using dual-porosity concept, which often involves calibration of relatively large number of  
14 model parameters, potentially undermining applications to large regions. In this study, a  
15 simplified parameterization, namely the Bulk Conductivity (BC) model is proposed for  
16 simulating hydrology in chalk unsaturated zone. This new parameterization introduces only  
17 two additional parameters (namely the macroporosity factor and the soil wetness threshold  
18 parameter for fracture flow activation) and uses the saturated hydraulic conductivity from  
19 chalk matrix. The BC model is implemented in the Joint UK Land Environment Simulator  
20 (JULES) and applied to a study area encompassing the Kennet catchment in the Southern  
21 UK. This parameterization is further calibrated at a point-scale using soil moisture profile  
22 observations. The performance of calibrated BC model in JULES is assessed and compared  
23 against the performance of both the default JULES parameterization and the uncalibrated  
24 version of BC model implemented in JULES. Finally, the model performance at the

25 catchment-scale is evaluated against independent data sets (e.g., runoff and latent heat flux).  
26 The results demonstrate that the inclusion of the BC model in JULES improves simulated  
27 land surface mass and energy fluxes over the chalk-dominated Kennet catchment. Therefore,  
28 the simple approach described in this study may be used to incorporate the flow processes  
29 through chalk unsaturated zone in large-scale land surface modelling applications.

30 **Keywords:** Chalk hydrology, macroporosity, land surface model, bulk conductivity model.

## 31 **1. Introduction**

32 Chalk can be described as a fine-grained porous medium traversed by fractures [*Price et al.*,  
33 1993]. Previous studies showed that the unsaturated zone of the chalk aquifers plays an  
34 important role on groundwater recharge in the UK [e.g., *Lee et al.*, 2006; *Ireson et al.*, 2009].  
35 Therefore, both monitoring [e.g., *Bloomfield*, 1997; *Ireson et al.*, 2006] and modelling [e.g.,  
36 *Bakopoulou*, 2015; *Brouyère*, 2006; *Ireson and Butler*, 2011, 2013; *Sorensen et al.*, 2014]  
37 strategies have been adapted previously to understand the governing hydrological processes  
38 in the chalk unsaturated zone.

39 In chalk, the matrix provides porosity and storage capacity, while the fractures greatly  
40 enhance permeability [*Van den Daele et al.*, 2007]. Water movement through chalk matrix is  
41 slow due to its relatively high porosity (0.3-0.4) and low permeability ( $10^{-9}$ - $10^{-8}$  ms<sup>-1</sup>). A  
42 fractured chalk system, in contrast, conducts water at a considerably higher velocity because  
43 of relatively high permeability ( $10^{-5}$ - $10^{-3}$  ms<sup>-1</sup>) and low porosity (of the order  $10^{-4}$ ) of  
44 fractures [*Price et al.*, 1993].

45 Simulating water flow through the matrix-fracture system of chalk has been the subject of  
46 research for some time. Both conceptual [e.g., *Price et al.*, 2000; *Haria et al.*, 2003] and  
47 physics-based [e.g., *Mathias et al.*, 2006; *Ireson et al.*, 2009] models have been proposed  
48 previously to describe water flow through chalk unsaturated zone. The physics-based models  
49 mentioned above were developed based on dual-continua approach and required relatively

50 large numbers of parameters (i.e., on the order of 20-30 parameters) that were calibrated via  
51 inverse modelling using observed soil moisture and matric potential data [e.g., *Ireson et al.*,  
52 2009; *Mathias et al.*, 2006].

53 In recent years, representation of chalk has gained attention in land surface modelling. For  
54 example, *Gascoign et al.* [2009] applied the Catchment Land Surface Model (CLSM) over the  
55 Somme River basin in northern France. A linear reservoir was included in the TOPMODEL  
56 based runoff formulation of CLSM to account for the contribution of chalk aquifers to river  
57 discharge. *Le Vine et al.* [2016] applied the Joint UK Land Environment Simulator (JULES  
58 [*Best et al.*, 2011]) over the Kennet catchment in southern England to evaluate the  
59 hydrological limitations of land surface models. In that study, two intersecting Brooks and  
60 Corey curves were proposed, which allowed a dual curve soil moisture retention  
61 representation for the two distinct flow domains of chalk (i.e., matrix and fracture) in the  
62 model. Considering this dual Brooks and Corey curve, a three-dimensional groundwater flow  
63 model (ZOOMQ3D [*Jackson and Spink*, 2004]) was coupled to JULES to demonstrate the  
64 strong influence of representing chalk hydrology and groundwater dynamics on simulated  
65 soil moisture and runoff.

66 The above mentioned studies illustrate the importance of representing chalk in land surface  
67 modelling. However, including chalk hydrology in large-scale land surface modelling using  
68 the contemporary dual-porosity concept can be complicated due to large number of additional  
69 parameters. In this context, we propose a new parameterization, namely the Bulk  
70 Conductivity (BC) model as a first step towards a simple chalk representation suitable for  
71 land surface modelling. In order to test the proposed parameterization, the BC model is  
72 included in JULES (version 4.2), which, by default (i.e., uniform soil column representation  
73 using general soil database as typically applied in land surface models), does not represent  
74 any chalk feature. In this study, the BC model (included in JULES) is applied at two distinct

75 spatial scales (i.e., point and catchment). At the point-scale, the proposed parameterization is  
 76 calibrated using observed soil moisture profile data. This is achieved by randomly sampling  
 77 the parameter space and extensively running the model in order to minimize the differences  
 78 between observed and simulated soil moisture variability at different depths. Finally, the  
 79 proposed model is applied to the Kennet catchment in the Southern England and the fluxes  
 80 and states of the hydrological cycle are simulated for multiple years. The simulation results  
 81 are evaluated using observed latent heat flux (LE) and runoff data to assess the performance  
 82 of the BC model in simulating land surface processes at the catchment scale.

## 83 **2. A model of flow through chalk unsaturated zone**

84 In this study, the *Bulk Conductivity* (BC) model based on the work by *Zehe et al.* [2001] is  
 85 incorporated in JULES to represent the flow of water through the fractured chalk unsaturated  
 86 zone. According to this approach, if the relative saturation ( $S$ ) exceeds a certain threshold ( $S_0$ )  
 87 at a soil grid, the saturated hydraulic conductivity of chalk matrix ( $K_s$ ) is increased to a bulk  
 88 saturated hydraulic conductivity ( $K_{sb}$ ) as follows

$$89 \quad K_{sb} = K_s + K_s f_m \frac{S - S_0}{1 - S_0} \quad \text{if } S > S_0 \quad (1)$$

$$90 \quad K_{sb} = K_s \quad \text{if } S \leq S_0 \quad (2)$$

$$91 \quad \text{with } S = \frac{\theta - \theta_r}{\theta_s - \theta_r}$$

92 where  $f_m$  is a macroporosity factor (-),  $\theta$  is soil moisture ( $\text{m}^3\text{m}^{-3}$ ),  $\theta_s$  is soil moisture at  
 93 saturation ( $\text{m}^3\text{m}^{-3}$ ) and  $\theta_r$  is the residual soil moisture ( $\text{m}^3\text{m}^{-3}$ ). Note that  $S$  ranges from zero in  
 94 case of completely dry soils to one for fully wet soils.

95 At the first step of evaluation, the  $K_s$ ,  $S_0$  and  $f_m$  parameters are estimated based on existing  
 96 literature to assess the performance of the uncalibrated BC model. For the matrix saturated  
 97 hydraulic conductivity ( $K_s$ ), we use  $K_s = 1.0 \text{ mmd}^{-1}$  following *Mathias et al.* [2006]. In

98 addition, Equation 1 indicates that the onset of water flow through the fracture system of  
99 chalk is controlled by the threshold  $S_0$ . According to *Wellings and Bell* [1980], water flow  
100 through fractures dominates over matrix flow in chalk when the pressure head in soil  
101 becomes higher than  $-0.50 \text{ mH}_2\text{O}$ . We consider a value of  $S_0 = 0.80$  for the uncalibrated BC  
102 model, which is based on observed soil moisture-matric potential relationship in the study  
103 area.

