



1 Identifying uncertainties in simulated streamflow from hydrologic 2 model components for climate change impact assessments

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8 **Abstract:** Assessing the impacts of climate change on hydrologic systems is critical for developing
9 adaptation and mitigation strategies for water resource management, risk control and ecosystem
10 conservation practices. Such assessments are commonly accomplished using outputs from a
11 hydrologic model forced with future precipitation and temperature projections. The algorithms
12 used in the hydrologic model components (e.g., runoff generation) can introduce significant
13 uncertainties in the simulated hydrologic variables, yet the identification and quantification of such
14 uncertainties is rarely studied. Here, a modeling framework is developed that integrates multiple
15 runoff generation algorithms with a routing model and associated parameter optimizations. This
16 framework is able to identify uncertainties from both hydrologic model components and climate
17 forcings as well as associated parameterization. Three fundamentally different runoff generation
18 approaches: runoff coefficient method (RCM, conceptual), variable infiltration capacity (VIC,
19 physically-based, infiltration excess) and simple-TOPMODEL (STP, physically-based, saturation
20 excess), are coupled with Hillslope River Routing model to simulate streamflow. A case study
21 conducted in Santa Barbara County, California, reveals that the median changes are 1-10%
22 increases in mean annual discharge (Q_m) and 10-40% increases in annual maximum daily
23 discharge (Q_p) and 100-yr flood discharge (Q_{100}). The Bayesian Model Averaging analysis
24 indicates that the probability of increase in streamflow can be up to 85%. However, the simulated
25 discharge uncertainties are large (i.e., 230% for Q_m and 330% for Q_p and Q_{100}) with general
26 circulation models (GCMs) and emission scenarios accounting for more than half of the total
27 uncertainty. Hydrologic process models contribute 10-30% of the total uncertainty, while
28 uncertainty due to hydrologic model parameterization is almost negligible (<1%), limiting the
29 impacts of hydrologic model parameter equifinality in climate change impact analysis. This study
30 also provides insights on how to optimize the selection of hydrologic models for projecting future
31 streamflow conditions.



32 1. Introduction

33 Streamflow is essential to humans and ecosystems, supporting human's life and economic
34 activities, providing habitat for aquatic creatures, and exporting sediment/nutrients to coastal
35 ecosystems (Feng et al., 2016; Barnett et al., 2005; Milly et al., 2005). Understanding streamflow
36 characteristics is important for water-resources management, civil infrastructure design and
37 making adaptation strategies for economic and ecological practices (Feng et al., 2019). With
38 economic development and population growth, the emission of greenhouse gas is likely to increase
39 during 21st century (IPCC, 2014). The increase in global surface temperature is projected to exceed
40 2°C by the end of 21st century even under moderate emission scenarios (e.g., Representative
41 Concentration Pathways, RCPs, 4.5 and 6.0) (IPCC, 2014). Intensified hydro-meteorological
42 processes, altered precipitation forms and patterns, and intensified atmospheric river events and
43 oceanic anomalies (e.g. El Nino events) are projected and likely to causes substantial impacts on
44 hydrologic fluxes (e.g., streamflow) (Barnett et al., 2005; Tao et al., 2011; Dai, 2013; Dettinger,
45 2011; Vicky et al., 2018; Cai et al., 2014; Feng et al., 2019).

46 The integration of climate projections and hydrologic models enables the investigation of
47 streamflow dynamics under the future climate conditions. However, the simulated streamflow
48 contains uncertainties from various sources. Due to epistemic limitations (e.g., human's lack of
49 knowledge about hydrologic processes and boundary conditions) and the complexities in nature
50 (e.g., temporal and spatial heterogeneity), hydrologic models are simplified representations of
51 natural hydrologic processes (Beven and Cloke, 2012). Generally, hydrologic models have
52 modules simulating atmosphere-land interactions associated with water and energy partitioning
53 (named as runoff generation process in this study), and modules simulating the water
54 transportation along terrestrial hillslopes and channels (named as routing process here). Each
55 process can be represented in different ways, which thus results in uncertainties in simulated
56 streamflow. For the runoff generation process, surface runoff is mainly represented as infiltration
57 excess overland flow (or Hortonian flow (Horton, 1933)) or saturation excess overland flow.
58 Infiltration excess overland flow occurs when water falls on the soil surface at a rate higher than
59 that the soil can absorb. Saturation excess overland flow occurs when precipitation falls on
60 completely saturated soils. In addition, surface runoff can also be quantified conceptually, for
61 example, a runoff coefficient can be used to generate surface runoff as a proportion of precipitation



62 rate. Subsurface runoff is generally represented as functions of soil characteristics and topographic
63 features. The complexity of these functions varies significantly, from simple linear to
64 combinations of multiple non-linear. The lateral routing processes are generally represented using
65 various approximations of the Saint-Venant equations (Reed et al., 2004). Different choices of
66 these process models may achieve different results and thus cause uncertainties in outputs.
67 Parameterization can be another uncertainty source. Due to the nonlinearity of hydrologic
68 processes, different combinations of model parameters can achieve similar, if not identical, model
69 performance. Model parameter selections based on statistical metrics obtained from calibration
70 can result in different optimal parameter values (i.e., parameter equifinality). When it comes to
71 hydrologic impact assessments, the model forcings, which differ among General Circulation
72 Models (GCMs) due to the model discrepancy and the uncertainty of future emission scenarios,
73 also contribute to the uncertainties in simulated streamflow. Without appropriate assessment of
74 these uncertainties, standalone studies on the climate change impacts, using a particular hydrologic
75 model forced by select GCMs' projections under some emission scenarios, can be difficult to
76 interpret. Systematic assessments of the relevant uncertainties associated with simulated
77 streamflow are needed.

78 Some studies have been performed considering the above at both regional and global scales
79 (for example, (Wilby and Harris, 2006;Vetter et al., 2015;Valentina et al., 2017;Kay et al.,
80 2009;Eisner et al., 2017;Su et al., 2017;Schewe et al., 2014;Hagemann et al., 2013;Asadieh and
81 Krakauer, 2017)). Most previous studies integrated multiple hydrologic models individually.
82 However, the hydrologic model structures can be significantly different, which may limit the
83 ability to quantify relative uncertainty contributions from different model components (e.g., runoff
84 generation model and routing model) and associated parametrizations. Troin et al. (2018) tested
85 the impacts by using different hydrologic model components to simulate streamflow, but they only
86 focused on snow and potential ET methods. In this study, a consistent hydrologic modeling
87 framework that integrates multiple runoff generation process models with surface, subsurface and
88 channel routing processes and associated parameter uncertainties is developed. This framework
89 enables uncertainties from different components representing hydrologic processes and associated
90 model parameters as well as model forcings (e.g., precipitation and temperature) to be quantified
91 and compared in a consistent manner. In this framework, three runoff generation process models
92 which represent the three fundamentally different approaches mentioned above are used. The



93 conceptual frameworks are adapted from the variable infiltration capacity model (Wood et al.,
94 1992;Liang et al., 1996) (infiltration excess), simple-TOPMODEL (Niu et al., 2005) (saturation
95 excess), and the runoff coefficient method (Feng et al., 2019) (conceptual). Each approach is
96 coupled within one routing model (i.e., Hillslope River Routing model, HRR (Beighley et al.,
97 2009)) to investigate the impacts of model structures and associate parameters on simulated
98 streamflow. Compared to runoff generation process models, routing process models have less
99 variants with most models using approximations of the Saint-Venant equations (Reed et al., 2004).
100 Therefore, only one routing model is included in this study, however, this modeling framework is
101 suitable to integrate different routing process models (e.g., diffusive wave and full dynamic
102 solutions) and runoff schemes in future studies. This modeling framework is also coupled with a
103 Bayesian model averaging (BMA) analysis to assess the performance of different model-forcing-
104 parameter combinations and to provide actionable information (e.g., probability of estimated
105 changes) for associated practices, such as water resource management and ecology conservation.

106 A case study is presented for Santa Barbara County, CA, a biodiverse region under a
107 Mediterranean climate with a mix of highly developed and natural watersheds. To estimate future
108 streamflow and associate uncertainties, the hydrologic models are forced with climate projections
109 from 10 GCMs selected for their good performance in representing historical meteorological
110 characteristics in the study region, under 2 emission scenarios (RCP 4.5 and RCP 8.5) (Feng et al.,
111 2019). The main objectives of this study are to: (1) evaluate and compare the performance of
112 hydrologic models with different approaches representing runoff generation process using a
113 consistent modeling framework; (2) quantify the relative contributions of different sources
114 (including hydrologic process models, parameterizations, GCM forcings and emission scenarios)
115 to the total uncertainty in simulated streamflow; and (3) provide actionable information and
116 suggestions for studies and practices associated with the hydrologic impacts of climate change in
117 the study region.