104 Finally, In *Zehe et al.* [2001],  $f_m$  was defined as the ratio of the saturated water flow rate in all  
105 macropores in a model element to the corresponding value in soil matrix, which can be  
106 determined based on the density and length of fractures at small scales. In addition,  $f_m$  has  
107 also been considered as a calibration parameter previously [e.g., *Blume, 2008; Zehe et al.,*  
108 *2013*]. In this study, we define  $f_m$  as a characteristic soil property reflecting the influence of  
109 fractures on soil water movement [*Zehe and Blöschl, 2004*] and estimate it from the relative  
110 difference of permeability between chalk matrix and fractured chalk system that can be of the  
111 order  $10^4$ - $10^6$  according to *Price et al.* [1993]. Consequently, we consider a macroporosity  
112 factor of  $f_m = 10^5$  for the uncalibrated BC model. In the following step, the BC model is  
113 calibrated to minimize the differences between the variability of observed and simulated soil  
114 moisture at individual depths. The calibration strategy will be discussed elaborately in section  
115 3.5.

### 116 **3. Methods**

#### 117 **3.1. Study area**

118 The study area encompasses the Kennet catchment located in the Southern England with an  
119 area of about  $1033 \text{ km}^2$  (Figure 1a). Generally, Kennet is rural in nature with scattered  
120 settlements and has a maximum altitude of approximately 297 m (Above Ordnance Level).  
121 The River Kennet discharges into the North Sea through London. The major tributaries of

122 this river are Lambourn, Dun, Enborne, and Foudry Brook. An average annual rainfall of  
123 approximately 760 mm was recorded in the catchment over a 40 year period from 1961-1990.  
124 Solid geology of the Kennet catchment is dominated by chalk, which is overlain by thin soil  
125 layer. While lower chalk outcrops along the northern catchment boundary, progressively  
126 younger rocks are found in the southern part. In general, surface runoff production is very  
127 limited over the regions of the catchment where chalk outcrops. The flow regime shows a  
128 distinct characteristics of slow response to groundwater held within the chalk aquifer [*Le*  
129 *Vine et al.*, 2016]. According to *Ireson and Butler* [2013], the unsaturated zone of chalk  
130 shows slow drainage over summer and bypass flow during wet periods in this catchment.

### 131 **3.2. Field measurements and remotely sensed data**

132 Table 1 summarizes the field measurements and remote sensing data used in this study. We  
133 use *in-situ* soil moisture and runoff measurements along with remotely sensed LE data to  
134 assess model performance in simulating the mass and energy balance components of the  
135 hydrological cycle. Point scale soil moisture measurements at two adjacent sites (~20 m  
136 apart) at the Warren Farm (Figure 1) were provided by Centre for Ecology and Hydrology  
137 (CEH). A Didcot neutron probe was used at these locations to measure fortnightly soil  
138 moisture at different depths below land surface (10 cm apart down to 0.8 m, 20 cm apart  
139 between 0.8-2.2 m, and 30 cm apart between 2.2-4.0 m) [*Hewitt et al.*, 2010].

140 The National River Flow Archive (NRFA) coordinates discharge measurements from the  
141 gauging station networks across UK. These networks are operated by the Environmental  
142 Agency (England), Natural Resources Wales, Scottish Environment Protection Agency, and  
143 Rivers Agency (Northern Ireland). We use discharge measurements provided by NRFA to  
144 assess model performance in simulating runoff over the Kennet catchment in this study.

145 The MOD16 product of the Moderate Resolution Imaging Spectroradiometer (MODIS) is a  
146 part of NASA/EOS project that provides estimation of global terrestrial LE. The LE  
147 estimation from MOD16 is based on remotely sensed land surface data [e.g., *Mu et al.*, 2007].  
148 In this study, the 8-day and monthly LE data products from MODIS is used to evaluate the  
149 model performance in simulating land surface energy fluxes.

### 150 **3.3. Land surface model**

151 In this study, we use the Joint UK Land Environment Simulator (JULES [e.g., *Best et al.*,  
152 2011; *Clark et al.*, 2011]) version 4.2. JULES is a flexible modelling platform with a modular  
153 structure aligned to various physical processes developed based on the Met Office Surface  
154 Exchange Scheme (MOSES [e.g., *Cox et al.*, 1999; *Essery et al.*, 2003]). Meteorological data  
155 including precipitation, incoming short- and longwave radiation, temperature, specific  
156 humidity, surface pressure, and wind speed are required to drive JULES. Each grid box in  
157 JULES can comprise nine surface types (broadleaf trees, needle leaf trees, C3 grass, C4 grass,  
158 shrubs, inland water, bare soil, and ice) represented by respective fractional coverage. Each  
159 surface type is represented by a tile and a separate energy balance is calculated for each tile.

160 Subsurface heat and water transport equations are solved based on finite difference  
161 approximation in JULES as described in *Cox et al.* [1999]. Moisture transport in the  
162 subsurface is described by the finite difference form of Richards' equation. The vertical soil  
163 moisture flux is calculated using the Darcy's law. While the top boundary condition to solve  
164 the Richards' equation is infiltration at soil surface, the bottom boundary condition in JULES  
165 is free drainage that contributes to subsurface runoff.

166 Surface runoff is calculated by combining the equations of throughfall and grid box average  
167 infiltration in JULES. In order to direct the generated runoff to a channel network, river  
168 routing is implemented based on the discrete approximation of one-dimensional kinematic

169 wave equation [e.g., *Bell et al.*, 2007]. In this approach, river network is derived from the  
170 digital elevation model (DEM) of the study area and different wave speeds are applied to  
171 surface and subsurface runoff components and channel flows [e.g., *Bell and Moore*, 1998]. A  
172 return flow term accounts for the transfer of water between subsurface and land surface [e.g.,  
173 *Dadson et al.*, 2010, 2011].

#### 174 **3.4. Model configurations and input data**

175 In this study, simulations are performed at two distinct spatial scales, namely point and  
176 catchment. At the point scale, JULES is configured to simulate the mass and energy fluxes at  
177 the Warren Farm site (Figure 1a). A total subsurface depth of 5 m is considered in the model  
178 with a vertical discretization ranging from 10 cm at the land surface to 50 cm at the bottom of  
179 the model domain. Note that this discretization is consistent with the soil moisture  
180 measurement depths mentioned in section 3.2. The vegetation type is implemented as C3  
181 grass using the default parameters in JULES. Point scale simulations were performed over 2  
182 consecutive years from 2003-2005 at an hourly time step. Except for precipitation, hourly  
183 atmospheric forcing data to drive JULES was obtained from an automatic weather station  
184 operated by the CEH at Warren Farm. In order to estimate hourly precipitation data to run  
185 JULES, rain gauge measurements from the Met Office [*Met Office*, 2006] were used. Inverse  
186 distance interpolation technique [e.g. *Garcia et al.*, 2008; *Ly et al.*, 2013] was applied on  
187 rainfall measurements from 13 gauges closest to Warren Farm (distance varies from 25-60  
188 km) to obtain hourly precipitation for the point scale simulations.

189 At the catchment scale, JULES is configured over a study area encompassing the Kennet  
190 catchment (Figure 1a) considering a uniform lateral grid resolution of 1 km with 70 x 40 cells  
191 in  $x$  and  $y$  dimensions, respectively. The total subsurface depth and vertical discretization are  
192 identical to those of the point scale simulations. Spatially distributed vegetation type

193 information for the study area (Figure 1b) is obtained from the Land Cover Map 2007  
194 (LCM2007) dataset [Morton *et al.*, 2011]. Simulations were performed over 5 consecutive  
195 years from 2006-2011 at the catchment scale. Note that the simulation periods of catchment  
196 and point scale (2003-2005) do not coincide due to the availability of soil moisture  
197 measurements described in section 3.2. Spatially distributed meteorological data from the  
198 Climate, Hydrology and Ecology research Support System (CHESS) was used to obtain the  
199 atmospheric forcing to drive JULES at the catchment scale. The CHESS data includes 1 km  
200 resolution gridded daily meteorological variables [Robinson *et al.*, 2015]. This daily data is  
201 downscaled using a disaggregation technique described in Williams and Clark [2014] to  
202 obtain hourly atmospheric forcing. The flow direction required for river routing is extracted  
203 from the USGS HydroSHEDS digital elevation data [Lehner *et al.*, 2008].