118 2. Methods

119 2.1 Study region

120 The study region is located in coastal Santa Barbara County (SBC), California, where
121 watersheds drain into the Santa Barbara Channel from just west of the Ventura River to just east



122 of Point Conception (Fig. 1). The combined land area is roughly 750 km² with 135 watersheds
123 ranging from 0.1 to 123 km². The local climate is Mediterranean, with an average annual
124 precipitation of roughly 600 mm (Feng et al., 2019). Most of the annual precipitation occurs in
125 fall/winter with 85% of rainfall occurring in the November-March period. Thus, it is characterized
126 by the intense and flashy floods in winter time. More than 80% of annual discharge occurs in only
127 a few number of large events during January-March and a large fraction of annual discharge
128 happens within one day (Beighley et al., 2003). River channels are typically filled with sediment
129 during dry season (April-October) and are scoured with the initiation of wet season floods (Scott
130 and Williams, 1978; Keller and Capelli, 1992). River flow is the major source of sediment exported
131 to the coastal sandy beaches in SBC. Therefore, the timing of seasonality and magnitudes of flood
132 events are critical to both local community and coastal ecosystems.

133 2.1 Data

134 Daily precipitation and temperature with a spatial resolution of 0.0625° x 0.0625° (roughly
135 6 by 6 km) (Livneh et al., 2015), and daily streamflow from 4 USGS gauges for the period 1984-
136 2013 are used to calibrate and validate the hydrologic model. The Global Soil Dataset for use in
137 Earth system models (GSDE) is used to estimate saturated hydraulic conductivity and saturated
138 moisture content. The 16-day composite albedo product (MCD43C3) with a spatial resolution of
139 0.05° x 0.05° and the monthly aerosol optical depth product (MOD08M3) with a spatial resolution
140 of 1.0° x 1.0° both derived from NASA's Moderate Resolution Imaging Spectroradiometer
141 (MODIS) are used to determine net radiation for evapotranspiration (PET) estimation.

142 For the historical (1986-2005) and future climate simulations (2081-2100), downscaled
143 precipitation and temperature from ten climate models (please refer to Pierce et al. (2014) and
144 Pierce et al. (2015) for model details) in Coupled Model Inter-Comparison Project, Phase 5,
145 (CMIP5) (Taylor et al. 2012) for two emission scenarios RCP 4.5 and RCP 8.5 (Moss et al. 2010)
146 are used. These 10 GCMs are selected because they have the best performance in representing
147 historical climate dynamics at southwest U.S. and California state scales (Pierce et al., 2018).

148 2.2 Hydrologic modeling framework

149 2.2.1 Hydrologic model development



150 This modeling framework is developed on the basis of the Hillslope River Routing model
151 (HRR) (Beighley et al., 2009). The watershed is delineated based on the Digital Elevation Model
152 (DEM). The sub-basins are irregular-shape catchments defined by the flow accumulation area
153 threshold. In this study, the threshold is 1 km^2 , which means the sub-basins (model units) are in
154 size of roughly 1 km^2 . The hydrogeological model inputs, including surface roughness, saturated
155 hydraulic conductivity, soil thickness, porosity, plane slope, channel slope and channel roughness,
156 are averaged over each sub-basin. The geometry of each sub-basin (plane length and width) is
157 calculated based on an “open-book” assumption, which assumes each sub-basin is a rectangular
158 divided by the river channel into two identical parts like an open book. Please refer to Beighley et
159 al. (2009) for more details. The precipitation and ET are extracted from the grid-based datasets for
160 each sub-basin using an area-weighted average method. Then the water-balance model (i.e., runoff
161 generation method) is applied to each model unit to simulate runoff generation processes. Here,
162 three runoff generation methods: runoff coefficient (Feng et al., 2019), and the methods used in
163 Variable Infiltration Capacity (VIC) (Wood et al., 1992;Liang et al., 1996) and simple-
164 TOPMODEL model (Niu et al., 2005), are used to simulate the generation of surface and
165 subsurface runoff excess. The routing methods within the HRR model (i.e., kinematic wave for
166 surface and subsurface lateral routing and Muskingum-Cunge for channel routing) are used to
167 simulate the transport of runoff excess. To clarify, we denote the three runoff generation
168 algorithms: runoff coefficient, runoff generation method used in Variable Infiltration Capacity and
169 runoff generation method used in simple-TOPMODEL as RCM, VIC and STP, respectively. The
170 three hydrologic models which integrate each of these runoff generation methods with the routing
171 method used in HRR model are referenced as RCM-HRR, VIC-HRR and STP-HRR, respectively.
172 The differences between simulations from these three models are considered as the uncertainty
173 resulting from hydrologic model formulation. The three runoff generation algorithms, and the
174 surface, subsurface and channel routing are described below.

175 The RCM assumes water excess available for surface runoff (e_s) is proportional to
176 precipitation rate (P). The proportion is represented by a coefficient value (e.g., 0 to 100%) and is
177 dependent on land cover, soil and topographic characteristics. The coefficient value is smaller for
178 dry and flat areas with permeable soils and vegetated surfaces, as compared to that for wet and
179 steep areas with more impervious areas (e.g., roads, parking lots, roofs). In this work, a dual
180 runoff-coefficient method is used, which assigns a larger runoff coefficient (C_2) to wet soils



181 (relative soil moisture at upper soil layer $\theta_U \geq$ threshold θ_t) and smaller runoff coefficient (C_1) to
 182 dry soils (relative soil moisture $\theta_U <$ threshold θ_t) (Eq. (1)). The water excess available for
 183 subsurface runoff (e_{ss}) is a function of saturated hydraulic conductivity (k_{sat}) and relative soil
 184 moisture in lower soil layer (θ_L) (Eq. (2)).

$$\begin{aligned} e_s &= C_1 \times P \quad \text{for } \theta_U < \theta_t \\ &= C_2 \times P \quad \text{for } \theta_U \geq \theta_t \end{aligned} \quad (1)$$

$$e_{ss} = K_{sat_all} k_{sat} \times \left(\frac{\theta_L}{n}\right)^b \quad (2)$$

185 where e_s and e_{ss} are water excess available for surface and subsurface runoff, respectively, ($m \, d^{-1}$)
 186 1); P is precipitation rate ($m \, d^{-1}$); C_1 is dry runoff coefficient; C_2 is wet runoff coefficient; θ_U and
 187 θ_L are relative soil moisture at upper and lower soil layer, respectively; θ_t is relative soil moisture
 188 threshold differentiating dry and wet soil conditions; k_{sat} is saturated hydraulic conductivity (m
 189 d^{-1}); K_{sat_all} is a scaler; b is Clapp-Hornberger parameter and n is soil porosity. C_1 , C_2 , θ_t and
 190 K_{sat_all} are parameters needing calibration.

191 In the VIC algorithm, surface runoff is generated as infiltration excess where the infiltration
 192 rate is characterized by the variable infiltration curve (Wood et al., 1992). In this work, the
 193 framework of modified 2-layer VIC model (VIC-2L) (Liang et al., 1996) is used. The water excess
 194 available for surface runoff is calculated as shown in Eq. (3)-(4). The water excess available for
 195 subsurface runoff is a function of soil moisture in lower soil layer (Eq. (5)), which is a linear
 196 function of soil moisture when the soil is relatively dry and quadratic when the soil is close to
 197 saturation:

$$e_s = P - z(\theta_s - \theta_U)/\Delta t - z\theta_s \left(\max \left[0, \left[1 - \frac{i_o + P\Delta t}{i_m} \right] \right] \right)^{1+b_i} / \Delta t \quad (3)$$

$$i_o = i_m [1 - (1 - A)^{1/b_i}] \quad (4)$$

$$e_{ss} = \frac{D_S D_M}{W_S \theta_S} \theta_L + \left(D_M - \frac{D_S D_M}{W_S} \right) \left(\frac{\max |0, \theta_L - W_S \theta_S|}{\theta_S - W_S \theta_S} \right)^2 \quad (5)$$

198 where z is soil depth in upper layers (m); θ_s is relative soil moisture at saturation; i_m is maximum
 199 infiltration capacity (m); i_o is infiltration capacity (m); b_i is infiltration curve parameter; A is the
 200 fraction of saturation; D_M is maximum base flow ($m \, d^{-1}$); D_S is the fraction of D_M at which the non-



201 linear base flow begins; W_s is the fraction of saturation at which the non-linear base flow
202 occurs; Δt is time step (d). b_i , D_M , D_s and W_s are parameters which need calibration.

203 In STP algorithm, the surface runoff is generated as saturation excess overland flow (Eq.
204 (6)). The saturation fraction of the catchment f_{sat} is determined as a function of topographic index
205 (Eq. (7)-(8)).

$$e_s = f_{sat} * P \quad (6)$$

$$f_{sat} = f_{max} * \exp(-0.5 z_{\nabla} f_{over}) \quad (7)$$

206 where f_{sat} is the fraction of saturated area; f_{over} is a decay factor for surface runoff water excess
207 (m^{-1}); z_{∇} is groundwater table depth (m); f_{max} is the maximum saturated fraction and is defined
208 as the percent of grid cells in each sub-basin with a topographic index (τ) that is \geq the mean τ
209 determined by averaging all grid cell τ values:

$$\tau = \ln\left(\frac{a}{\tan(\beta)}\right) \quad (8)$$

210 where a is the specific catchment area (i.e., upslope area per unit contour length) and β is the pixel
211 slope. The specific catchment area a and slope β are calculated for grid cell using the gridded
212 elevation data and the TauDEM tools (Tarboton, 2003).