204 We estimate the soil hydraulic properties based on texture (Table 2). At the point scale, loam  
205 soil is dominant at the Warren Farm site. At the catchment scale, the Harmonized World Soil  
206 Database (HWSD) from the Food and Agricultural Organization of UNO (FAO) is used to  
207 obtain the texture of different soil types over Kennet (Figure 1c). The saturation-pressure  
208 head relationship for different soil types is described using the Van Genuchten [Van  
209 Genuchten, 1980] model with parameter values (Table 2) obtained from Schaap and Leij  
210 [1998].

211 The hydraulic properties for chalk used in this study are summarized in Table 3. These  
212 properties are obtained based on existing literature as a first step when evaluating the  
213 uncalibrated BC model. The BC model parameters are subsequently calibrated to minimize  
214 the differences between observed and simulated  $\Delta\theta$  (section 3.5) at various soil depths.

215 In this study, we consider two different model configurations, namely *default* and *macro*  
216 (Figure 2). The *default* configuration corresponds to the standard parameterizations of JULES

217 that does not represent chalk hydrology in the model. In this configuration, each soil column  
 218 in JULES is considered to be vertically homogeneous with the soil properties defined in  
 219 Table 2, which is motivated by the Met Office JULES Global Land 4.0 configuration  
 220 described in *Walters et al.* [2014]. The *macro* configuration, in contrast, explicitly represents  
 221 chalk by applying the BC model starting at 30 cm below land surface to the bottom of the  
 222 model domain (i.e. 500 cm). Therefore, the soil column in the *macro* configuration can be  
 223 divided into topsoil (0-30 cm) and chalk (30-500 cm). The topsoil depth of 30 cm is defined  
 224 based on several augured soil samples collected during a field campaign at Warren Farm in  
 225 2015 (Figure 2). This depth is corroborated by additional information from the British  
 226 Geological Survey (BGS) operated borehole records  
 227 ([http://www.ukso.org/pmm/soil\\_depth\\_samples\\_points.html](http://www.ukso.org/pmm/soil_depth_samples_points.html)), which show that topsoil depths  
 228 vary from 10-40 cm over the study area. We apply the *macro* configuration assuming a  
 229 spatially homogeneous topsoil depth of 30 cm for both point and catchment scale simulations.  
 230 Note that except for this inclusion of chalk, *default* and *macro* configurations are identical in  
 231 terms of model set up and input data. It should also be emphasized that *default* represents a  
 232 “naïve” configuration deprived of model calibration. Moreover, this configuration does not  
 233 represent chalk, which, according to previous studies [e.g., *Le Vine et al.*, 2016], substantially  
 234 affects the hydrology of the study area considered here.

### 235 **3.5. Calibration of the BC model**

236 We calibrate the BC model at the point-scale to minimize the differences between observed  
 237 and simulated soil moisture variability ( $\Delta\theta$ ) at different depths. The Root Mean Squared  
 238 Error (RMSE) is used as the objective function to optimize the BC model parameters [e.g.,  
 239 *Ireson et al.*, 2009]

$$240 \quad RMSE = \frac{1}{nd} \sum_1^{nd} \sqrt{\left( \frac{1}{nt-1} \sum_2^{nt} (\Delta\theta_{d,t}^{obs} - \Delta\theta_{d,t}^{sim})^2 \right)} \quad (3)$$

241 where  $nd$  is the number of soil layers,  $nt$  is the number of soil moisture observations available  
242 for a layer  $d$ ,  $\Delta\theta^{obs}$  is the observed variability of soil moisture and  $\Delta\theta^{sim}$  is the simulated  
243 variability of soil moisture. Note that we consider  $\Delta\theta$  for this optimization because of its  
244 relevance to the water flux and recharge through chalk unsaturated zone [e.g., *Ireson and*  
245 *Butler*, 2011].

246 Equation (1) reveals that the calibration of the BC model involves optimizing 3 parameters,  
247 namely the saturated hydraulic conductivity of chalk matrix ( $K_s$ ), saturation threshold ( $S_0$ )  
248 and macroporosity factor ( $f_m$ ). *Ireson et al.* [2009] suggested a range of 0.2-2.0  $\text{mmd}^{-1}$  for  $K_s$ .  
249 On the other hand, *Price et al.* [1993] argued that in general,  $K_s$  is around 3-5  $\text{mmd}^{-1}$  for most  
250 chalk soils. Therefore, we consider a range of 0.2-5.0  $\text{mmd}^{-1}$  in optimizing  $K_s$ . We consider  $S_0$   
251 range 0-1, representing the entire physical domain for soil wetness from fully dry to fully  
252 wet, respectively. For  $f_m$ , a range of  $10^4$ - $10^6$  is considered, which, as discussed earlier, is  
253 consistent with the relative difference between the permeability of fractured chalk and chalk  
254 matrix according to *Price et al* [1993]. Latin hypercube sampling technique [e.g., *McKay et*  
255 *al.*, 2016] is used to generate 2,000 random samples for each BC model parameter within the  
256 ranges discussed above. Note that for the  $K_s$  parameter, the random sampling was performed  
257 from a logarithmic distribution [*Ireson et al.*, 2009]. We perform simulations using these  
258 random samples and calculate model performance (Equation 3) to select the optimum  
259 parameter values for the BC model for each possible parameter combination as discussed in  
260 details in the following section.

## 261 **4. Results and discussion**

### 262 **4.1. Point scale simulations**

263 At the point scale, the simulation results are evaluated using soil moisture observations at the  
264 Warren Farm site. Figure 3a compares observed and simulated soil moisture ( $\theta$ ) from the

265 *default* and *macro* configurations at 2 m below land surface. Note that the *macro*  
266 configuration uses the chalk hydraulic parameters collected from existing literature (Table 3).  
267 This figure shows that the *default* configuration considerably underestimates  $\theta$  throughout the  
268 simulation period, which is improved remarkably in case of *macro*. Figure 3b plots observed  
269 and simulated soil moisture variability ( $\Delta\theta$ ) from the *default* and *macro* configurations  
270 ( $\Delta\theta_{default}$  and  $\Delta\theta_{macro}$ , respectively) at the Warren Farm site. In general, both configurations  
271 show discrepancies with observed  $\Delta\theta$  with *macro* showing relatively better model  
272 performance.

273 The results show that despite the *macro* configuration improves simulated  $\theta$ , it shows  
274 considerable discrepancies with observed  $\Delta\theta$ , which is consistent throughout the whole chalk  
275 profile (results from other model layers are not shown). In order to minimize the differences  
276 between observed and modelled  $\Delta\theta$  from the *macro* configuration, we calibrate the BC model  
277 following the methodology described in section 3.5. The optimization results are summarized  
278 in Figure 4. Note that for each combination considered in the optimization, 2,000 model runs  
279 were performed using randomly sampled parameters as discussed in section 3.5. In addition  
280 to the *default* and *macro* cases, the calibrated cases in Figure 4 correspond to the results from  
281 the model runs yielding the lowest RMSE for each parameter combination evaluated.

282 The RMSE between observed and simulated  $\Delta\theta$  for the model configurations considered in  
283 the optimization is shown in Figure 4a. This figure illustrates that the RMSE of the *default*  
284 configuration is larger than that of *macro*, indicating better model performance in  
285 reproducing  $\Delta\theta$  for the latter (corresponding to a reduction of 15% in RMSE compared to the  
286 *default* case). Therefore, the uncalibrated BC model (i.e., *macro* configuration) better  
287 reproduces the soil moisture variability compared to *default*. Concerning the calibration of  
288 single BC model parameters, Figure 4a shows that  $S_0$  results in a 46% reduction of RMSE

289 compared to the *macro* configuration. Calibrating  $K_s$  or  $f_m$  individually yields only about 25%  
290 reduction of RMSE compared to *macro*.

291 Optimizing both  $K_s$  and  $S_0$  simultaneously shows the largest reduction (50%) of RMSE  
292 compared to *macro* which coincides with the total RMSE reduction found when all  
293 parameters are calibrated. Arguably, the BC model can be implemented in other chalk  
294 regions by constraining only  $S_0$  parameter. Such result could potentially be advantageous for  
295 transferability to other regions in the UK in order to assess chalk hydrology at large-scale.  
296 However since this is the first time the BC model is introduced, we decide to take a  
297 conservative approach and select the *macro* configuration with optimized  $K_s$  and  $S_0$  (*macro<sub>opt</sub>*  
298 hereafter) to simulate chalk hydrology over the study area that ensures best overall model  
299 performance.

300 The lower three panels in Figure 4 presents the BC model parameter values for the *default*  
301 and uncalibrated *macro* cases as well as for different combinations of parameters calibrated.  
302 The red bars in Figure 4a, b and c highlight the cases in which a given parameter is  
303 constrained by optimization. In those cases, the calibrated parameter values are obtained from  
304 model runs producing the lowest RMSE. An interesting feature in Figure 4b (calibrating  $K_s$   
305 individually) is that the optimization suggests a compensation mechanism in which  $K_s$  is  
306 increased remarkably in order to physically represent the “effective” flow through the chalk  
307 fractures in the BC model. This is not surprising and arguably the simplest way to attempt to  
308 improve model performance. For *macro<sub>opt</sub>*, the values used for  $K_s$  is relatively lower than that  
309 of uncalibrated *macro* case nevertheless consistent with previous estimates [e.g., Ireson *et al.*,  
310 2009]. Figure 4c clearly shows the dominance of  $S_0$  in the BC model as all the relatively low  
311 RMSE bars in Figure 4a are associated with  $S_0$  calibration (see red bars in Figure 4c). In  
312 addition, calibrated  $S_0$  values for all cases show a consistent constraint around 0.50. Finally,  
313 Figure 4d indicates the lack of influence for  $f_m$  parameter on model performance.

314 Figure 5 compares  $\Delta\theta_{default}$ ,  $\Delta\theta_{macro}$  and  $\Delta\theta$  from the *macro<sub>opt</sub>* configuration ( $\Delta\theta_{opt}$ ) with  
315 observed soil moisture variability ( $\Delta\theta_{obs}$ ). As mentioned earlier,  $\Delta\theta_{default}$  and  $\Delta\theta_{macro}$  show  
316 considerable discrepancies with  $\Delta\theta_{obs}$  while the *macro* configuration exhibits relatively better  
317 performance (Figure 3). Figure 5 illustrates that the overall agreement between observed and  
318 simulated  $\Delta\theta$  improves substantially in case of *macro<sub>opt</sub>* compared to *default* and *macro*,  
319 which is pronounced especially in the deeper chalk layers. Therefore, this figure indicates  
320 that the performance of the BC model in simulating  $\Delta\theta$  is further improved by optimizing the  
321  $K_s$  and  $S_0$  parameters simultaneously at the Warren Farm site.

322 As mentioned earlier, efficiently reproducing soil moisture variability over the profile is  
323 important due to the fact that  $\Delta\theta$  significantly affects water flux and recharge through chalk  
324 unsaturated zone. The drainage flux through the bottom of soil column ( $d_b$ ) of a land surface  
325 model can be considered as the potential recharge flux to groundwater [e.g., *Sorensen et al.*,  
326 2014]. Figure 6 compares the daily sum of  $d_b$  from the *default* and *macro<sub>opt</sub>* configurations at  
327 the Warren Farm site. Daily rainfall at this site over the simulation period is shown in Figure  
328 6a. In Figure 6b, the *macro<sub>opt</sub>* configuration shows considerable  $d_b$  during the colder months,  
329 while relatively slow drainage is observed in summer. In contrast, the *default* configuration  
330 shows relatively high  $d_b$  in summer compared to the colder months. In general, the recharge  
331 rate through chalk unsaturated zone during the warmer periods of the year is lower than that  
332 in the winter months [*Wellings and Bell*, 1980; *Ireson et al.*, 2009]. Therefore, the *macro<sub>opt</sub>*  
333 configuration appears to be more consistent with the recharge mechanism in chalk compared  
334 to *default*.

335 In this section, the BC model was evaluated at the point scale. The results showed that in  
336 general, the *macro* configuration outperforms the *default* case in simulating  $\Delta\theta$ . In order to  
337 improve the model performance even further, model parameter calibration was performed to

338 minimize the differences between observed and simulated  $\Delta\theta$  at the point scale. In the next  
339 sections, the optimized model (*macro<sub>opt</sub>*) is evaluated at the catchment scale.

#### 340 **4.2. Catchment scale simulations**

341 At the catchment scale, simulation results from the *default* and *macro<sub>opt</sub>* configurations are  
342 compared with the observations over the Kennet catchment. In order to assess the differences  
343 between LE from the *default* and *macro<sub>opt</sub>* configurations at the catchment scale, Figure 7  
344 plots spatially averaged 8-day composites of LE from MODIS ( $LE_{MOD}$ ) against the LE from  
345 these configurations ( $LE_{default}$  and  $LE_{opt}$ , respectively) over Kennet. The agreement between  
346 simulated LE and  $LE_{MOD}$  is evaluated using the coefficient of determination ( $R^2$ , see  
347 Appendix) and mean bias. Comparison between  $LE_{default}$  and  $LE_{MOD}$  shows a coefficient of  
348 determination of  $R^2_{default} = 0.78$  and a mean bias of  $bias_{default} = 10.5 \text{ Wm}^{-2}$ . The agreement  
349 between simulated LE and  $LE_{MOD}$  improves in case of the *macro<sub>opt</sub>* configuration, which is  
350 reflected by an increased coefficient of determination of  $R^2_{opt} = 0.80$  and a reduced mean bias  
351 of  $bias_{opt} = 7.1 \text{ Wm}^{-2}$ .

352 Figure 7 shows considerable differences between  $LE_{default}$  and  $LE_{opt}$  for relatively high LE,  
353 indicating discrepancies especially during the warmer months of the year. Spatially averaged  
354 time series of monthly  $LE_{MOD}$ ,  $LE_{default}$  and  $LE_{opt}$  is presented in Figure 8a. This figure shows  
355 that the differences between  $LE_{default}$  and  $LE_{opt}$  increases in summer compared to the colder  
356 months of the year, which is consistent with Figure 7. Consequently, the *default* configuration  
357 underestimates LE in summer compared to  $LE_{MOD}$ , which is improved in case of the *macro<sub>opt</sub>*  
358 configuration. In contrast, the differences between  $LE_{default}$  and  $LE_{opt}$  are negligible during the  
359 colder months of the year.

360 In addition, Figure 8b compares the observed and simulated monthly average discharge from  
361 the two model configurations at the “Kennet at Theale” gauging station (Figure 1a). This

362 figure depicts that the *default* configuration generally overestimates discharge at this gauging  
363 station, which is improved considerably in the case of *macro<sub>opt</sub>*. We use the Kling-Gupta  
364 Efficiency criterion (KGE [Gupta *et al.*, 2009]) to compare the performance of the two model  
365 configurations in reproducing observed discharge variability. As mentioned above, the  
366 *default* configuration overestimates discharge with  $KGE_{\text{default}} = -0.17$ . On the other hand, the  
367 *macro<sub>opt</sub>* configuration improves the agreement between observed and simulated discharge,  
368 which is reflected by  $KGE_{\text{opt}} = 0.51$ .

369 In order to summarize the results at catchment scale, Table 4 compares observed and  
370 simulated runoff from the two model configurations over the Kennet catchment from 2006-  
371 2011. The runoff ratio (*RR*, see Appendix), which is equal to the mean volume of flow  
372 divided by the volume of precipitation [e.g., Kelleher *et al.*, 2015], assesses the partitioning  
373 of precipitation into runoff over the catchment. The *default* configuration ( $RR = 0.82$ ) shows  
374 considerably higher *RR* compared to observation ( $RR = 0.40$ ), indicating overestimation of  
375 runoff by the model that is consistent with Figure 8b. Including chalk hydrology in the model  
376 remarkably improves the agreement between observed and simulated mean runoff over the  
377 Kennet catchment, which is assessed from a runoff ratio of  $RR = 0.46$  for the *macro<sub>opt</sub>*  
378 configuration which is much closer to the observed *RR* value than *default*.

379 In Table 4, the relative bias ( $\Delta\mu$ ) of 1.04 between observed and simulated runoff from the  
380 *default* configuration again indicates the overestimation by the model. In comparison,  
381 *macro<sub>opt</sub>* shows a smaller relative bias of  $\Delta\mu = 0.12$ , indicating improved agreement between  
382 observed and simulated mean runoff volume compared to *default*. The relative difference in  
383 standard deviation ( $\Delta\sigma$ , see Appendix) compares the variability of observed and simulated  
384 flow in Table 4 relating directly to the seasonal change in runoff. This comparison shows that  
385 the *default* configuration overestimates the variability of runoff over the Kennet catchment

386 ( $\Delta\sigma = 2.04$ ), which is improved in case of *macro* ( $\Delta\sigma = 0.65$ ). This improvement in  
387 reproducing flow variability is also clearly observed in Figure 8b.