213 The water excess available for subsurface runoff is a function of maximum base flow rate
214 and groundwater table depth:

$$e_{ss} = Q_m * \exp(-f_{drain} * z_{\nabla}) \quad (9)$$

215 where f_{drain} is a decay factor for subsurface runoff water excess (m^{-1}), and Q_m is the maximum
216 baseflow rate ($m d^{-1}$). Water excess for both surface and subsurface runoff are dependent of the
217 groundwater table depth z_{∇} . Here, the water table depth z_{∇} is determined by applying the method
218 used in (Niu et al., 2005), which assumes the water head at depth z is in equilibrium with that at
219 ground water depth z_{∇} (Eq. (10)-(13)).

$$\varphi(z) - z = \varphi_{sat} - z_{\nabla} \quad (10)$$



220 where $\varphi(z)$ and φ_{sat} are the metric potentials at depth z and at groundwater table depth z_{∇} (m).
221 The soil at the groundwater table depth is assumed to be saturated. Based on Clapp-Hornberger
222 relationship (Clapp and Hornberger, 1978), $\varphi(z)$ can be expressed as:

$$\varphi(z) = \varphi_{sat} \left(\frac{\theta(z)}{\theta_{sat}} \right)^{-b} \quad (11)$$

223 where $\theta(z)$ and θ_{sat} are soil moisture content at depth z and groundwater table depth z_{∇} ,
224 respectively, b is a Clapp-Hornberger parameter. By substituting Eq. (10) with Eq. (11), the soil
225 matric profile at depth z can be expressed as:

$$\theta(z) = \theta_{sat} \left(\frac{\varphi_{sat} - (z_{\nabla} - z)}{\varphi_{sat}} \right)^{-1/b} \quad (12)$$

226 Then, the groundwater table depth (z_{∇}) can be determined by solving Eq.13 iteratively.

$$D_{\theta} = \int_0^{z_{\nabla}} (\theta_{sat} - \theta(z)) dz \quad (13)$$

227 where D_{θ} is the soil moisture deficit, which can be calculated in Eq.14:

$$D_{\theta} = \sum_{i=1}^m (\theta_{sat} - \theta_i) \nabla z_i \quad (14)$$

228 where θ_i is the soil moisture content at the i^{th} soil layer; ∇z_i is the soil thickness of i^{th} soil layer, m
229 is the number of soil layer, $m=2$ in this study. In STP algorithm, f_{over} , f_{drain} , Q_m and φ_{sat} are
230 parameters to be calibrated.

231 The water movement between soil layers in the soil matrix is similar to that in the modified
232 VIC-2L model (Liang et al., 1996). The soil is divided into 2 layers: upper layer (0.6 m) and lower
233 layer (2.6 m). The soil thickness data is determined based on a previous study (Feng et al., 2019).
234 After the surface runoff is determined using the methods mentioned previously, the infiltrated
235 water is added to the upper soil layer, and the soil moisture is updated. If the upper soil is
236 oversaturated, the excess water is returned to surface. The interaction between upper and lower
237 soil layers is determined using the Clapp-Hornberger equation (Eq. (15)-(16)). Subsurface runoff
238 is generated from the bottom of the lower soil layer. The water flux from the upper layer does not



239 contribute to runoff and is only lost to evapotranspiration and/or drainage to the lower soil layer.
240 A conceptual illustration of the runoff generation process for each method and the water movement
241 in soil matrix can be found in *Supporting Information Fig. S1*.

$$K = k_{sat} \times \left(\frac{\theta_U}{n}\right)^c \quad (15)$$

$$D = k_{sat} \times \left(\frac{\theta_L}{n}\right)^c \quad (16)$$

242 where K is the water flux from the upper soil layer to the lower soil layer (m d^{-1}); and D is the
243 water flux transported from the lower soil layer to the upper soil layer due to diffusion (m d^{-1}).

244 After water excess for surface and subsurface runoff is determined, the kinematic wave
245 approach is used to simulate the transport of runoff from the planes (surface and subsurface), and
246 the Muskingum Cunge method is used for channel routing following the below conservation
247 equations (Beighley et al., 2009):

248 Plane Routing:

$$\frac{\partial y_s}{\partial t} + \frac{\partial q_s}{\partial x_p} = e_s \quad (17)$$

$$\frac{\partial y_{ss}}{\partial t} + \frac{\partial q_{ss}}{\partial x_p} = e_{ss} \quad (18)$$

249 Channel Routing:

$$\frac{\partial A_c}{\partial t} + \frac{\partial Q_c}{\partial x_c} = q_s + q_{ss} \quad (19)$$

250 where y_s and y_{ss} are water depth (or thickness) of surface and subsurface runoff, respectively (m);
251 q_s and q_{ss} are surface and subsurface runoff flow rates per unit width of plane ($\text{m}^2 \text{s}^{-1}$); dx_p is the
252 distance step along the plane (m); A_c is the cross section area of flow in the channel (m^2); Q_c is
253 the flow rate in channel ($\text{m}^3 \text{s}^{-1}$); dx_c is the distance step along the channel (m); and dt is the time
254 step (s).

255 2.3.2 Model calibration



256 After the models are setup, a state-of-the-art optimization algorithm, Borg Multiobjective
257 Evolutionary Algorithm (Borg MOEA) (Hadka and Reed, 2013), is adopted to optimize the model
258 parameters (Table 1). The parameters of the three models are calibrated separately. For each
259 model, there are 4 parameters calibrated for runoff generation processes and 2 parameters
260 calibrated for routing processes. K_{s_all} and K_{ss_all} are conceptual parameters which account for
261 spatial heterogeneity at the model unit scale and uncertainties in the hydro-geologic inputs
262 associated with the plane routing processes (e.g., surface roughness and saturated hydraulic
263 conductivity). They can be different for different model structures even for the same study region.
264 Therefore, they are calibrated for each model separately. The Nash–Sutcliffe model efficiency
265 coefficient (NSE) (Eq. (20)) is used to assess model performance, as it accounts for model
266 performance in terms of both timing and magnitudes of peak flow and base flow that are
267 particularly important in this study. The optimal parameter set is determined after the improvement
268 of error is minimized (here it is defined as $\Delta NSE < 0.005$). To quantify the uncertainties from model
269 parameters, 3 optimal parameter sets with similar performance are selected for each model. The
270 selected parameter sets are then used for simulation with different climate forcings.

$$NSE = 1 - \frac{\sum_{t=1}^T (Q_s^t - Q_o^t)^2}{\sum_{t=1}^T (Q_o^t - \overline{Q_o})^2} \quad (20)$$

271 where Q_s^t and Q_o^t are simulated and observed discharge at time t, respectively, ($m^3 s^{-1}$); and $\overline{Q_o}$ is
272 the mean discharge during the study period of length T, ($m^3 s^{-1}$).

273 2.3 Uncertainty Analysis

274 The uncertainty is quantified by running each of the 9 hydrologic model-parameter sets
275 (i.e., 3 hydrologic models and 3 parameter sets, $3 \times 3 = 9$) with each of the 20 forcing sets (i.e., 10
276 GCMs and 2 emission scenarios, $10 \times 2 = 20$) for a total of 180 simulations.

277 To evaluate the uncertainty sources and their relative significance in simulated discharges
278 for the future period, the analysis of variance (ANOVA) (Vetter et al., 2015) is used. The
279 contribution of each uncertainty source for a particular variable (e.g., annual mean discharge,
280 annual peak discharge or 100-yr flood discharge) is defined as the fraction of its variance to the
281 total variance. The total variance is quantified as the total sum of squares (SS_{total}) of differences
282 between the simulations and the mean of all simulations (Eq. (21)):



$$SS_{Total} = \sum_{i=1}^{N_{Hyd}} \sum_{j=1}^{N_{para}} \sum_{k=1}^{N_{GCM}} \sum_{l=1}^{N_{RCP}} (q_{ijkl} - q_{0000})^2 \quad (21)$$

283 where q_{ijkl} is the simulated value of a particular variable by i^{th} hydrologic model with j^{th} parameter
 284 set, forced by k^{th} GCM projection under l^{th} RCP scenario; q_{0000} is the overall average of the
 285 simulated variable. Next, the SS_{Total} can be divided into 15 parts representing the 4 main effects
 286 (or first-order effects), 6 second-order, 4 third-order and 1 fourth-order interaction effects. For
 287 clarity, the third and fourth orders of interaction effects are combined and represented as $SS_{3,4}$ in
 288 Eq. (22).

$$\begin{aligned} SS_{Total} = & SS_{Hyd} + SS_{para} + SS_{GCM} + SS_{RCP} + SS_{Hyd.para} + SS_{Hyd.GCM} \\ & + SS_{Hyd.RCP} + SS_{para.GCM} + SS_{para.RCP} + SS_{GCM.RCP} \\ & + SS_{3,4} \end{aligned} \quad (22)$$

289 where SS_{Hyd} , SS_{para} , SS_{GCM} and SS_{RCP} are the main effects (i.e., uncertainties or variance) from
 290 hydrologic models, parameterization, GCMs and RCPs, respectively; $SS_{Hyd.para}$, $SS_{Hyd.GCM}$,
 291 $SS_{Hyd.RCP}$, $SS_{para.GCM}$, $SS_{para.RCP}$ and $SS_{GCM.RCP}$ are uncertainties from interactions between
 292 the hydrologic models and parameterization, hydrologic models and GCMs, hydrologic models
 293 and RCPs, parameterization and GCMs, parameterization and RCPs, and GCMs and RCPs,
 294 respectively. The calculation of the effect of each order is illustrated in Eq. (23)-(25).

$$SS_{Hyd} = N_{para} N_{GCM} N_{RCP} \sum_{i=1}^{N_{Hyd}} (q_{i000} - q_{0000})^2 \quad (23)$$

$$SS_{Hyd.para} = N_{GCM} N_{RCP} \sum_{j=1}^{N_{para}} \sum_{i=1}^{N_{Hyd}} (q_{ij00} - q_{i000} - q_{0j00} + q_{0000})^2 \quad (24)$$

$$\begin{aligned} SS_{3,4} = & SS_{Total} - (SS_{Hyd} + SS_{para} + SS_{GCM} + SS_{RCP} + SS_{Hyd.para} \\ & + SS_{Hyd.GCM} + SS_{Hyd.RCP} + SS_{para.GCM} + SS_{para.RCP} \\ & + SS_{GCM.RCP}) \end{aligned} \quad (25)$$



295 where q_{i000} is the average of all simulations from the i^{th} hydrologic model with all combinations
296 of parameter sets, GCMs and RCPs; q_{0j00} is the average of all simulations from the j^{th} parameter
297 set with all combinations of hydrologic models, GCMs and RCPs; q_{ij00} is the average of all
298 simulations from the i^{th} hydrologic model and j^{th} parameter set with all combinations of GCMs and
299 RCPs. Other terms in Eq. (22) can be calculated similarly using Eq. (23)-(24).

300 To avoid bias from the difference in sample sizes of uncertainty sources (i.e., 3 hydrologic
301 models, 3 parameter sets, 10 GCMs and 2 RCPs), a subsampling step is performed by following
302 Vetter et al. (2015). In the subsampling step, 2 samples (the minimum number of uncertainty
303 source, here it is RCPs) from each source are randomly selected, that is, 2 hydrologic models, 2
304 parameter sets, 2 GCMs and 2 RCPs, which means N_{Hyd} , N_{para} , N_{GCM} and N_{RCP} in Eq. (21), (23)-
305 (24) are all equal to 2. This produces $C_3^2 \times C_3^2 \times C_{10}^2 \times C_2^2=405$ subsamples. For each subsample,
306 the fractional sum of squares is calculated for each effect using Eq. (23)-(25), and then the average
307 of variance fractions of each source is used as the uncertainty contribution from that source using
308 Eq. (26):

$$\delta_e = \frac{1}{405} \sum_{m=1}^{405} \frac{SS_e(m)}{SS_{Total}(m)} \quad (26)$$

309 where δ_e is the average fractional effect of term e (i.e, each of 11 terms in Eq. (22)); $SS_e(m)$ is
310 the sum of variance of effect e in the m^{th} subsample, and the $SS_{Total}(m)$ is the total variance in
311 the m^{th} subsample. So in this study, there are 11 δ_e values in total, representing the uncertainty
312 contributions of 11 terms in Eq. (22), with a sum of 1.0.

313 2.5 Probability of estimated changes

314 In addition to the quantification of uncertainties and associated contributions from different
315 sources, an evaluation on the probability of uncertain changes in discharge can be useful to provide
316 actionable information for the stakeholders such as water resource managers. In this study, the
317 Bayesian model averaging (BMA) (Duan et al., 2007) is used to evaluate the model performance
318 in reproducing historical hydrologic conditions and then assign weights to each of them based on
319 their performance. A model with better performance will be assigned a higher weight, which
320 assumes it has a higher probability to be the truth. Note, there is no RCPs for historical period, so
321 only combinations of hydrologic models, parameter sets and GCMs ($3 \times 3 \times 10=90$) are evaluated.



322 Here the models' performance in representing annual mean discharge (Q_m) and annual maximum
323 daily discharge (Q_p) is evaluated. The details of this procedure can be found in Chapter 6 in Feng
324 (2018).

325 After the weights of model ensemble are obtained using the BMA method, the statistics of
326 posterior probability distribution (here it is assumed to be normal distribution) of estimated
327 changes in Q_m , Q_p and Q_{100} in the future (2081-2100) relative to historical period 1986-2005 are
328 calculated using Eq. (27)-(32).

329 For Q_m , the statistics are:

$$\mu_m = \sum_{k=1}^K w_{k,m} \times c_{k,m} \quad (27)$$

$$\sigma_m^2 = \sum_{k=1}^K w_{k,m} \times (c_{k,m} - \mu_m)^2 \quad (28)$$

330 where μ_m and σ_m are the mean and standard deviation of posterior distribution of relative changes
331 in Q_m ; $w_{k,m}$ is the weight of model k in terms of Q_m ; $c_{k,m}$ is the relative change in Q_m predicted
332 by model k; K is the total number of models, and here it is 90.

333 For Q_p , the statistics are:

$$\mu_p = \sum_{k=1}^K w_{k,p} \times c_{k,p} \quad (29)$$

$$\sigma_p^2 = \sum_{k=1}^K w_{k,p} \times (c_{k,p} - \mu_p)^2 \quad (30)$$

334 where μ_p and σ_p are the mean and standard deviation of posterior distribution of relative changes
335 in Q_p ; $w_{k,p}$ is the weight of model k in terms of Q_p ; $c_{k,p}$ is the relative change in Q_p predicted by
336 model k.

337 For Q_{100} , the statistics are:



$$\mu_{100} = \sum_{k=1}^K w_{k,p} \times c_{k,100} \quad (31)$$

$$\sigma_{100}^2 = \sum_{k=1}^K w_{k,p} \times (c_{k,100} - \mu_{100})^2 \quad (32)$$

338 where μ_{100} and σ_{100} are the mean and standard deviation of posterior distribution of relative
339 changes in Q_{100} ; $w_{k,p}$ is the weight of model k for Q_p ; $c_{k,100}$ is the relative change in Q_{100}
340 predicted by model k . Here, the weights for Q_p are used because Q_{100} is estimated based on the
341 statistics of Q_p series, so it is reasonable to assume that the model having a better ability in
342 reproducing the annual peak discharge should also have a better ability in reproducing the Q_{100} .

343 3 Results and Discussion

344 3.1 Hydrologic model performance

345 The three hydrologic models perform well in representing streamflow dynamics in the study
346 region. The NSE varies within 0.57-0.61 and 0.53-0.62 for calibration and validation periods,
347 respectively, in Mission Creek (gauge NO. 11119750) (Fig. 2). At other calibrated watersheds, the
348 models perform similarly well with NSE varying between 0.45-0.60 for calibration period and
349 0.42-0.62 for validation period (Fig. S2-S4). Simulated streamflow from the three models matches
350 the in-situ measurements in both magnitudes and timing of hydrographs at event scales (Fig. 2b).
351 At annual scale, simulated annual peak flows are comparable to the observations in most years.
352 However, in some years with extremely high events, for example in 1995 January, 1998 February
353 and 2005 January (highlighted in Fig. 2c), the simulated peaks are much lower than the gauge
354 records. This disparity can be attributed to the input bias (e.g., precipitation or streamflow
355 measurements). This is identified using an ‘extreme scenario’ simulation, which assumes 100%
356 precipitation is transformed to surface runoff (i.e., without any loss due to, for example, infiltration
357 or evapotranspiration) and transported immediately to river channels and represents the maximum
358 streamflow considering groundwater is minimal in the study region (Beighley et al., 2003). Even
359 for this extreme scenario, the simulated peaks were still lower (events highlighted in red in Fig.
360 2c) or slightly higher (event highlighted in blue in Fig. 2c) than the gauge observations. This is
361 likely because that model forcings are bias low for these events. One possible source of this bias
362 can be the grid-based precipitation dataset which averages the precipitation rates over the grid



363 masking spatial heterogeneity and thus reducing precipitation rates at some locations. The
364 uncertainties in gauge measurements can also be a bias source. For example, in typical conditions
365 the uncertainty in streamflow measurements ranges between 6%-19% in small watersheds, but it
366 can be higher during large storm events when accurate stage measurements are more difficult
367 (Harmel et al., 2006). Beighley et al. (2003) also identified the overestimation of gauge records at
368 Gauge 11119940 during the 1995 January event. As for mean annual discharge, all three models
369 tend to overestimate for the study period, mainly due to the overestimation of subsurface flow
370 during dry seasons (Fig. 2d). This highlights challenges of simulating hydrologic processes in
371 semiarid regions under a Mediterranean climate.