388 In this section, the BC model is evaluated using observed mass and energy fluxes over the  
389 Kennet catchment. The *default* configuration suggested relatively low summertime LE over  
390 the catchment. The agreement between observed and simulated LE was improved in case of  
391 the *macro<sub>opt</sub>* configuration compared to *default*. It was also observed that the overall runoff  
392 prediction was considerably improved by *macro<sub>opt</sub>* compared to *default*. Given its simplicity,  
393 our results indicate that the proposed parameterization is suitable for use in land surface  
394 modelling applications.

## 395 **5. Summary and Conclusions**

396 In this study, we proposed a simple parameterization, namely the *Bulk Conductivity* (BC)  
397 model to simulate water flow through the matrix-fracture system of chalk in large scale land  
398 surface modelling applications. This parameterization was implemented in the Joint UK Land  
399 Environment Simulator (JULES) and applied to the Kennet catchment located in the southern  
400 UK to simulate the mass and energy fluxes of the hydrological cycle for multiple years. Two  
401 model configurations, namely *default* and *macro* were considered with the latter using the BC  
402 model to simulate chalk hydrology.

403 The proposed BC model is a single continuum approach of modelling preferential flow [e.g.,  
404 *Beven and Germann, 2013*] that involves only 3 parameters, namely the saturated hydraulic  
405 conductivity of chalk matrix ( $K_s$ ), macroporosity factor ( $f_m$ ) and relative saturation threshold  
406 ( $S_0$ ). Initially, these parameters were estimated from existing literature to assess the  
407 performance of the uncalibrated BC model. Finally, the BC model parameters were optimized  
408 to minimize the differences between observed and simulated soil moisture variability. Our  
409 results indicated that  $S_0$  is by far the most influential parameter in the model when

410 representing water movement through a soil-chalk column. This highlights the simplicity of  
411 the proposed BC model for large-scale studies and potential ease in transferability. In  
412 comparison,  $K_s$  and  $f_m$  showed secondary (low) sensitivity on the model performance. Since  
413 this study introduces the BC model, we decided however to take a conservative approach.  
414 We optimized  $K_s$  and  $S_0$  simultaneously for our catchment scale simulations since this  
415 combination resulted in the best overall model performance.

416 At the catchment scale, the proposed BC parameterization improved simulated latent heat  
417 flux (especially in summer) and the overall runoff compared to *default*. Note that the  
418 complexity (i.e., number of parameters) of the BC model for simulating water flow through  
419 chalk unsaturated zone is substantially lower compared to more commonly used models for  
420 this purpose (e.g., dual-porosity models). Despite its simplicity, the proposed  
421 parameterization considerably improves the key hydrological fluxes simulated by JULES at  
422 the catchment-scale. Therefore, the BC model can potentially be useful for land surface  
423 modelling applications over large-scale chalk-dominated areas.

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435 that added to the quality of this manuscript.

## 436 **Appendix**

### 437 **Definition of Statistical Metrics**

438 Coefficient of determination ( $R^2$ ) for observation  $y = y_1, \dots, y_n$  and prediction  $f = f_1, \dots, f_n$   
439 is defined as

$$440 \quad R^2 = 1 - \frac{SS_{res}}{SS_{tot}}$$

441 where  $SS_{res}$  is the residual sum of square and  $SS_{tot}$  is the total sum of square.  $SS_{res}$  and  $SS_{tot}$   
442 are defined as

$$443 \quad SS_{res} = \sum_{i=1}^n (y_i - f_i)^2 \quad \text{and}$$

$$444 \quad SS_{tot} = \sum_{i=1}^n (y_i - \bar{y})^2 \quad \text{with } \bar{y} \text{ being the mean of } y.$$

445 Runoff ratio (RR) assesses the portion of precipitation that generates runoff over the  
446 catchment. RR is defined as

$$447 \quad RR = \frac{\mu_{runoff}}{\mu_{rain}}$$

448 where  $\mu_{runoff}$  is mean runoff and  $\mu_{rain}$  is mean precipitation [e.g., *Kelleher et al.*, 2015].

449 Relative bias ( $\Delta\mu$ ) between observed and simulated time series can be defined as

$$450 \quad \Delta\mu = \frac{\mu_{mod} - \mu_{obs}}{\mu_{obs}}$$

451 where  $\mu_{obs}$  and  $\mu_{mod}$  are the mean of observed and simulated time series, respectively. While  
452 the optimal value of  $\Delta\mu$  is zero, negative (positive) values indicate an underestimation  
453 (overestimation) by the model [e.g., *Gudmundsson et al.*, 2012].

454 Relative difference in standard deviation ( $\Delta\sigma$ ) between observed and simulated time series  
455 can be defined as

$$456 \quad \Delta\sigma = \frac{\sigma_{mod} - \sigma_{obs}}{\sigma_{obs}}$$

457 where  $\sigma_{obs}$  and  $\sigma_{mod}$  are the standard deviation of observed and simulated time series,  
458 respectively [e.g., *Gudmundsson et al.*, 2012].

459

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604

605 **Tables**

606 Table 1. Field measurements and remote sensing data.

Data	Spatial scale	Temporal extent	Frequency	Source
Soil moisture	Point <sup>a</sup>	2003-2005	15 day	N. Hewitt (CEH)
Latent heat flux	Global	2006-2011	8 day, 1 month	MODIS
Discharge	Point <sup>b</sup>	2006-2011	1 day	NRFA

607 <sup>a</sup>Measured at Warren Farm.

608 <sup>b</sup>Locations are shown in Figure 1a.

609 Table 2. Hydraulic properties for different soil types (refer to Figure 1c). Saturated hydraulic  
 610 conductivity ( $K_s$ ) and porosity data are obtained from *Rawls et al.* [1982]. The Van Genuchten  
 611 parameters are acquired from *Schaap and Leij* [1998].

Texture	$K_s$ (mmd <sup>-1</sup> )	Porosity (-)	$\alpha$ (m <sup>-1</sup> )	$n$ (-)
Loam	320	0.463	3.33	1.56
Silt loam	172	0.50	1.2	1.39
Clay	15	0.475	2.12	1.2

612

613 Table 3. Hydraulic properties of chalk

Properties	Uncalibrated		Range for calibration	Calibrated value
	Value	Source		
$K_s$ (mmd <sup>-1</sup> )	1.0	Price et al., 1993	0.2 - 5.0	0.31
$S_0$ (-)	0.8	Observations	0.0 - 1.0	0.46
$f_m$ (-)	$10^5$	Price et al., 1993	$10^4 - 10^6$	$10^{5*}$
$\alpha$ (m <sup>-1</sup> )	3.0	Le Vine et al., 2016	-	-
$n$ (-)	1.4	Le Vine et al., 2016	-	-

614 \*  $f_m$  parameter not calibrated

615 Table 4. Comparison between observed and simulated daily average runoff from the two  
 616 configurations over the Kennet catchment. Metrics include the Runoff Ratio (RR), relative bias ( $\Delta\mu$ ),  
 617 and relative difference in standard deviation ( $\Delta\sigma$ ) (refer to appendix for further information).