372 Among the three hydrologic models, STP-HRR has the best overall performance (i.e.,
373 highest average NSE), mainly due to its better ability for capturing flood peaks than the other 2
374 models (Fig. 2, S2-S4). The peak performance is likely a result of the STP-HRR representing the
375 runoff generation process as an exponential relationship between soil moisture and runoff rates,
376 which makes runoff generation more sensitive to soil moisture dynamics as compared to the other
377 2 models. This algorithm is well suited to represent the significant nonlinearity of hydrologic
378 response to rainfall in the study region. RCM-HRR and VIC-HRR have similar overall
379 performance (i.e., similar average NSE), however, they represent hydrologic dynamics differently.
380 VIC-HRR tends to perform better in representing small peak flows than RCM-HRR while worse
381 in simulating mean flow (or total discharge volume) (Fig. 2, S2-S4). This is because as the wet
382 season proceeds, the lower soil layer is close to saturation (i.e., relative soil moisture is higher than
383 the threshold W_s for VIC-HRR) which initiate the quadratic relationship between soil moisture and
384 subsurface runoff in VIC-HRR. This quadratic response to soil moisture conditions can lead to
385 much higher subsurface runoff (2-3 magnitudes higher than that of RCM-HRR), which contributes
386 to the lower performance in reproducing the total volume of discharge. This also explains that
387 VIC-HRR generates the highest subsurface runoff during the wet season (Fig. 3). In addition, VIC-
388 HRR also generates the most surface runoff during wet season (Fig. 3). This is because when soil
389 is almost saturated, surface runoff in VIC-HRR is almost a linear function of precipitation with a
390 coefficient of 1 (much larger than RCM-HRR which is 0.2 (C_2) and STP-HRR which is around
391 0.5 depending on the watershed topography). The higher surface and subsurface runoff generated
392 by VIC-HRR leads to the overestimation of mean annual flow (Fig. 2d). However, there are no in-



393 situ measurement of surface and subsurface fluxes, and it is difficult to evaluate model
394 performance for these quantities individually or as a ratio. In Fig. 3, the simulated surface and
395 subsurface runoff from National Land Data Assimilation Systems VIC model (NLDAS-VIC)
396 output is also shown for purpose of comparison. A similar pattern, i.e., a very high subsurface
397 runoff, even higher than surface runoff, during wet season, can be found from NLDAS-VIC
398 simulations. The surface runoff of NLDAS-VIC is lower than those generated by the models in
399 this study, which is probably because of the difference in precipitation inputs. The NLDAS
400 precipitation input is lower during wet season than that used in this study for the study region. In
401 addition, the difference in spatial resolutions of precipitation (0.125° for NLDAS vs. 0.0625° for
402 this study) can also contribute to the difference in simulated runoff.

403 These results may suggest that STP-HRR is more suitable than VIC-HRR in representing
404 hydrologic processes in Mediterranean regions where 80% annual precipitation is concentrated in
405 a short period (roughly 3 months). As the wet season proceeds, the soil is close to saturation
406 conditions, under which the saturation excess overland flow is dominant. That explains why STP-
407 HRR performs best in this study region. VIC-HRR is probably more suitable to the regions where
408 precipitation events are sparsely distributed where soil is not easy saturated. Although RCM is an
409 empirical method, it performs fairly well in this study, mainly because it captures the nonlinearity
410 of hydrologic processes through a switch between dry and wet surface runoff coefficients (C_1 and
411 C_2) based on the soil moisture conditions.

412 Three sets of parameters with the best performance (assessed by NSE) were selected for
413 each model (Fig. 4). For most of parameters, the selected optimal values are very close, except C_1
414 and K_{s_all} in RCM-HRR, suggesting that most parameters are important factors controlling model
415 performance. For some parameters whose optimal values are close to their range boundaries, for
416 example, K_{ss_all} , W_s , ϕ_{sat} and Q_m , wider and physically acceptable ranges were tested and similar
417 results were obtained, which suggests the ranges of these parameters defined in this study (Table
418 1) are reasonable.

419 3.2 Uncertainty analysis

420 For the 28 major watersheds in SBC, the projected changes in Q_m during 2081-2100 as
421 compared to historical period 1986-2005, range from -80% to 150% (Fig. 5). The median changes
422 for each of these major watersheds are slightly above 0%, varying between 1% and 10%. The



423 major uncertainty sources are GCM and RCP, which account for about 50% of the total
424 uncertainty. Among the first order factors (i.e., GCM, RCP, hydrologic model and
425 parameterization), hydrologic model ranks third after GCM and RCP, accounting for about 10-
426 20% of total uncertainty. In contrast, parameterization only introduces less than 1% of the total
427 uncertainty. The remaining 30-40% uncertainty is from the second, third and fourth order
428 interactions between the four major sources. The projected relative changes in Q_p and Q_{100} during
429 2081-2100 compared to 1986-2005 are similar in magnitudes, both varying from -90% to 240%
430 (Fig. 6 and Fig. 7). The median changes in Q_p and Q_{100} for each watershed are higher than those
431 of Q_m , ranging between 10-40%. For most of watersheds, GCM and RCP are the two major
432 uncertainty contributors for Q_p and Q_{100} , accounting for 40-60% of total uncertainties. Hydrologic
433 model contributes about 10-30% of total uncertainties in Q_p and Q_{100} . Compared to Q_m , Q_p and
434 Q_{100} get more uncertainty from the hydrologic models, which is likely due to highly nonlinear
435 rainfall-runoff behavior and larger differences between runoff generation methods in generating
436 peak flows as compared to average flow conditions.

437 Changes in Q_m , Q_p and Q_{100} are higher under RCP 8.5, but the uncertainties are also higher
438 (Fig. 8), which suggests the uncertainties from RCPs are mainly introduced by RCP 8.5. In Mission
439 Creek watershed (USGS gauge No. 11119750), the probability of increase in Q_m under RCP 4.5
440 is only 51%. However, this probability increases to 64% under RCP 8.5. For the less frequent
441 events (Q_p and Q_{100}), the probabilities of positive changes are higher: 78% and 85% for Q_p and
442 Q_{100} , respectively, under RCP 8.5. This implies that if RCP 8.5 happens in the future, the extreme
443 events will probably get intensified.

444 Compared to previous studies (e.g., Vetter et al. (2015), Schewe et al. (2014), Hagemann
445 et al. (2013);(Troin et al., 2018), and Asadieh and Krakauer (2017)), this work identifies relatively
446 lower uncertainty contributions from hydrologic models. This is mainly because in this study the
447 models use the same model configuration including the model unit definition (irregular
448 catchments) and the hillslope routing scheme (“open-book” assumption), which reduces the
449 difference between hydrologic models. Here, a common calibration approach is also used to
450 eliminate user/method bias which is common in studies that consider more than one hydrologic
451 model. In contrast, the hydrologic models used in previous studies are the individual models which
452 use their own model configurations. For example, the VIC model (here VIC refers to the original
453 VIC models, and is different from the model used in this study; to clarify, in following text, VIC



454 refers to the original VIC model while VIC-HRR refers to the model used in this study) uses the
455 grid-based model units ignoring the spatial arrangement and has its own routing scheme which
456 adopts the synthetic unit hydrograph concept. These differences between models probably resulted
457 in the larger uncertainties in the simulation from hydrologic models in previous studies.

458 Different from previous studies, the hydrologic model uncertainty in this study only comes
459 from the runoff generation algorithms. This can provide useful information for selecting
460 hydrologic models for climate change impact analysis. The results in this study imply that selecting
461 an appropriate runoff generation algorithm suitable to the regions of interest and the study targets
462 (e.g., total volume or extremes) can reduce uncertainties by 10-30%, especially for the extreme
463 quantities (e.g., 100-yr flood discharge). Compared to the runoff generation algorithms, model
464 parameterization plays a negligible role (less than 1%) in the total streamflow uncertainty. This
465 suggests that the parameter equifinality (or non-uniqueness) is less of a concern when quantifying
466 climate change impacts on hydrologic fluxes using an ensemble of GCM forcings. In this study,
467 only one routing scheme is investigated. Although there are fewer variants of routing algorithms
468 as compared to runoff generation methods, the choice of different routing methods can still make
469 a difference in the total uncertainties in streamflow simulation, especially when the model
470 configurations are different, for example, the routing schemes in VIC model and HRR model.
471 Therefore, further study integrating different routing algorithms should be conducted to evaluate
472 the uncertainties in simulated streamflow resulted from both process models (runoff generation
473 model and routing model), which can be useful to guide stakeholders to select appropriate
474 hydrologic algorithms for climate change impacts analysis and develop actionable adaptation and
475 mitigation strategies.

476 **4 Conclusions**

477 A modeling framework which integrates multiple runoff generation algorithms (VIC, STP
478 and RCM) with the Hillslope River Routing model (HRR) is developed. Forced with an ensemble
479 of GCM projections under different emission scenarios, this framework is able to quantify the
480 climate change impacts on streamflow and evaluate the associated uncertainties from different
481 sources (i.e., RCPs, GCMs, hydrologic process models and parameterization). The results in this
482 study show that the median changes in mean annual discharge for the major watersheds in SBC
483 are 1-10%, with an uncertainty of 230% (-80% to +150%); the median changes in annual peak



484 discharge and 100-yr flood discharge are higher than those of mean annual discharge, varying
485 between 10% and 40%, but with a higher uncertainty of 330% (-90% to +240%). For these
486 uncertainties, GCM and RCP are the first two major contributors, accounting for more than 50%
487 of total uncertainties at most watersheds in SBC, while hydrologic process models (i.e., runoff
488 generation modules) contribute between 10% and 30% among watersheds with the remaining 20-
489 40% of the uncertainty coming from the interactions between these individual sources. Hydrologic
490 model parameters alone contribute less than 1% of the uncertainty, which suggests the parameter
491 equifinality should not be a concern when analyzing climate change impacts using ensembles of
492 climate models projections. The results based on the BMA analysis indicate that there is a high
493 probability (up to 85%) that streamflow, especially the extreme quantities like Q_{100} under RCP
494 8.5, will increase in SBC.

495 Unique to this framework, the uncertainties from different hydrologic model components
496 (e.g., runoff generation process) and associated model parameterizations can be identified and
497 quantified. In this study, only one routing scheme is integrated, however, this framework is capable
498 of incorporating multiple routing methods to quantify their contributions to the uncertainties in
499 simulated hydrologic variables. These information can help stakeholders with different focuses
500 (e.g., water resources, risk controls or ecosystem conservation) customize and optimize their
501 selections of hydrologic models and make actionable adaptation decisions under the changing
502 climate.

503 **Code availability**

504 The source code supporting this work is available on Github:
505 <https://github.com/dongmeifeng-2019/HydroUncertainty>

506 **Author contribution**

507 D. Feng and E. Beighley designed the experiments and D. Feng developed the model code and
508 performed the simulations. D. Feng and E. Beighley prepared the manuscript.

509 **Competing interests**

510 The authors declare that they have no conflict of interest.



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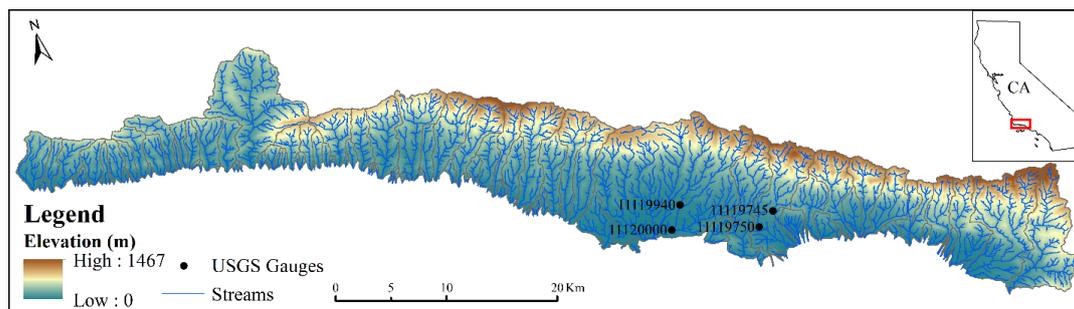


633 **Table 1:** Calibrated parameters for all 3 models

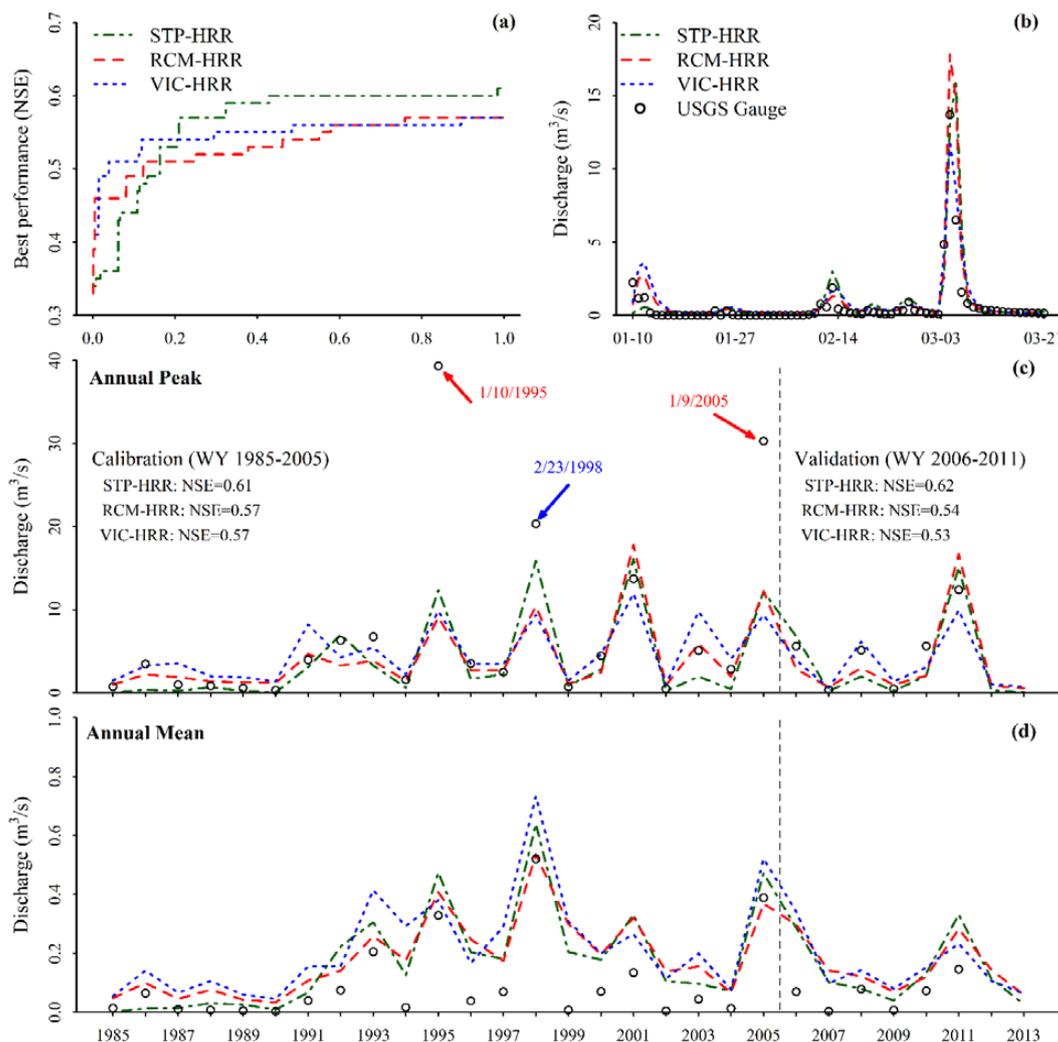
Parameters	Description	Unit	Range	RCM -HRR	VIC- HRR	STP- HRR
K_{s_all}	coefficient to adjust surface roughness	-	1-20	✓	✓	✓
K_{ss_all}	coefficient to adjust horizontal hydraulic conductivity	-	10-200	✓	✓	✓
K_{sat_all}	coefficient to adjust vertical hydraulic conductivity	-	0.01-5.0	✓		
C_1	dry runoff coefficient	-	0-0.3	✓		
C_2	wet runoff coefficient	-	0.2-0.8	✓		
θ_t	soil moisture threshold separating dry and wet conditions	-	0.2-0.8	✓		
b_{in}	Infiltration curve shape parameter	-	0.005-0.5		✓	
D_m	maximum baseflow	$m \cdot d^{-1}$	0 -0.037		✓	
D_s	fraction of D_M where non-linear baseflow begins	-	0 -0.005		✓	
W_s	fraction of the maximum soil moisture where non-linear baseflow occurs	-	0.92-1.0		✓	
f_{over}	Surface runoff coefficient	m^{-1}	0.1-5			✓
f_{drain}	Subsurface runoff coefficient	m^{-1}	0.1-5			✓
Q_m	maximum baseflow	$m \cdot d^{-1}$	0.864-1728			✓
φ_{sat}	Saturated suction head in the soil	m	-3.05-0			✓