Metric	Observed	Simulated ( <i>default</i> )	Simulated ( <i>macro</i> )
RR	0.40	0.82	0.46
$\Delta\mu$	-	1.04	0.12
$\Delta\sigma$	-	2.04	0.65

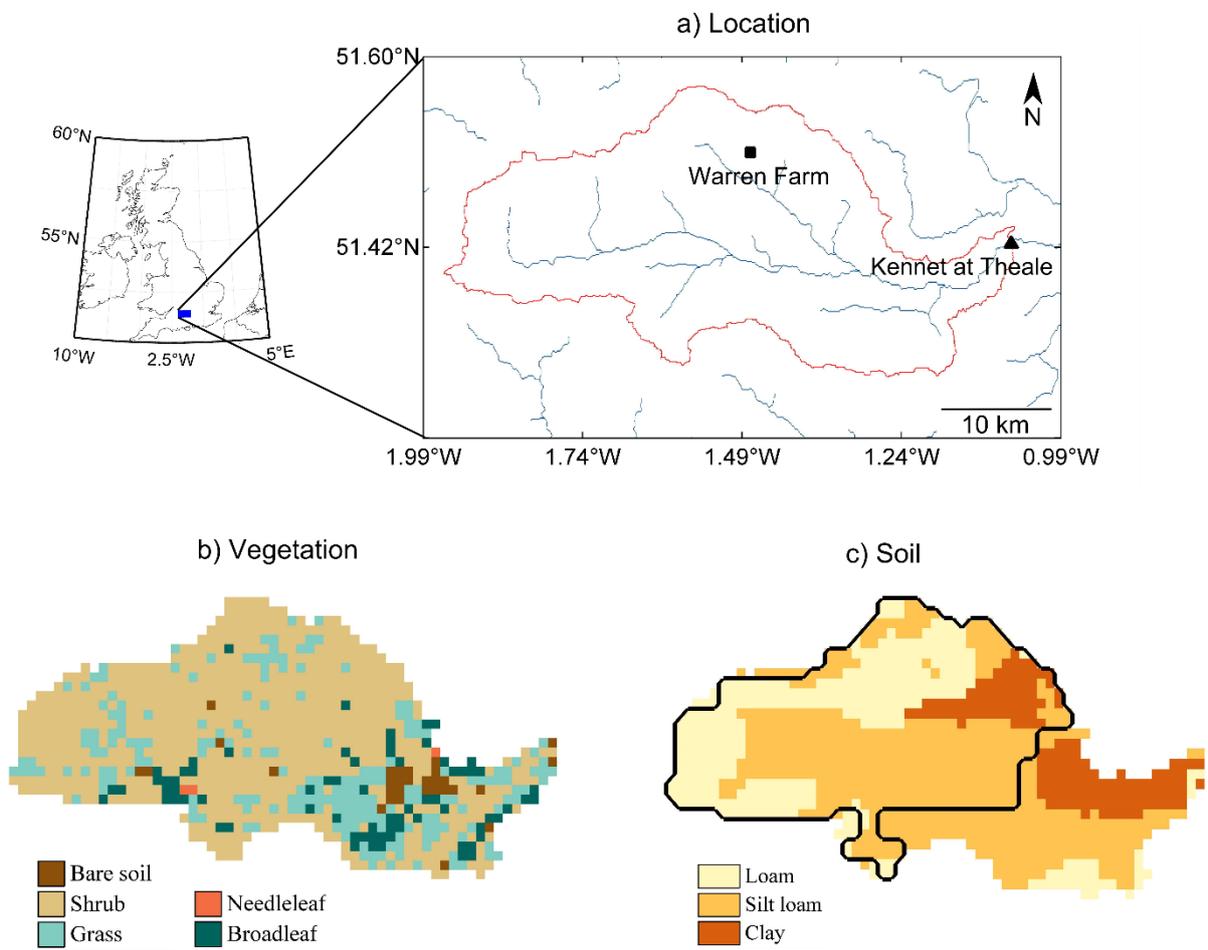
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620 **Figures**

621 Figure 1. (a) Location, (b) vegetation cover and (c) soil texture over the study area. The red  
622 line in (a) outlines the Kennet catchment boundary, while the river network is shown in blue.  
623 The black triangle in (a) shows the location of the discharge gauging station at the catchment  
624 outlet and the black square corresponds to Warren Farm location where point-scale  
625 simulations are carried out. The black line in (c) encloses the area of the catchment where  
626 chalk is present.

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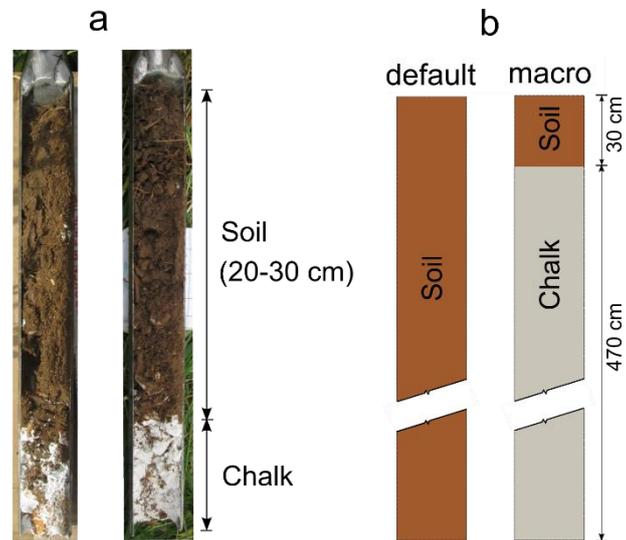
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631 Figure 2. (a) Example of soil profiles collected at Warren Farm during a field campaign in  
632 2015 and (b) the two model configurations.

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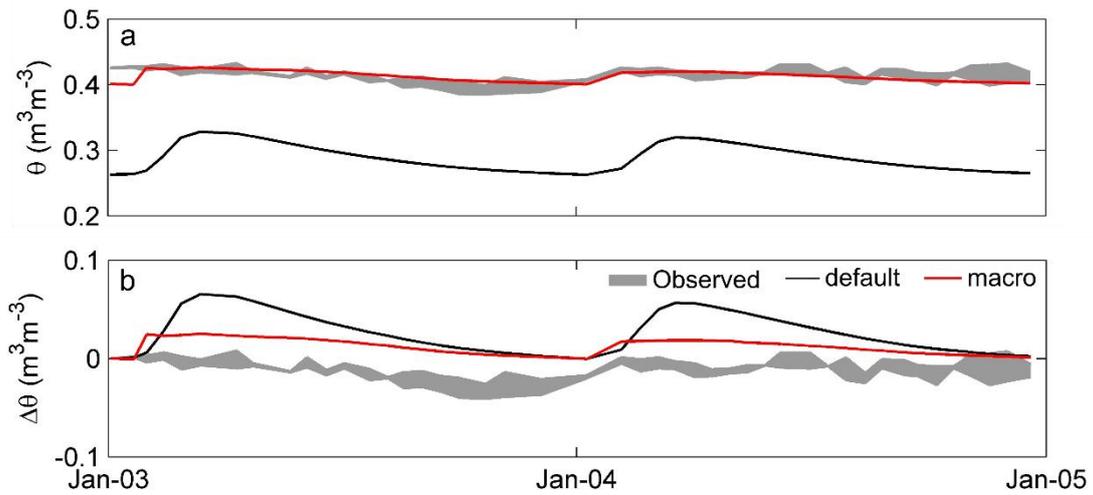
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646 Figure 3. Comparison between observed and simulated (a) soil moisture ( $\theta$ ) and (b) change in  
647 soil moisture ( $\Delta\theta$ ) from the *default* and *macro* configurations at a depth of 2 m below land  
648 surface at the Warren Farm site. The shaded areas constructed from 2 soil moisture probes at  
649 the Warren Farm site denote the range of observed data in these plots.