634

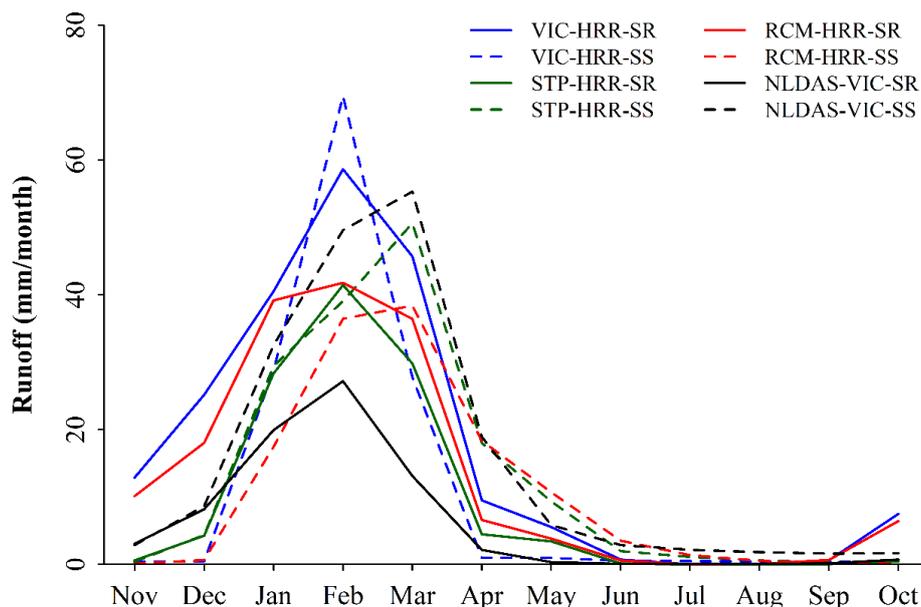
635



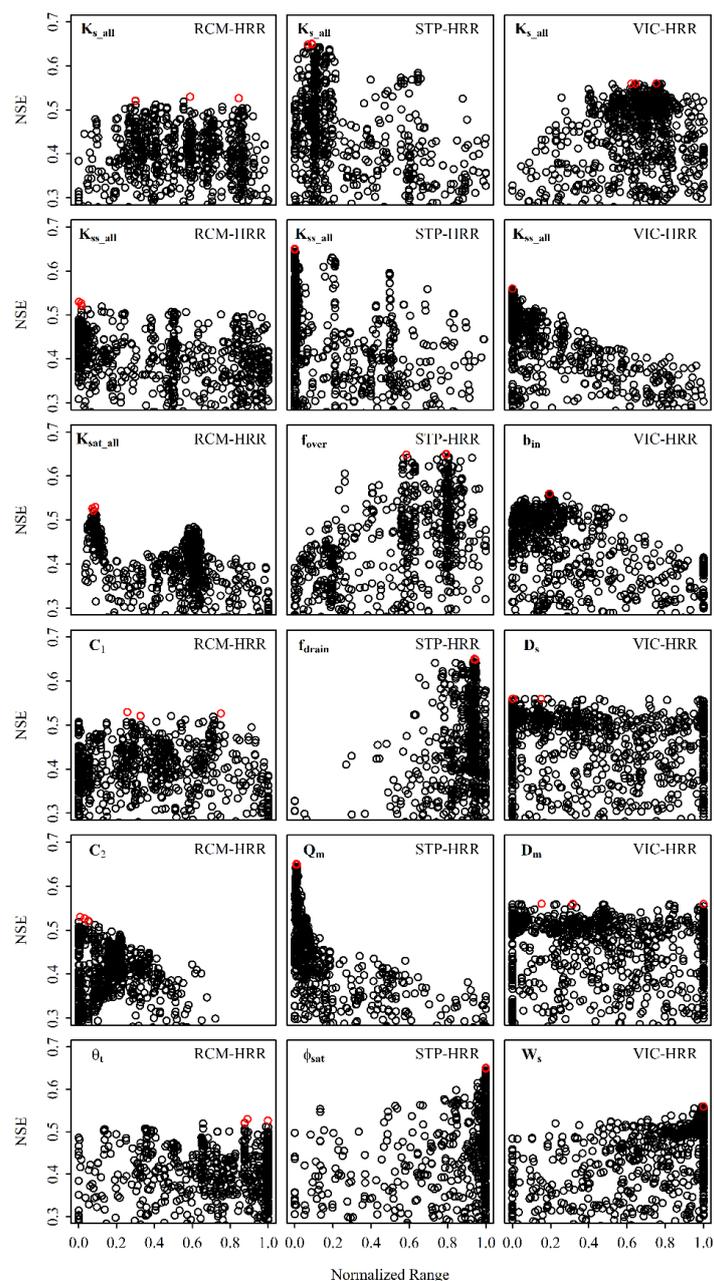
636 **Figure 1: Study region with USGS streamflow gauges. The inset figure indicates the**
637 **location of SBC within the state of California (CA).**



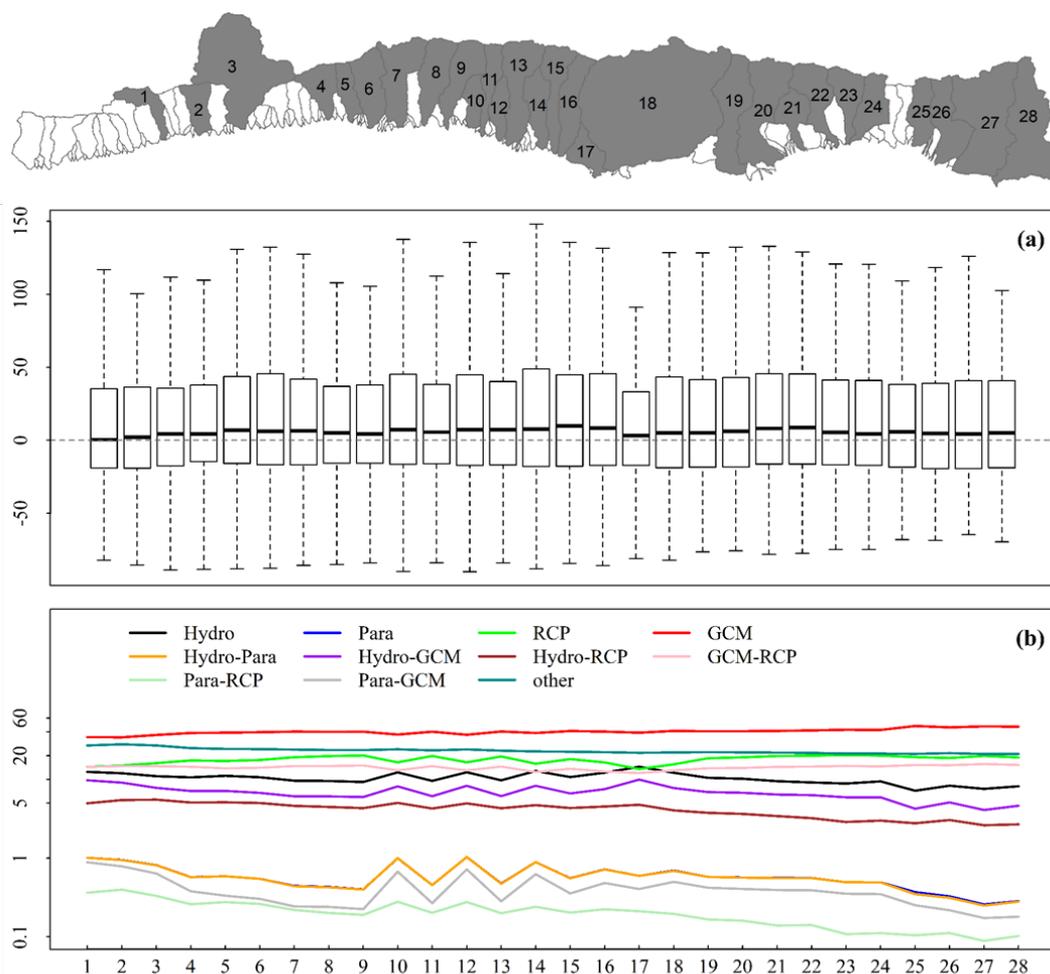
638 **Figure 2** Model performance for calibration and validation periods: (a) model performance
639 (assessed by NSE) during calibration process, x axis is the normalized calibration process; (b)
640 hydrographs simulated by 3 calibrated models and measured by USGS gauge; in order to show the
641 details of the hydrographs, they are zoomed in to the wet season in 2001; the model performance is
642 similar in other years; (c) simulated annual peak flow during calibration (water year 1985-2005)
643 and validation (water year 2006-2011) periods as compared with in situ observations; texts indicate
644 model performance (i.e., NSE) in reproducing historical hydrographs for both periods; the points
645 highlighted in red arrows indicate the events were not reproduced by models due to the input (e.g.,
646 precipitation or discharge observation) bias; the point highlighted in blue arrow is similar to those
647 in red but at a lower probability; and (d) simulated and observed annual mean flow during
648 calibration and validation periods. For clarity, only results for Mission Creek watershed (USGS
649 gauge NO. 11119750) are shown here; results for other gauged watersheds are similar and can be
650 found in the Supporting Information (Figure S1-S3).



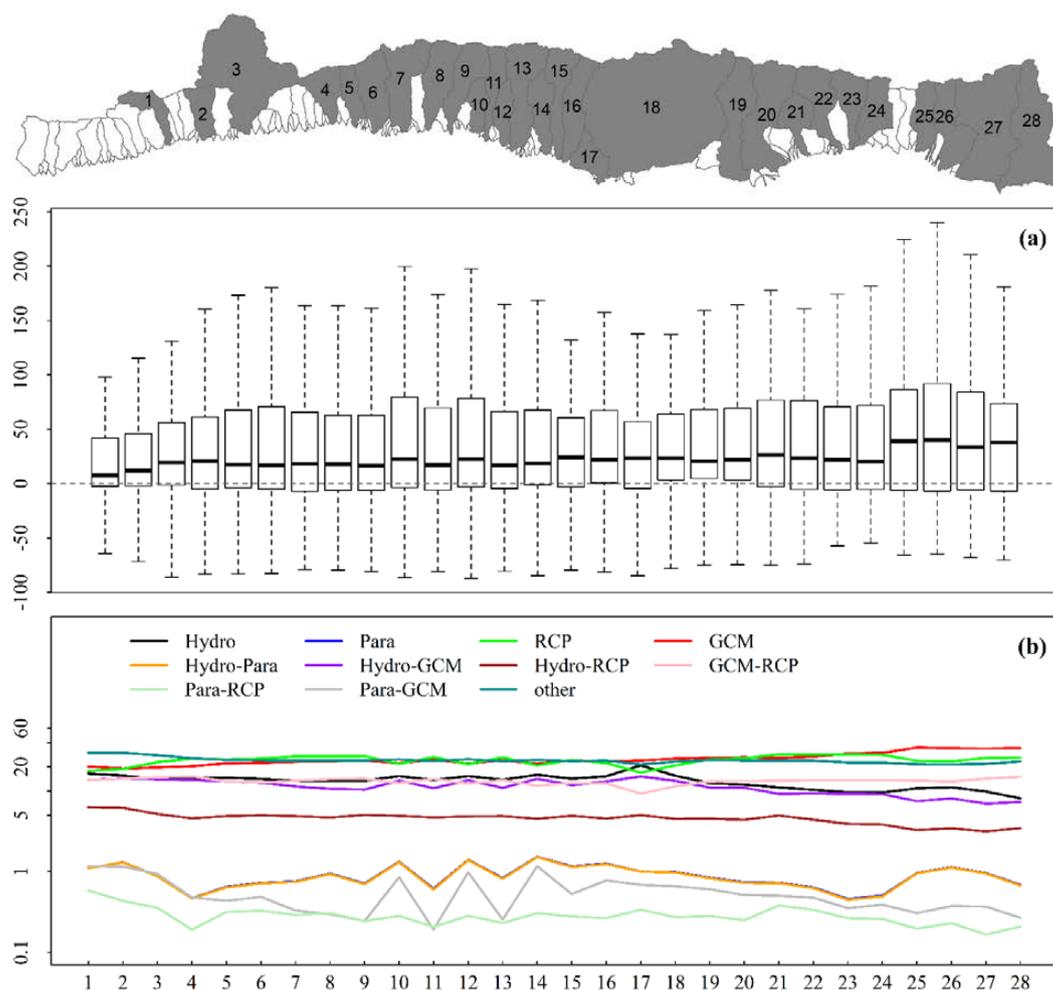
651 **Figure 3: Simulated monthly surface and subsurface runoff from Mission Creek watershed (USGS**
652 **gauge NO. 11119750) by three models for the calibration period (water year 1985-2005). Surface**
653 **runoff is denoted by ‘SR’ and subsurface runoff is denoted by ‘SS’ in this figure. Monthly surface**
654 **and subsurface runoff from National Land Data Assimilation Systems (NLDAS) VIC model**
655 **simulation for the same period are shown here for comparison purpose.**