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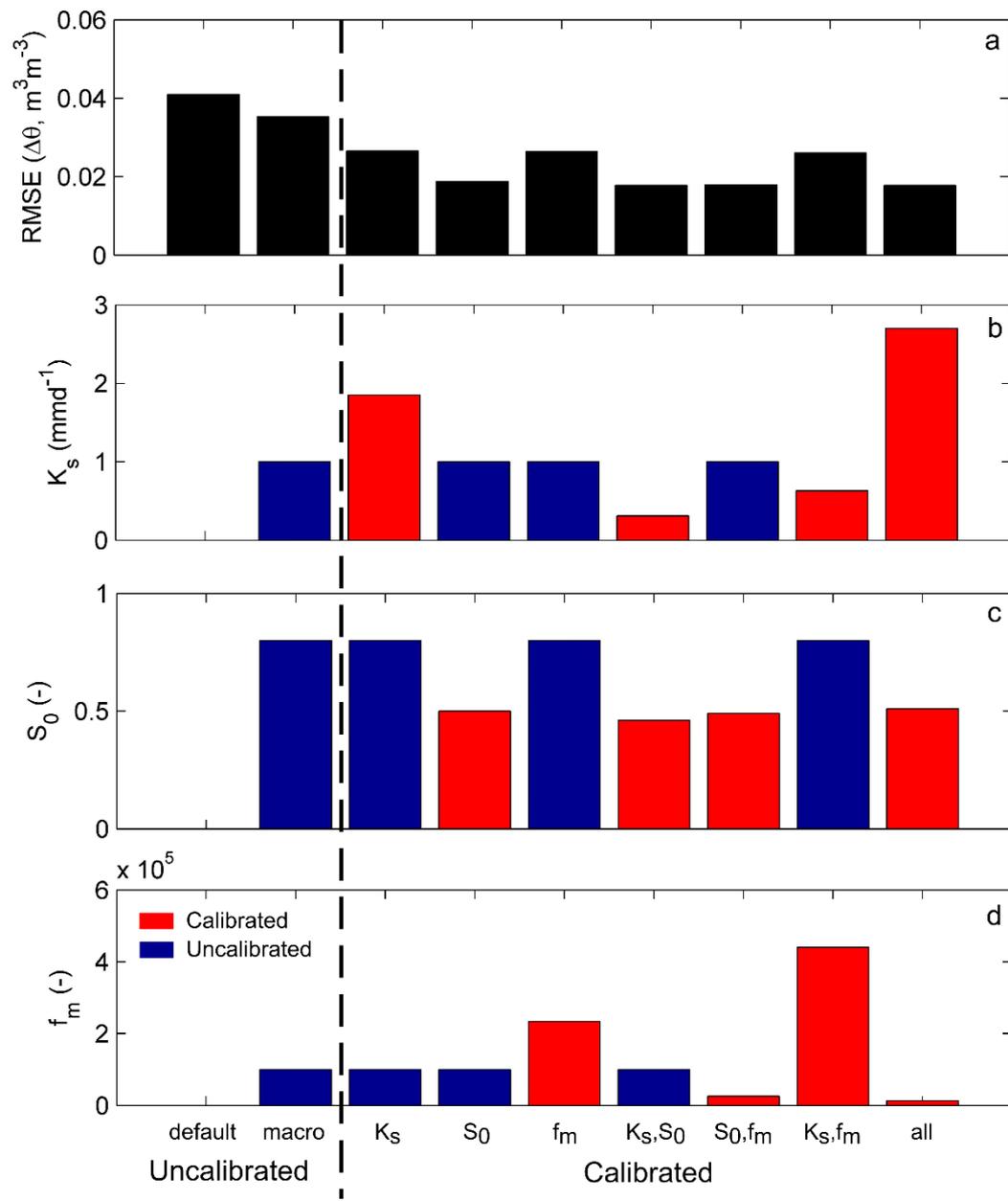
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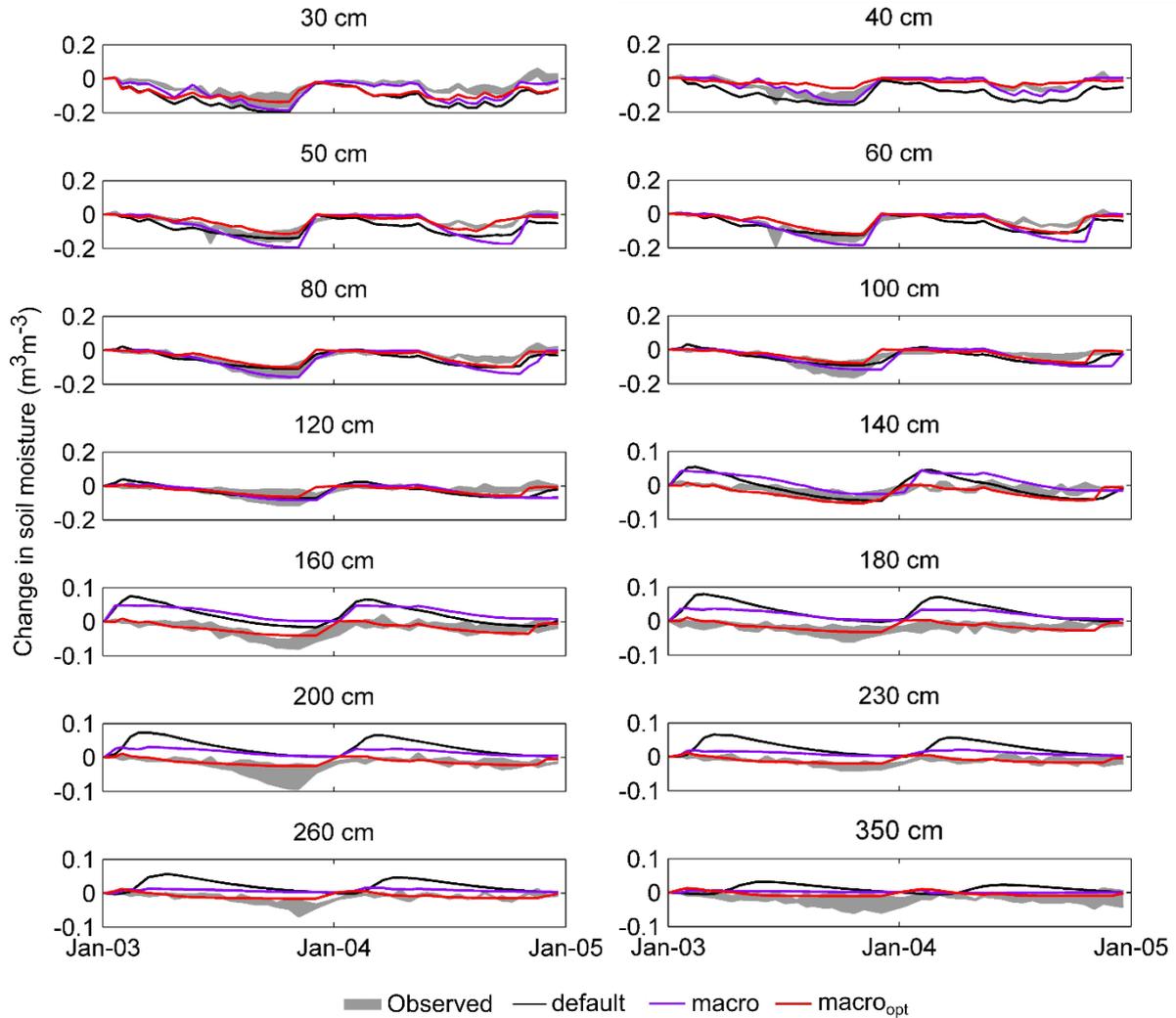
661 Figure 4. (a) Model performance in reproducing observed and simulated  $\Delta\theta$ , (b)  $K_s$ , (c)  $S_0$  and  
 662 (d)  $f_m$  for different parameter combinations considered in the optimization. For each  
 663 parameter (i.e., panels b, c, and d), red bars show cases in which the relevant parameter is  
 664 calibrated (either individually or in combination with others); while the blue bars correspond  
 665 to cases in which the selected parameter is not calibrated (i.e., fixed value according to  
 666 literature as in the *macro* case). Note that except for the *default* and *macro*, the simulation  
 667 yielding the lowest RMSE (out of 2,000 model runs) is presented in this plot.



668

669 Figure 5. Comparison between observed and simulated  $\Delta\theta$  from *default*, *macro* and *macro<sub>opt</sub>*  
 670 configurations at various depths below land surface. The shaded areas, which are constructed  
 671 from 2 soil moisture probes at the Warren Farm site, denote the range of  $\Delta\theta$ .