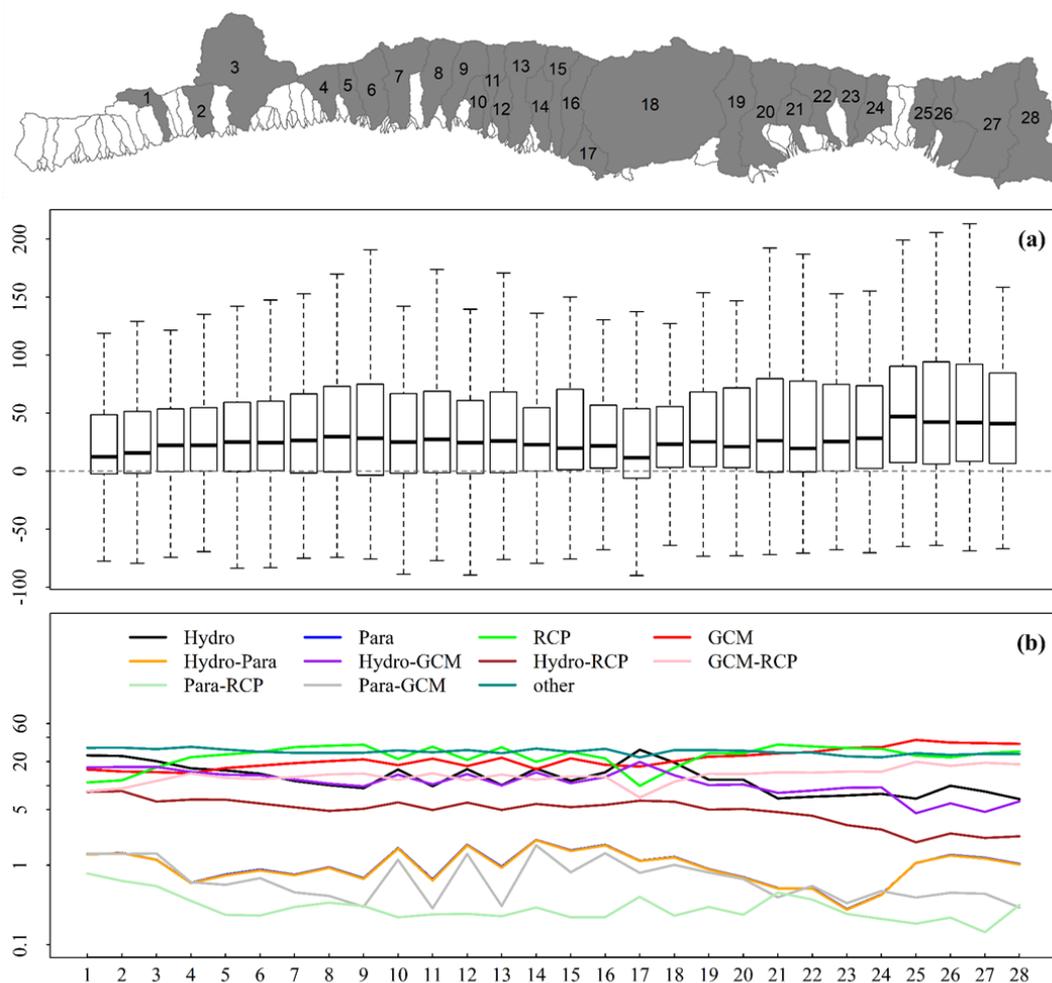
656 **Figure 4.** Parameters sampled during calibration process and their corresponding performance
 657 (assessed by NSE). The black circles are parameter samples within the predefined ranges (shown in
 658 Table 1) and the red circles indicate the optimal values used for further uncertainty analysis. The
 659 parameter values are normalized by their ranges, so the range of x axis in all plots is 0-1. The
 660 parameters were sampled throughout their whole ranges, however, for clarity, samples with NSE
 661 lower than 0.3 are not shown in this figure.



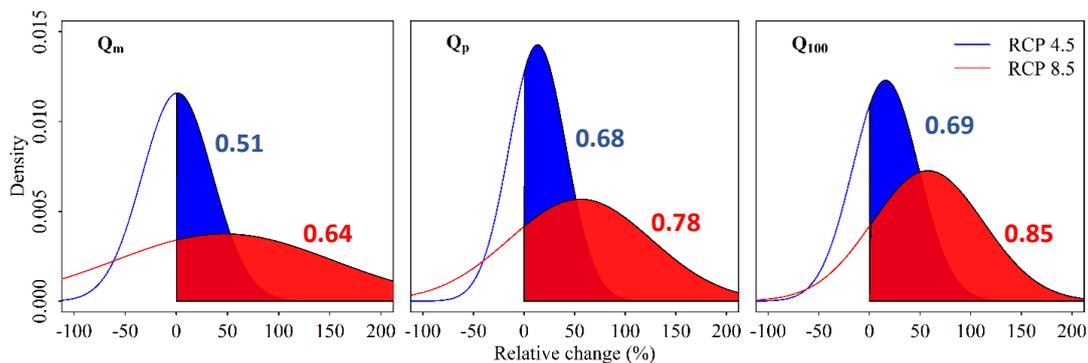
662 **Figure 5. (a) Projected relative changes (%) in annual mean discharge (Q_m) in the major**
 663 **SBC watersheds (indicated by the grey watersheds in the map) during 2081-2100 as**
 664 **compared to historical period (1986-2005); each bar depicts relative changes in minimum,**
 665 **maximum, median, 1st and 3rd quartiles for the ensemble outputs; bars from left to right**
 666 **spatially corresponding to watersheds from west to east. For clarity, only watersheds with**
 667 **drainage areas larger than 7 km², which account for roughly 83% of the study area, are**
 668 **shown. (b) Relative sources (%) of the uncertainties in the projected changes at each of these**
 669 **watersheds; the category “other” is the uncertainty from the 3rd and 4th orders of interactions**
 670 **between the 4 major sources (i.e., GCMs, RCPs, Hydrologic models, denoted by “Hydro”**
 671 **and parameters denoted by “Para”)**



672 **Figure 6. (a) Projected relative changes (%) in annual mean discharge (Q_p) in the major SBC**
 673 **watersheds (indicated by the grey watersheds in the map) during 2081-2100 as compared to**
 674 **historical period (1986-2005); each bar depicts relative changes in minimum, maximum,**
 675 **median, 1st and 3rd quartiles for the ensemble outputs; bars from left to right spatially**
 676 **corresponding to watersheds from west to east. For clarity, only watersheds with drainage**
 677 **areas larger than 7 km², which account for roughly 83% of the study area, are shown. (b)**
 678 **Relative sources (%) of the uncertainties in the projected changes at each of these**
 679 **watersheds; the category “other” is the uncertainty from the 3rd and 4th orders of interactions**
 680 **between the 4 major sources (i.e., GCMs, RCPs, Hydrologic models, denoted by “Hydro”**
 681 **and parameters denoted by “Para”)**



682 **Figure 7. (a) Projected relative changes (%) in 100-yr flood discharge (Q_{100}) in the major**
 683 **SBC watersheds (indicated by the grey watersheds in the map) during 2081-2100 as**
 684 **compared to historical period (1986-2005); each bar depicts relative changes in minimum,**
 685 **maximum, median, 1st and 3rd quartiles for the ensemble outputs; bars from left to right**
 686 **spatially corresponding to watersheds from west to east. For clarity, only watersheds with**
 687 **drainage areas larger than 7 km², which account for roughly 83% of the study area, are**
 688 **shown. (b) Relative sources (%) of the uncertainties in the projected changes at each of these**
 689 **watersheds; the category “other” is the uncertainty from the 3rd and 4th orders of interactions**
 690 **between the 4 major sources (i.e., GCMs, RCPs, Hydrologic models, denoted by “Hydro”**
 691 **and parameters denoted by “Para”)**



692 **Figure 8. Probability of changes in Q_m , Q_p and Q_{100} at Mission Creek watershed (No. 20 in**
693 **Figure 5 map). The numbers in the plot are the probabilities of positive changes in Q_m , Q_p**
694 **and Q_{100} (areas of shaded regions) under each emission scenario (blue numbers are for RCM**
695 **4.5 and red numbers are for RCP 8.5).**