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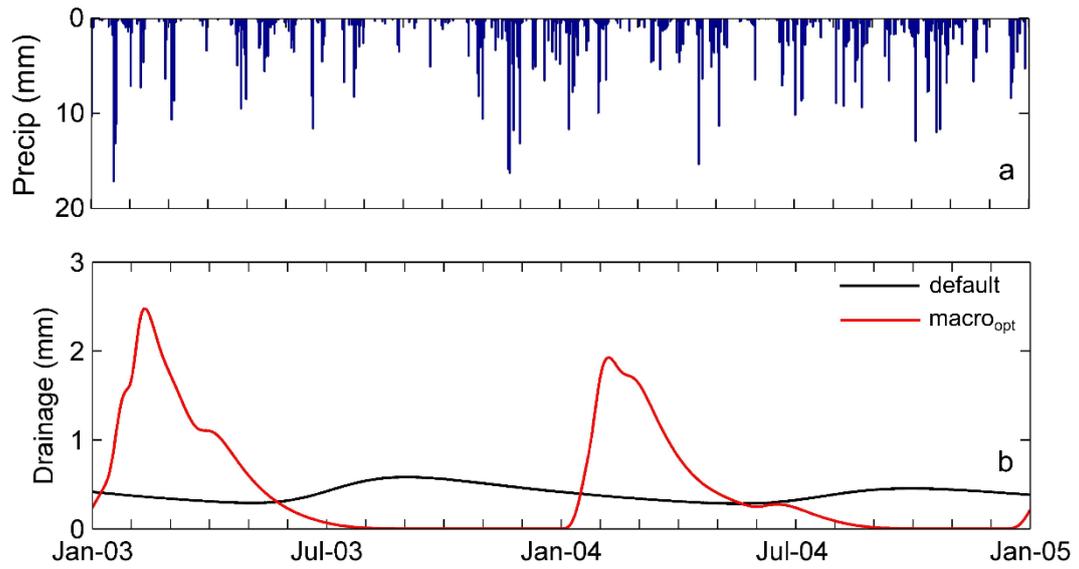
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679 Figure 6. (a) Daily precipitation and (b) daily drainage through the bottom of the soil column  
680 at Warren Farm over the two simulated years (2003-2005).

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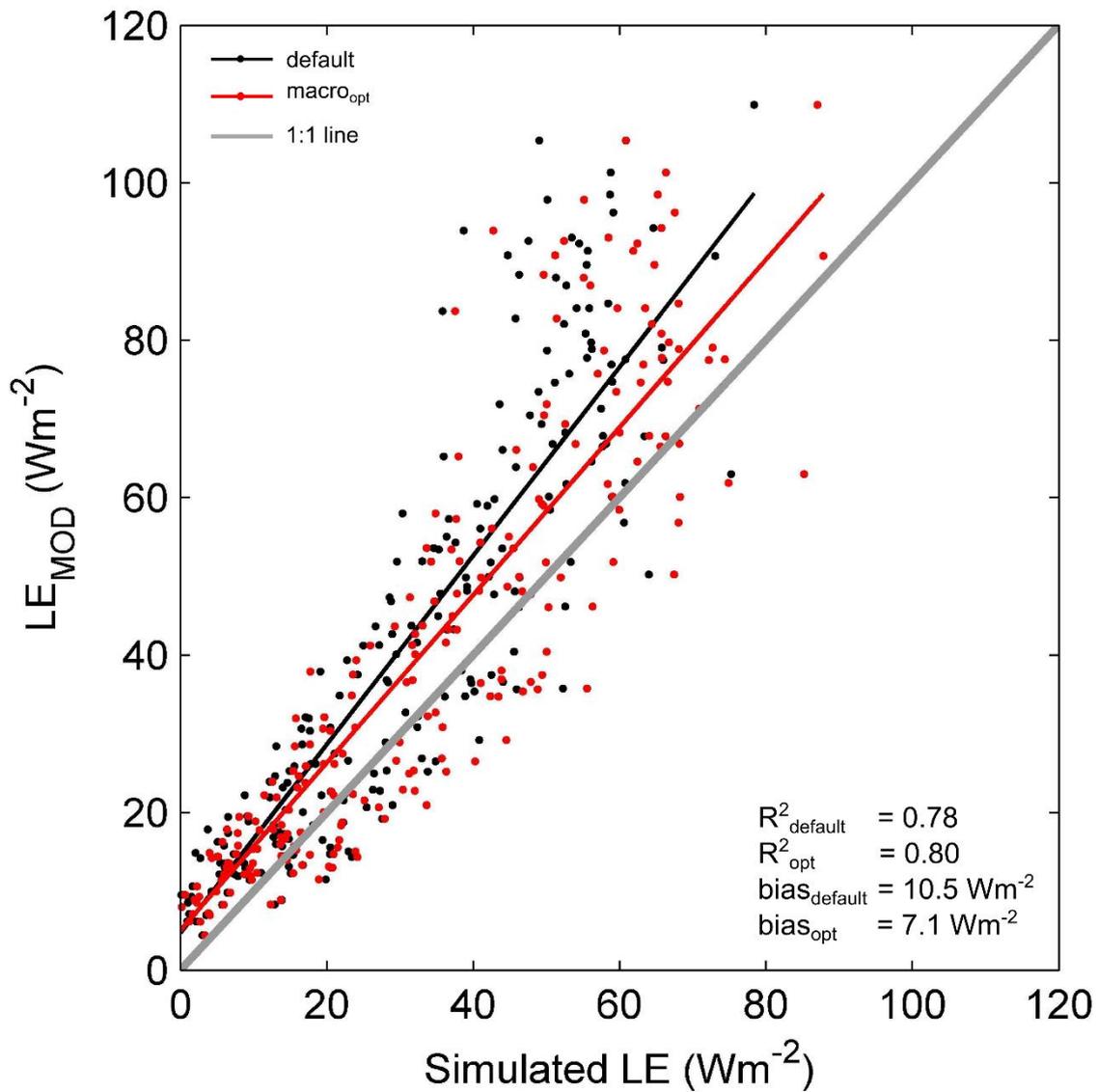
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693 Figure 7. Catchment average 8 day composites of MODIS estimated  $LE$  ( $LE_{MOD}$ ) against  
694 simulated  $LE$  from *default* and *macro* configurations ( $LE_{default}$  and  $LE_{macro}$ , respectively) along  
695 with the linear models fitted for  $LE_{default}$  (black line) and  $LE_{macro}$  (red line). The 1:1 line is  
696 shown in grey, which represents the perfect fit between  $LE_{MOD}$  and simulated  $LE$ .

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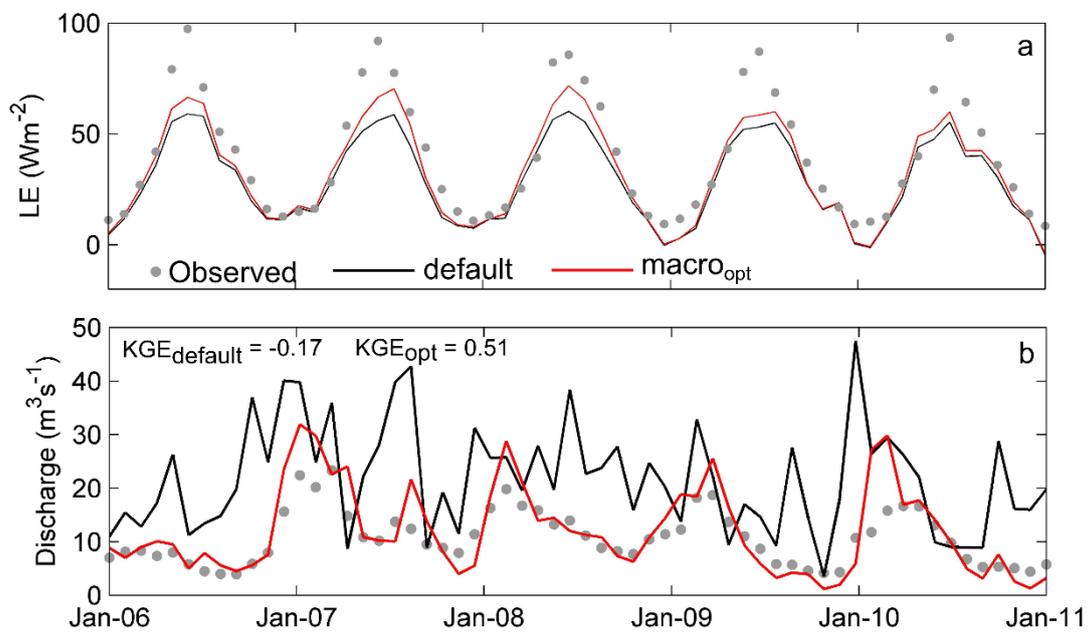
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701 Figure 8. (a) Spatially averaged monthly latent heat flux ( $LE$ ) from MODIS, *default* and  
702 *macro<sub>opt</sub>* configurations over the Kennet catchment and (b) monthly average observed and  
703 simulated discharge from the *default* and *macro<sub>opt</sub>* configurations at the “Kennet at Theale”  
704 gauging station.

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